



HAL
open science

The structure and robustness of ecological networks with two interaction types

Virginia Domínguez-García, Sonia Kéfi

► **To cite this version:**

Virginia Domínguez-García, Sonia Kéfi. The structure and robustness of ecological networks with two interaction types. *PLoS Computational Biology*, 2024, 20 (1), pp.e1011770. 10.1371/journal.pcbi.1011770 . hal-04547972

HAL Id: hal-04547972

<https://hal.umontpellier.fr/hal-04547972>

Submitted on 16 Apr 2024

HAL is a multi-disciplinary open access archive for the deposit and dissemination of scientific research documents, whether they are published or not. The documents may come from teaching and research institutions in France or abroad, or from public or private research centers.

L'archive ouverte pluridisciplinaire **HAL**, est destinée au dépôt et à la diffusion de documents scientifiques de niveau recherche, publiés ou non, émanant des établissements d'enseignement et de recherche français ou étrangers, des laboratoires publics ou privés.

THE STRUCTURE AND ROBUSTNESS OF ECOLOGICAL NETWORKS WITH TWO INTERACTION TYPES

LETTER

Virginia Domínguez-García, ORCID 0000-0002-4591-4186

ISEM, CNRS, Univ. Montpellier, IRD, Montpellier, France;

Estación Biológica de Doñana (EBD-CSIC) , Avda. Américo Vespucio s/n, Isla de la Cartuja, E-41092 Sevilla, Spain;

`virginia.dominguez@ebd.csic.es`

Sonia Kéfi, ORCID 0000-0002-9678-7770

ISEM, CNRS, Univ. Montpellier, IRD, Montpellier, France;

Santa Fe Institute, 1399 Hyde Park Road, Santa Fe, NM 87501, USA

`sonia.kefi@umontpellier.fr`

April 11, 2024

1 **Full Title:** The structure and robustness of ecological networks with two interaction types

2 **Short Title:** Structure and robustness of ecological networks with two interactions

3 **Author contribution:**

4 • Conceptualization: S.K. and V.D.-G.

5 • Data Curation: V.D.-G.

6 • Formal Analysis: V.D.-G.

7 • Funding Acquisition: S.K.

8 • Investigation: V.D.-G. and S.K.

9 • Software: V.D.-G.

10 • Supervision: S.K.

11 • Visualization: V.D.-G.

12 • Writing – Original Draft Preparation: V.D.-G. and S.K.

13 • Writing – Review & Editing : V.D.-G. and S.K.

14 **Data accessibility statement:** The data and code supporting the results are available in Zenodo: [10.5281/zenodo.10198613](https://zenodo.org/doi/10.5281/zenodo.10198613)

15 **Corresponding author:** Virginia Domínguez-García. Estación Biológica de Doñana (EBD-CSIC), Avda. Américo Vespucio
16 s/n, Isla de la Cartuja, E-41092 Sevilla, Spain. virginia.dominguez@ebd.csic.es

ABSTRACT

17 Until recently, most ecological network analyses investigating the effects of species' declines and extinctions
18 have focused on a single type of interaction (e.g. feeding). In nature, however, diverse interactions co-occur,
19 each of them forming a layer of a 'multilayer' network. Data including information on multiple interaction
20 types has recently started to emerge, giving us the opportunity to have a first glance at possible commonalities
21 in the structure of these networks. We studied the structural features of 44 tripartite ecological networks from
22 the literature, each composed of two layers of interactions (e.g. herbivory and pollination), and investigated
23 their robustness to species losses. Considering two interactions simultaneously, we found that the robustness
24 of the whole community is a combination of the robustness of the two ecological networks composing it.
25 The way in which the layers of interactions are connected to each other affects the interdependence of their
26 robustness. In many networks, this interdependence is low, suggesting that restoration efforts would not
27 automatically propagate through the whole community. Our results highlight the importance of considering
28 multiple interactions simultaneously to better gauge the robustness of ecological communities to species loss
29 and to more reliably identify key species that are important for the persistence of ecological communities.

30 Author Summary

31 In the face of the current biodiversity crisis, predicting how species loss will affect ecological communities is becoming increasingly
32 relevant. Previous studies including only one type of ecological interactions (e.g. feeding or pollination) revealed the relevance
33 of the structure of ecological networks for the persistence of ecological communities. However, there is mounting evidence
34 that considering multiple interactions simultaneously can alter the results based on a single interaction. Here, we study the
35 robustness of ecological networks with two interaction types to the loss of plant species, and we show that it is a combination
36 of the robustness of the two bipartite ecological networks composing the ecological community. By analyzing networks from
37 multiple communities, we are able to identify commonalities across interaction types, as well as singularities specific to a given
38 interaction type, caused by underlying biological constraints. Our results highlight that a multi-interaction approach is crucial to
39 better gauge the overall robustness of ecological communities, and to correctly determine the relative importance of different
40 plants species at the whole community level, which can be key for biodiversity conservation.

41 INTRODUCTION

42 The rate of decline of many species populations is accelerating [1], and species extinctions are seriously threatening the functioning
43 of ecological communities worldwide. Understanding how species interact and how this affects the robustness of ecological
44 communities to species loss is essential to anticipate the consequences of biodiversity losses and extinction cascades as well as
45 to design protection and restoration plans. The study of ecological networks – where species are represented by nodes and the
46 ecological interactions by links between these nodes – have contributed significantly to the understanding of how ecological
47 interactions are structured and have unveiled important relationships between network structure and their robustness to species
48 loss [2, 3, 4, 5]. However, while the ecological network literature has long been dominated by studies of networks containing
49 a single interaction type, it has become increasingly clear that species in nature are connected by a myriad of interaction
50 types simultaneously and that considering networks which include this diversity of interaction types could greatly improve our
51 knowledge of the structure and dynamics of ecological communities [6, 7, 8, 9, 10, 11, 12, 13].

52 A number of previous studies have investigated the effect of including multiple interaction types on the functioning of ecological
 53 communities, especially on their stability [14, 15, 16, 17, 18, 19, 20]. Yet the vast majority of these studies have so far remained
 54 theoretical. With the publication of the first multi-interaction empirical networks, we begin to know more about their structure
 55 [6, 21, 22, 10, 23, 24, 25, 26, 27, 13], and how this structure affects their persistence [6, 10] and robustness [21, 28, 24, 27]. In
 56 particular, studies on multi-interaction networks have provided new insights on whether the inclusion of several interactions
 57 can significantly alter their robustness to species loss [24] and how extinctions propagate through such networks [21]. However,
 58 in spite of these pioneering studies, there is currently no consensus about the structure of multi-interaction networks and its
 59 consequences for the robustness of ecological communities, in part due to the lack of data sets, whose amount has only recently
 60 started to increase.

61 A key question, of relevance given the current biodiversity crisis, is how robustness varies across network types, and what we can
 62 learn from including multiple interactions simultaneously. With this in mind, we gathered ecological networks with multiple
 63 interaction types currently available in the literature. More specifically, we focused on tripartite networks because they were
 64 the most abundant in the literature, allowing us to compare a wide variety of ecological systems. Tripartite ecological networks
 65 are composed of two interaction layers (e.g. pollination and herbivory), each of the bipartite kind [29]. They therefore contain
 66 three different *species sets* (e.g. plant, pollinator and herbivore guilds in a pollination-herbivory network), one of which is shared
 67 between the two interaction layers (e.g. plant species can interact with both pollinators and herbivores in a pollination-herbivory
 68 network). We call the set of nodes that can have interactions in both interaction layers the *shared set*, and the subset of nodes in
 69 the shared set that have interactions in both interaction layers the *connector nodes* (see Fig 1A and B).

70 Our data set consists of 44 tripartite networks from 6 different studies, in which the interaction layers include mutualistic
 71 (pollination, seed-dispersal and ant-mutualism) and antagonistic (herbivory and parasitism) interactions (see Table 2 and Methods).
 72 To identify possible generalities across interaction types as well as singularities specific to a given interaction type, we divided the
 73 networks in three types according to the signs of the interactions involved: mutualism-mutualism (MM) if both interactions were
 74 positive, antagonism-antagonism (AA) if both interactions were negative, and mutualism-antagonism (MA) if one interaction was
 75 positive and the other negative, given that interaction type can determine network architecture through the underlying biological
 76 constraints [31].

77 Using this data set, we investigated how the two interaction layers are connected and the consequences for the robustness of these
 78 networks to plant loss. Robustness was assessed by sequentially removing plants in a random order and estimating secondary
 79 extinctions (Fig. 1C and Methods). Although this approach lacks realism (since there are no underlying temporal dynamics), it
 80 has proven useful in understanding the threat that biodiversity loss poses to ecosystem services and functioning [32, 3, 33, 21].
 81 Furthermore, it provides a lower bound on the damages that may be caused to an ecological community since it relies on the
 82 conservative hypothesis that secondary extinctions happen only when an animal species has lost all its links. We focused on
 83 the extinctions of plants because they are the only group of species, whose disappearance can potentially harm all other species
 84 groups, and also because plants can be managed more directly [21]. Note that while plants are not the shared set of species in
 85 all networks (see Fig. 1), it is still possible to quantify robustness to plant loss in all the networks of our data set (Methods).
 86 Extending the study of robustness to include multiple interactions simultaneously allowed us to study the interdependence of
 87 the robustness of animal species sets (Fig. 1D), which is relevant to know how cascading extinctions will propagate through a
 88 multi-interaction network [21], and to better identify keystone plant species [21, 13], of importance when designing protection
 89 and restoration interventions. We used four null models with increasing constraints (Methods) to study how different structural
 90 properties could determine the interdependence and robustness in the tripartite networks.

91 Taken together, our results suggest that considering multiple ecological interactions simultaneously does not have a dramatic
 92 impact on the robustness of tripartite networks to plant losses. However, a multi-interaction approach is crucial to better gauge the
 93 overall robustness of ecological communities, to know the interdependence of the robustness of the different animal sets, and
 94 to correctly determine the relative importance of different plants species at the whole community level, which can be key for
 95 biodiversity conservation.

96 RESULTS

97 Different ways of connecting the interaction layers

98 We gathered a total of 44 ecological networks, each containing two types of ecological interactions, including mutualistic
 99 (pollination, seed-dispersal and ant-mutualism, corresponding to respectively 19, 3 and 1 networks) and antagonistic (herbivory
 100 and parasitism, corresponding to respectively 41 and 24 networks) interactions. We divided these networks in three types according
 101 to the signs of their interactions: mutualistic-mutualistic, mutualistic-antagonistic, and antagonistic-antagonistic (see Methods).

102 To study how the interaction layers are connected, we focused our attention on the shared set of species between the two interaction
 103 layers. We measured three structural properties of the shared species: the proportion of the shared species that are connector
 104 nodes, i.e. that have links in both interaction layers (C); the proportion of shared species hubs, i.e. 20% of the shared species with
 105 the most connections, that are connectors nodes (H_C); and the participation coefficient of the connector nodes between the two
 106 interaction layers, i.e. how well split between the two interaction layers are their links (PC_C) (see Methods).

107 This revealed fundamental differences across the three types of tripartite networks (Fig. 2). In antagonistic-antagonistic networks,
 108 $\sim 35\%$ of the shared species (herbivore hosts) are involved in both parasitic and herbivory interactions (i.e. are connector nodes).
 109 Moreover, most of the shared species hubs ($\sim 96\%$) are acting as connectors between interaction layers, and they have their links
 110 equally split among the two interaction layers (average PC_C of 0.89). We found a very different pattern in mutualistic-mutualistic
 111 networks, for which only $\sim 10\%$ of the shared species (plants in this case) are involved simultaneously in the two types of
 112 mutualistic interactions, and only 32% of shared species hubs act as connector nodes. Also, the connector nodes have their links
 113 less equally split among the two interaction layers (average PC_C of 0.59). Mutualistic-antagonistic networks are not significantly
 114 different from mutualistic-mutualistic networks and tend to have values intermediate between those of antagonistic-antagonistic
 115 and mutualistic-mutualistic networks (Fig. 2A-C). About $\sim 22\%$ of the shared species are involved simultaneously in the two
 116 types of mutualistic interactions, $\sim 56\%$ of shared species hubs act as connector nodes and the average PC_C is ~ 0.59 . An

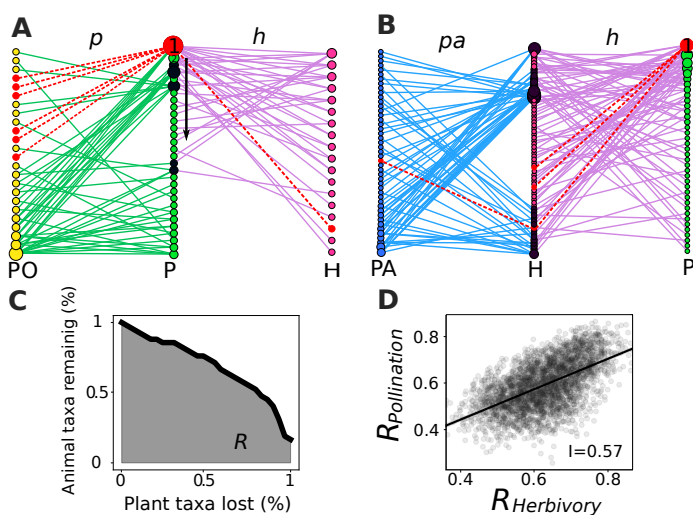


Fig 1: Tripartite networks, robustness and interdependence. A) An Herbivory(h) - Pollination(p) tripartite network, where plants (P) are the shared set of species. B) An Herbivory(h) - Parasitism(pa) tripartite network, where herbivores (H) are the shared set of species. Link colours represent the two interaction layers, and node colours the three sets of species. Connector nodes in the shared set of species are highlighted in black. C) Extinction curve showing the fraction of surviving animal species as a function of plant loss for a given plant extinction sequence in network A. The robustness to plant loss, R , is the area under the curve. Extinction protocol: plants (green nodes) are progressively removed from the community in the prescribed order, their corresponding links are erased (colored in red) and animal species are declared extinct (colored in red) whenever they lose all their feeding links. D) Pairwise correlation in the robustness of the two animal sets – interdependence, I – resulting from 3,000 simulations of random sequential loss of plant taxa in network A.

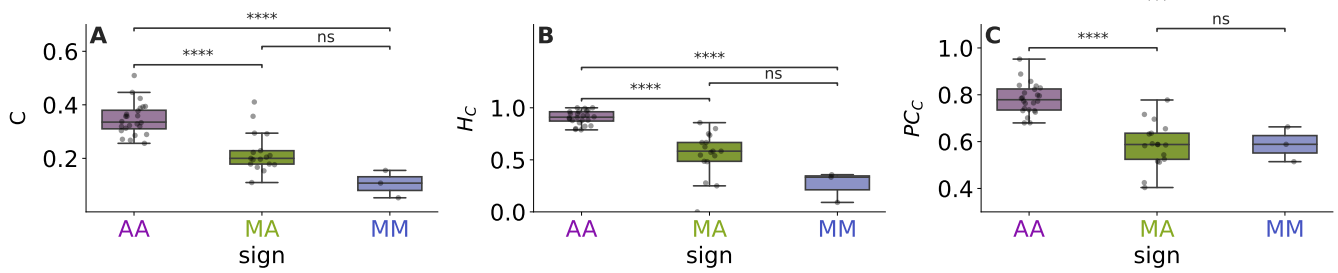


Fig 2: How does the shared set of nodes connect the network? A) Proportion of connector nodes in the shared set, B) Proportion of shared set hubs that are connector nodes, C) Average participation coefficient of the connector nodes. Boxplots are color-coded by network type: AA: Antagonistic-Antagonistic, MA: Mutualistic-Antagonistic, and MM: Mutualistic-Mutualistic. Differences among categories are measured by independent t-tests (**** $p < 1e^{-4}$, *** $p < 1e^{-3}$, *ns* not significant).

117 example of this contrasting structure is visible at a glance in the way the connector nodes link the interaction layers differently in
 118 the two networks in Fig. 1.

119 Interdependence of the robustness of animal species

120 We expected these differences in structure to affect the correlation of the robustness of the two animal species sets. Following
 121 recent studies [21, 28], for each network, we measured the robustness of the two animal species sets following the extinction of
 122 plants, and we investigated whether they were correlated, i.e. if they were interdependent (Fig. 1.D and Methods). When driving
 123 plants to extinction, a ‘high’ correlation between the robustness of the two animal species sets implies that the same plants that are
 124 important for one of the species set are also important in the other species set [21], (e.g. the plants whose extinctions lead to a
 125 relatively high number of secondary extinctions of pollinators also do so for herbivores).

126 We found that, in general, when plants are driven to extinction in a random order, interdependence (I) is either positive or null
 127 (Fig. 3A), with, again, fundamental differences between antagonistic-antagonistic networks and the two other types of networks.
 128 The value of interdependence found in antagonistic-antagonistic networks is on average significantly higher from that found
 129 in the other two network types, which is consistent with our results on hubs and connectors, suggesting that the two layers in
 130 antagonistic-antagonistic networks tend to be strongly interconnected. Note that data collection for parasitoids relies on their
 131 sampling on herbivores found on leaves. This sampling difference (compared to the two other types of networks where species
 132 sets can be collected independently from each other) could potentially introduce a positive correlation. However, the correlation

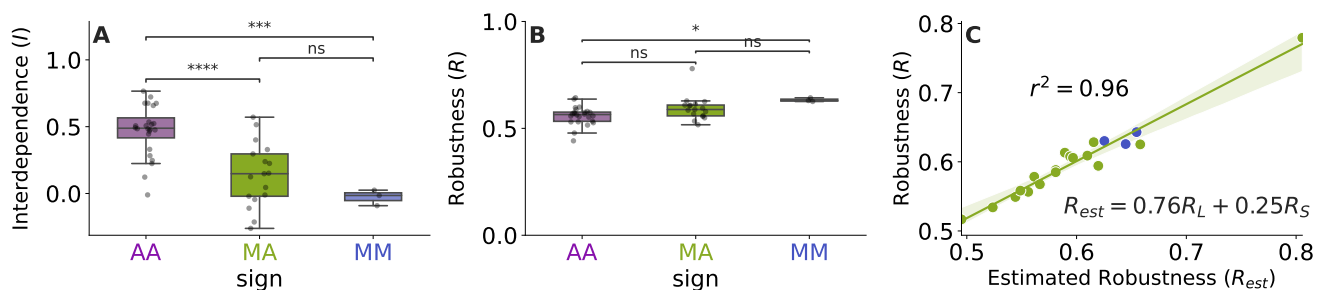


Fig 3: Interdependence and robustness of tripartite networks. A) Interdependence (I) of the tripartite networks in our data set. As $I \rightarrow 1$ the importance of plants for the maintenance of the two animal species sets becomes more similar. B) Robustness of the tripartite networks in our data set (R) when plants are randomly driven to extinction. As $R \rightarrow 1$, animal groups are increasingly robust to the simulated sequential loss of plant taxa. Grey points represent the values in each network. All boxplots are color-coded based on the type of tripartite network. Differences among the categories are measured by independent t-tests (**** $p < 1e^{-4}$, *** $p < 1e^{-3}$, * $p < 5e^{-2}$, *ns* not significant). C) Robustness (R) vs Estimated Robustness (R_{est}) in the empirical MA and MM networks of our database. The text shows the best estimation of the robustness as a combination of the robustness of the larger (R_L) and smaller (R_S) bipartite networks that compose the tripartite network, and the correlation coefficient. Each point represents a network, color coded based on network type. AA: Antagonistic-Antagonistic in purple, MA: Mutualistic-Antagonistic in green, and MM: Mutualistic-Mutualistic in blue.

we found is not significantly higher from what is expected in the null models. The presence of this positive correlation in the four null models considered (Fig. S7B-D, SI 5) suggests that it is due to the particular layout of these networks, more specifically, to the cascading extinction process characteristic of these tripartite antagonistic-antagonistic networks, in which plants are not the shared set of species, meaning that their extinctions sequentially spread from plants to herbivores and to parasites.

In mutualistic-mutualistic networks, the interdependence is close to null, meaning that the robustness of the two species sets seem largely decoupled from each other (but more correlated than expected by chance if we do not control for degree heterogeneity, i.e. the heterogeneity of the number of links each species has (Fig. S7B-C).

Mutualistic-antagonistic networks exhibit a range of values going from moderate correlations ($I \sim 0.5$) to weak negative correlations ($I \sim -0.2$), and comparisons with null models showed a similar trend as in mutualistic-mutualistic networks, with empirical networks being more correlated than their randomized counterparts without taking degree heterogeneity into account (Fig. S7B-C.)

Studying how interdependence relates to the three structural features we measured revealed differences among network types as well. More specifically, in antagonistic-antagonistic networks, interdependence is correlated (albeit weakly) with the proportion of connectors (C), while in the other two network types it varies with the proportion of hubs that are connectors (H_C) and their (un)balanced participation in the two interaction layers (PC_C) (Table 1, and SI 5).

Tripartite networks' robustness

The robustness of antagonistic-antagonistic networks was found to be lower than that of Mutualistic-Mutualistic networks when plants were randomly driven to extinction (Fig. 3B), although differences among the three types of networks are overall not significant. Surprisingly, this suggests that even if the different ways in which the tripartite networks are connected seem to have a significant effect on interdependence, this difference does not translate into significant differences in the global robustness of the tripartite networks. In other terms, a higher interdependence between the interaction layers does not cause a lower overall robustness. As expected, all tripartite ecological networks were most fragile when plants were selectively attacked targeting the most connected plants first, and the least fragile when plants were attacked selecting the specialists plants first, as previously reported in networks with only one interaction type [2, 3, 34, 35, 36] (Fig.S9).

The structural features that most determine the robustness of the networks are the degree heterogeneity and the proportion of connector nodes in mutualistic-antagonistic and in mutualistic-mutualistic networks, as well as the even split of links between the two interaction layers in antagonistic-antagonistic networks (Table 1, and SI 6). We included the degree heterogeneity of nodes in the analysis (i.e. the variance of the interaction degree divided by the average degree) because broad degree distributions are known to make ecological networks with one interaction type more robust to random deletion of species [37, 3], a result we recover here in the case of tripartite networks. Comparison with the null models further corroborates this result, since the robustness of mutualistic-mutualistic and mutualistic-antagonistic networks was not significantly different from that of their randomized counterparts when degree heterogeneity is conserved (Fig. S10C-E).

Furthermore, the robustness of the tripartite networks could be predicted by the robustness of the two bipartite networks composing it (Fig. 3C). The estimated overall robustness, a combination of the robustness of the two bipartite networks (Methods), is in very good agreement ($R^2 = 0.96$) with the robustness of the tripartite networks. When the robustness of only one bipartite network was used, R^2 was at most 0.8 (SI 6, Fig. S11). While in the main text we only consider the classical co-extinction algorithm in unweighted networks because it is the more parsimonious and offers a lower bound to the damage the community can suffer, we show that the results hold when using a stochastic version in weighted networks[42] (Fig. S14 and 15).

Table 1: Table of regression of interdependence (I) and robustness (R) on the structural features we studied: degree heterogeneity ($\sigma_k / \langle k \rangle$), proportion of connector nodes (C), proportion of shared species hubs that are connectors (H_C), and (un)even split of interactions among interaction layers (PC_C).

	I		R	
	AA	MA & MM	AA	MA & MM
$\sigma_k / \langle k \rangle$		0.30	0.68***	0.38*
C	0.40*		0.24**	-0.51**
H_C		0.70***		
PC_C		0.50**	-0.25*	
Observations	24	20	24	20
R^2	0.16	0.70	0.76	0.38
Adjusted R^2	0.12	0.64	0.73	0.31
F Statistic	4.20*	12.41***	21.39***	5.18**

Note: *p<0.1; **p<0.05; ***p<0.01

171 Plant importance for robustness

172 The results on interdependence suggest that the important plants for one set of animal species may not always be as important for
 173 the other species set (e.g. important plants for pollinators may not be important for herbivores and vice versa). We investigated this
 174 point further and asked which plants were more important for the survival of the whole ecological community, and to what extent
 175 those plants were the same for the two animal species sets. We therefore built three rankings of plant importance – one for each
 176 animal species set and one for the whole community – in which a plant is considered to be more important if robustness is lower
 177 when that plant is attacked earlier in the extinction sequence [21] (Methods). For example, a plant can be considered important
 178 based on the pollinator and whole community rankings (e.g. plant 1, Fig. 4A, B), but not so based on the herbivore ranking (Fig.
 179 4C). Other plants can be important based on the three rankings (e.g. plant 2, Fig. 4D-F). Comparing the three rankings in the
 180 example shows that plant importance when the two interaction layers are considered simultaneously (whole community) is not
 181 just a simple combination of the ranking of plant importance for each set of animal species (Fig. 4G). While it is more difficult to
 182 differentiate between the less important plants (those with lower values of importance), the ranking is well defined, as can be seen
 183 from the correlation values between the importance and the ranking based on importance (Fig.S16). Interestingly, it becomes
 184 better defined when the two interaction layers are considered simultaneously.

185 We studied to what extent the importance of a given plant at the whole community level was driven by its importance for the
 186 two animal species sets (Fig. 4H). In the majority of networks ($\sim 63\%$), the importance of a plant for the whole community is a
 187 mixture between its importance for the two animal sets (i.e. the similarity between the ranking in the whole community and in the
 188 animal sets (S_{set}) is between 0.5 and 0.9; Methods), while in $\sim 25\%$ of the networks it is mostly driven by its ranking in one of
 189 the animal species sets (i.e S_{set} of one animal species set is above 0.9). This was especially relevant in mutualistic-mutualistic
 190 networks, where 2 out of the 3 networks lie in this category, probably because of the high dissimilarity between the sizes of the
 191 two animal sets (180 pollinators vs 27 seed-dispersers and 173 pollinators vs 30 ants). In a few cases ($\sim 12\%$), the ranking of plant
 192 importance for the whole community did not resemble any of the rankings for the animal sets (i.e. both S_{set} were below 0.5),
 193 meaning that the importance of a plant when the two interactions are considered simultaneously changes dramatically compared to
 194 its importance when the interactions are considered separately. While in the main text we only consider the classical co-extinction
 195 algorithm in unweighted networks because it is the more parsimonious and offers a lower bound to the damage the community
 196 can suffer, we show that these results hold when using a stochastic version in weighted networks (Fig. S14-15).

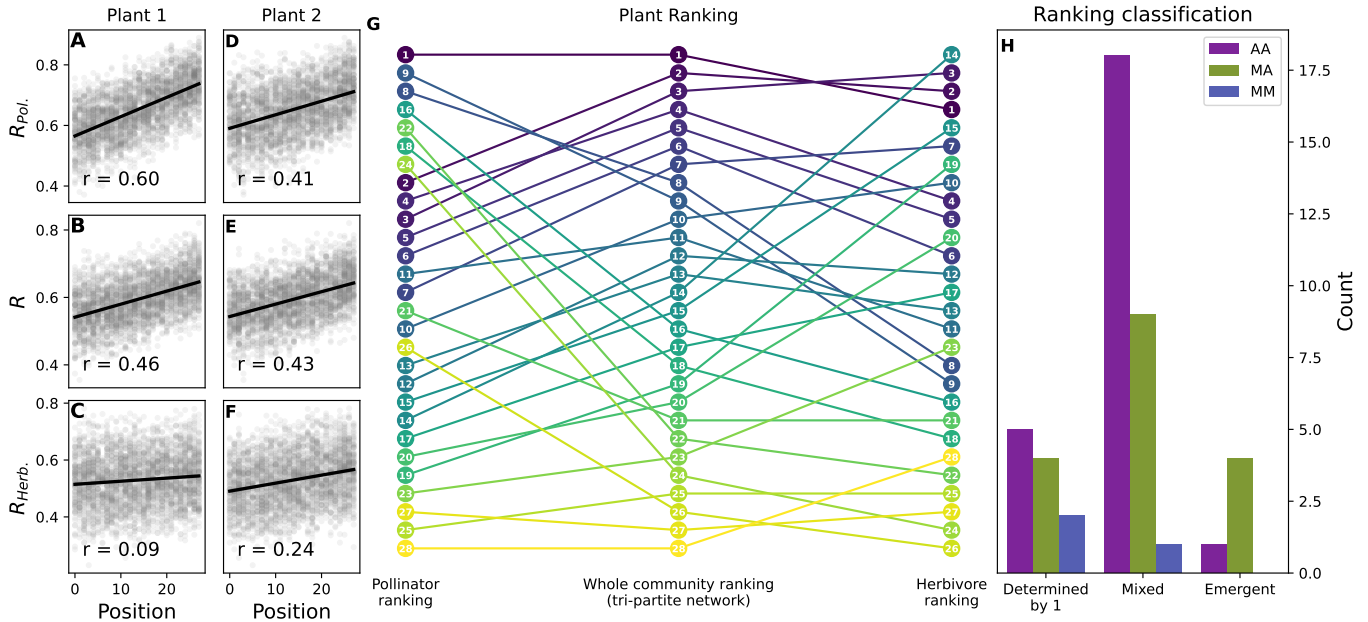


Fig 4: Plant importance rankings. A-F: Scatter plot of the robustness of pollinators (R_P), of the tripartite network (R), and of herbivores (R_H) vs the order of two plants (plant 1 and plant 2, chosen as an illustrative example) in the extinction sequence. The correlation coefficients are used to determine the ranking of importance of plant species. G: Ranking of plant importance for pollinators (left), for the whole community (center) and for herbivores (right). Each plant is represented by a disk whose number reflects its order in the ranking of importance of the whole community (in the tripartite network). The height of the disk represents its order in each of the three different rankings (i.e. the higher the position, the more important). Lines between balls are a visual help to track changes in the rankings. H: Classification of the tripartite networks in our database according to S_{set} (similarity between the ranking of plant importance in the whole community and in the animal sets), illustrating if the ranking of plant importance in the whole community is mainly determined by only one animal set, is a mixture of the rankings of importance in the two animal sets (mixed), or does not resemble any of the rankings of importance in the two animal sets (emergent).

197 DISCUSSION

198 We gathered 44 tripartite ecological networks composed by two types of ecological interactions (including herbivory, parasitism,
 199 pollination, seed dispersal, and ant-mutualism) to investigate how different interaction types were connected to each other in
 200 tripartite ecological networks and to study how considering multiple interactions simultaneously changed our knowledge of
 201 their robustness to plant loss. While multi-interaction network data sets have been gradually appearing in the literature in the
 202 last years, only a few studies have compared several of them [13]. Such comparison allows us to reveal possible commonalities
 203 of network properties (or particularities) across the different types of tripartite networks, categorized based on the sign of the
 204 ecological interactions composing them. The rationale behind this categorization is that previous studies showed that the structure
 205 of mutualistic and antagonistic ecological networks was clearly different [31].

206 We found fundamental differences in the way the two interaction layers are connected in the different types of tripartite networks
 207 (Fig. 2A-C), possibly as a consequence of underlying biological constraints. In antagonistic-antagonistic networks, the shared
 208 species hubs are almost all connectors (meaning that generalist herbivores tend to have more parasitoids, maybe because they tend
 209 to be more abundant too, or maybe due to the sampling procedure in which parasitoids can only be reared out of the sampled
 210 herbivores), while in mutualistic-mutualistic networks most shared species hubs are not connectors (meaning that generalist plants
 211 tend not to be involved in two types of mutualism simultaneously, which hints at trade-offs in the type of interactions a given
 212 species can invest in, making it unlikely that a species can e.g. invest in attracting both pollinators and ant bodyguards [38]). The
 213 more varied behaviour of mutualistic-antagonistic networks may be related to highly complex trade-offs between herbivory and
 214 pollination [39].

215 Intuitively, we expected these differences in the connection patterns to affect the correlation between the robustness of the animal
 216 species sets in the different types of tripartite networks. These correlations (which we named ‘interdependence’) suggest that in

217 antagonistic-antagonistic networks the same plant species are important for both animal sets (in terms of secondary extinctions),
218 whereas this is not the case in mutualistic-mutualistic networks. Our results add to previous evidence showing that the benefits of
219 an intervention are not always expected to propagate throughout the whole network [21], which has implications for biodiversity
220 conservation. They highlight the relevance of knowing the type of ecological interactions involved in an ecological community
221 before planning restoration efforts, since, in the analysed networks containing mutualistic interactions, positive cascading effects
222 could only be expected if the generalist plants acted as connector nodes and were the focus of the restoration plan.

223 Surprisingly, we found that more interdependent communities are not necessarily less robust to plant losses. Rather, robustness of
224 the overall tripartite network is determined by the particular organization of each network, with degree heterogeneity playing an
225 important role, especially in antagonistic-antagonistic networks. The positive effect of degree heterogeneity on the robustness
226 of food webs and bipartite mutualistic networks was already reported in [37] and in [3, 40] (in mutualistic networks through
227 nestedness, but it was also shown that nestedness is a consequence of degree heterogeneity [41]). It is worth noting that the
228 robustness of mutualistic-mutualistic and mutualistic-antagonistic tripartite networks was found to be a combination of the
229 robustness of the two bipartite networks composing them, stressing the relevance of knowing the structure of connections in both
230 interaction layers to better quantify the robustness of the whole tripartite network. This is good news for ecologists, because it
231 means that when measuring overall robustness to plant loss it is still possible to use multiple bipartite networks (with only one
232 interaction type) and assume their effects are additive, as long as we know how plants connect them. Interestingly, looking at the
233 two interaction layers simultaneously did not result in a dramatic change in the robustness of the whole community, as already
234 reported for one of the networks in the database [24]. Nonetheless, considering the two interactions simultaneously improved the
235 quantification of the overall robustness and is crucial to identify the most important plants in a given community.

236 The approach we used to study robustness also allowed us to identify keystone species in the whole community. In most tripartite
237 networks, the ranking of plant importance in the whole community is determined by the importance of plants for both animal sets
238 (with the exception of mutualistic-mutualistic networks, that are mostly driven by one interaction layer, probably because of their
239 disproportionate size and low connection among interaction layers). In a few cases, considering the whole community could even
240 alter the picture considerably, since the ranking of plant importance in the whole community is emergent, i.e. it is not similar to
241 the ranking of importance for neither of the animal sets. This evidences that considering multiple interactions simultaneously can
242 be crucial for correctly identifying keystone species in a community.

243 The results we present here advance our knowledge of how different interactions connect ecological communities, and how that
244 affects the robustness of tripartite networks to plant losses. Taken together, our results suggest that considering multiple ecological
245 interactions simultaneously does not have a dramatic impact on the overall robustness of multi-interaction networks to plant losses.
246 However, a multi-interaction approach is crucial to know the interdependence of the robustness of the different animal sets, to
247 better gauge the overall robustness, and to correctly determine the importance of the plants at the whole community level.

248 **METHODS**

249 **Data set**

250 We gathered from the literature ecological networks which included different types of interactions. Because most studies only
251 provided two interactions simultaneously, we decided to study networks with two interaction layers. Also, we only considered
252 unweighted networks because not all studies provided interaction strengths. From all the networks we found, we only kept those
253 which had at least 5 connector nodes. In the end, our data set contains 44 unweighted networks from 6 studies (see Table 2). Each
254 network is composed of two ecological bipartite layers including mutualistic (pollination, seed-dispersal and ant-mutualism) and

255 antagonistic interactions (herbivory and parasitism). In the cases where multiple types of herbivory were present, all interactions
 256 were combined in a single herbivory layer. See SI S1 and Table S3 for more details.

Table 2: Tripartite networks included in our analyses, indicating the sign of the interactions (i.e. if the tripartite network has both mutualistic and antagonistic interactions (MA), only antagonistic interactions (AA), or only mutualistic interactions (MM)), the two ecological interactions composing the tripartite network, the number of network of each type, and the reference.

Sign	Interactions (Acronym)	Number of networks	references
MA	herbivory-pollination (H-P)	16	[6] [43] [21] [44]
	herbivory-seed dispersal (H-SD)	1	[6]
AA	herbivory-parasitism (H-Pa)	24	[45] [44]
	pollination-seed dispersal (P-SD)	2	[24] [6]
MM	pollination-ant mutualism (P-A)	1	[24]

257 Structural metrics of the connector nodes

258 We were interested in studying how the two different interactions of the tripartite networks were interconnected through the
 259 connector nodes. We used three metrics to quantify this:

- 260 • The proportion of connectors nodes in the shared set of species (C), i.e. the proportion of shared species that have links
 261 simultaneously in the two interaction layers [25].
- 262 • The proportion of shared species hubs that are connectors (H_C), i.e. the 20% of the species in the shared set of species
 263 with the highest degree that are connector nodes. Note that the degree of a node is the number of links it has with other
 264 species. We used a threshold of 20% to ensure that all networks had at least 1 “most connected” node, but the results are
 265 robust to that choice (Fig. S6B).
- 266 • The participation coefficient. This species-level metric quantifies whether the links of node i are primarily concentrated
 267 in one interaction layer or if they are well distributed among the two interaction layers [46, 47]. We quantified it as two
 268 times the ratio between the lowest degree in both interaction layers divided by the total degree of the node ($2 \frac{k_{min}}{k_{tot}}$).
 269 Hence $PC = 1$ if the links are perfectly split among the two interaction layers, and it approaches 0 as the split grows
 270 more uneven. We obtained the participation coefficient of the connector nodes (PC_C) by computing the average value
 271 over the connector nodes.

272 Quantifying robustness

273 We simulated plant loss following an established method [2, 3] and assuming bottom-up control of the animals, as justified
 274 by [48, 21]. To quantify robustness to plant loss we sequentially removed plants in a given order (the ‘extinction sequence’)
 275 keeping track of the number of secondary extinctions of animal species at each step. We considered that an animal species
 276 undergoes extinction when it has lost all its links. Note that secondary extinctions work differently in mutualistic-antagonistic
 277 and mutualistic-mutualistic networks compared to antagonistic-antagonistic networks. In the former, after removing a plant, all
 278 herbivores that no longer have resources go extinct and so do all pollinators without any resources, which means that erasing
 279 a plant may generate *simultaneous* secondary extinctions in the two animal species set (Fig. 1A). In antagonistic-antagonistic
 280 networks herbivores are the shared set of species, so when a plant disappears all herbivores without resources go extinct, which
 281 may subsequently trigger extinctions of parasitoids. In this case, removal of a plant will generate *cascading* extinctions (Fig. 1B).
 282 By plotting the proportion of remaining animal species as a function of the proportion of deleted plant species and measuring the
 283 area under the curve, we obtained the ‘robustness’ (R) (Fig. 1C). This is a standard way of measuring the efficiency of a given
 284 extinction sequence to tear down an ecological community [49, 50]: as $R \rightarrow 0$ the most impact a given extinction sequence has

285 on the community, indicating that it targets the species following the ‘correct’ order of importance for the maintenance of the
286 community.

287 Working with multipartite networks such as those in [21, 24], several robustness metrics can be measured depending on the species
288 set on which secondary extinctions are considered.

289 Here, we measured:

- 290 • the robustness of the tripartite network (R): we kept track of the proportion of remaining animal species as a function of
291 the proportion of deleted plant species, where the proportions are measured with respect to the total number of animals
292 (irrespective of their species set) and plants.
- 293 • the robustness of the two animal species sets (R_P, R_H): we measured the proportion of remaining animal species with
294 respect to the total number of animals in each species set (e.g. how many pollinators remain from the original number of
295 pollinators), and the proportion of deleted plants is measured with respect to the total number of plants in the tripartite
296 network.
- 297 • the robustness of the two bipartite networks: In this case, the tripartite network is split in two bipartite networks, on which
298 the same protocol as above is performed. These two networks are not identical to the two interaction layers because the
299 shared set of species that are not connected in a given layer are not considered in the bipartite network, which affects the
300 calculation of the robustness. We thereby obtain two robustness (R_L and R_S , respectively for the smaller and larger
301 networks, in terms of species number). Note that in antagonistic-antagonistic networks, the protocol can be performed
302 only on the herbivory network since there is no direct link between plants and parasitoids.

303 We applied 3000 random extinction sequences of plants to each of the tripartite networks in the data set, and for each extinction
304 sequence we measured the different robustness measures above. Here, results are presented for random extinction sequences but
305 results for other extinction scenarios (increasing or decreasing degree of plants) are presented in SI S6.

306 We also measured the robustness of the mutualistic-antagonistic network using an stochastic version of the co-extinction algorithm
307 [42] and weighted networks (when available) to compare with the results of the classic co-extinction algorithm (see SI 6, Fig.
308 S14 and 15). In this stochastic version, a species i will undergo a secondary extinction following the extinction of plant j with
309 a probability $P_{ij} = R_i \cdot d_{ij}$, where d_{ij} is the dependency of species i on j (interaction weight), and R_i represents the intrinsic
310 demographic dependence of species i on mutualism (we considered $R_i = 1$ for animals and $R_i = 0$ for plants, to keep the
311 bottom-up control of animals).

312 **Interdependence**

313 We measured the correlation between the robustness of the two different species sets (other than plants) in the tripartite networks,
314 hereafter called ‘interdependence’ (I) (Fig. 1D). When driving plants to extinction, a ‘high’ correlation between the robustness of
315 pollinators and herbivores implies that the same plants that are relevant for one of the species set will also be relevant in the other
316 species set [21], (i.e. sequences of plant loss that were relatively benign for pollinators were also benign for herbivores). If, on the
317 other hand, the relevant plants are not the same in the two species sets, we expect a low correlation in robustness.

318 **Plant importance rankings**

319 The importance of each plant species for the different animal sets and for the whole community (i.e. for the tripartite network)
320 was quantified based on the correlation coefficient between robustness and the position of the plant in the extinction sequence [21].
321 The rationale is that the ‘importance’ of a plant cannot be directly assessed from the number of secondary extinctions caused by

322 its loss because if lost at the start (rather than at the end) of the extinction sequence, fewer secondary extinctions are expected;
 323 however, if a plant is ‘important’, then robustness is expected to be lower when it is lost earlier in the sequence than when it is
 324 lost later. Hence, the lower the robustness to an extinction sequence, the better that extinction sequence actually resembles the
 325 importance of plants for the survival of the community. To obtain the plant importance rankings (three in total: one for each of the
 326 two interaction layers and one for the whole community), we ranked each plant species by increasing correlation between its order
 327 of appearance in extinction sequences and the corresponding robustness (i.e. plants that have a larger negative correlation are
 328 considered more important; Fig. 4A-G).

329 To assess to what extent one of the two interaction layers was driving the robustness of the whole community we measured the
 330 similarity between the importance of a plant for one animal set (for example for pollinators or herbivores) and the importance of a
 331 plant in the whole community, namely S_P or S_H . We quantified S_{set} as the square of the correlation coefficient between the
 332 ranking of the plant in each species set and the ranking in the whole community.

333 We then classified the networks in three categories: those where one interaction layer was driving the process (one S_{set} was above
 334 0.9, meaning that 90% of the variance in the importance ranking in the whole community can be traced to one of the two animal
 335 rankings), those where the ranking in the whole community was a mixture of the rankings in both animal sets (both S_{set} were
 336 between 0.5 and 0.9) and finally those where the importance ranking was emergent (both S_{set} were below 0.5, meaning that no
 337 animal set ranking was able to explain at least 50% of the ranking of importance in the whole community).

338 **Estimating tripartite robustness from networks with one interaction**

We also tested whether one can express the robustness of the whole community (R) as a combination of the robustness of the two
 independent bipartite network composing the tripartite network. To do that we performed the following linear regression:

$$R^{(est)} = a.R_L + b.R_S$$

339 where R_L and R_S are the robustness of the two bipartite networks composing the tripartite network under study, respectively the
 340 larger (i.e. with more species) and the smaller one.

341 **Multiple regressions**

342 We performed a multiple regression of interdependence and robustness based on the structural features we measured in the
 343 tripartite networks using the package *statsmodel* in Python. We selected the structural features that were more relevant for
 344 interdependence or robustness by choosing the model with a lowest AIC.

345 **Null models**

346 To assess the importance of network structure in determining a certain network feature, we compared measurements of that
 347 feature performed on empirical networks with measurements performed on randomized versions of those networks keeping some
 348 properties fixed. We used four different null-models, represented in Fig. 5, which – going from the least to the most constraining –
 349 are as follows: "1" keeps the number of species constant in each species set and the number of links constant in each interaction
 350 layer, "2" adds the constraint of keeping the degree distribution of the animal nodes constant, "3" keeps the degree distribution of
 351 animals and plants but not the total degree of the shared set species (i.e. it breaks the correlation between the degree of the shared
 352 set of species in the two interaction layers), and "4" keeps the degree of each node constant while links are reshuffled within a
 353 layer (see SI 2 for more details).

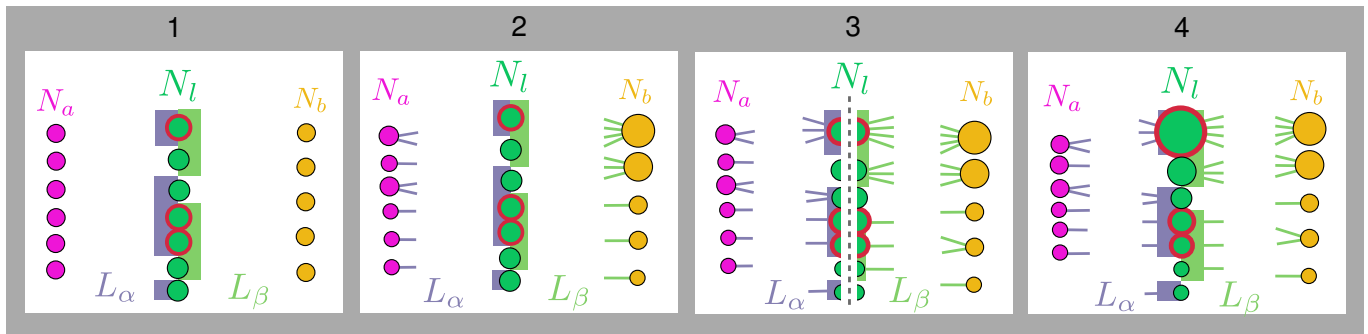


Fig 5: The 4 different null models used in this study. Each figure represents what is kept fixed in each null model, going from the less restrictive on the left, to the more restrictive on the right. N_x is the number of nodes in the species set, L_x the number of links in the interaction layer, the color of the nodes represent the different species set, the colour of the link the two different ecological interactions, the size of the node is proportional to its degree (when kept), and connector nodes are highlighted in red.

References

- 354
- 355 [1] IPBES. Global assessment report on biodiversity and ecosystem services of the Intergovernmental Science-Policy Platform
356 on Biodiversity and Ecosystem Services; 2019. Available from: <https://zenodo.org/record/3831673>.
- 357 [2] Dunne JA, Williams RJ, Martinez ND. Network structure and biodiversity loss in food webs: robustness increases with
358 connectance. *Ecology Letters*. 2002;5(4):558–567. doi:10.1046/j.1461-0248.2002.00354.x.
- 359 [3] Memmott J, Waser NM, Price MV. Tolerance of pollination networks to species extinctions. *Proceedings of the Royal
360 Society of London Series B: Biological Sciences*. 2004;271(1557):2605–2611. doi:10.1098/rspb.2004.2909.
- 361 [4] Montoya JM, Pimm SL, Solé RV. Ecological networks and their fragility. *Nature*. 2006;442(7100):259–264.
362 doi:10.1038/nature04927.
- 363 [5] Kaiser-Bunbury CN, Muff S, Memmott J, Müller CB, Caffisch A. The robustness of pollination networks to the loss of
364 species and interactions: a quantitative approach incorporating pollinator behaviour. *Ecology Letters*. 2010;13(4):442–452.
365 doi:10.1111/j.1461-0248.2009.01437.x.
- 366 [6] Melián CJ, Bascompte J, Jordano P, Krivan V. Diversity in a complex ecological network with two interaction types. *Oikos*.
367 2009;118(1):122–130. doi:10.1111/j.1600-0706.2008.16751.x.
- 368 [7] Fontaine C, Guimaraes PR, Kéfi S, Loeuille N, Memmott J, van der Putten W, et al. The ecological and evolutionary
369 implications of merging different types of networks. *Ecology Letters*. 2011;14(11):1170–1181.
- 370 [8] Kéfi S, Berlow EL, Wieters EA, Navarrete SA, Petchey OL, Wood SA, et al. More than a meal... integrating non-feeding
371 interactions into food webs. *Ecology Letters*. 2012;15(4):291–300. doi:10.1111/j.1461-0248.2011.01732.x.
- 372 [9] Kéfi S, Berlow EL, Wieters EA, Joppa LN, Wood SA, Brose U, et al. Network structure beyond food webs: mapping
373 non-trophic and trophic interactions on Chilean rocky shores. *Ecology*. 2015;96(1):291–303. doi:10.1890/13-1424.1.
- 374 [10] Kéfi S, Miele V, Wieters EA, Navarrete SA, Berlow EL. How Structured Is the Entangled Bank? The Surprisingly
375 Simple Organization of Multiplex Ecological Networks Leads to Increased Persistence and Resilience. *PLOS Biology*.
376 2016;14(8):e1002527. doi:10.1371/journal.pbio.1002527.
- 377 [11] García-Callejas D, Molowny-Horas R, Araújo MB. Multiple interactions networks: towards more realistic descriptions of
378 the web of life. *Oikos*. 2017;127(1):5–22. doi:10.1111/oik.04428.
- 379 [12] Kéfi S. Ecological Networks: from structure to dynamics. In: *Theoretical ecology: Concepts and applications*. Oxford
380 University Press; 2020. p. 143–160.

- 381 [13] Timóteo S, Albrecht J, Rumeu B, Norte AC, Traveset A, Frost CM, et al. Tripartite networks show that keystone species can
382 multitask. *Functional Ecology*. 2022;doi:10.1111/1365-2435.14206.
- 383 [14] May RM. Will a Large Complex System be Stable? *Nature*. 1972;238(5364):413–414. doi:10.1038/238413a0.
- 384 [15] Mougi A, Kondoh M. Diversity of Interaction Types and Ecological Community Stability. *Science*. 2012;337(6092):349–351.
385 doi:10.1126/science.1220529.
- 386 [16] Allesina S, Tang S. Stability criteria for complex ecosystems. *Nature*. 2012;483(7388):205–208. doi:10.1038/nature10832.
- 387 [17] Sauve AMC, Fontaine C, Thébault E. Structure-stability relationships in networks combining mutualistic and antagonistic
388 interactions. *Oikos*. 2013;123(3):378–384. doi:10.1111/j.1600-0706.2013.00743.x.
- 389 [18] Lurgi M, Montoya D, Montoya JM. The effects of space and diversity of interaction types on the stability of complex
390 ecological networks. *Theoretical Ecology*. 2015;9(1):3–13. doi:10.1007/s12080-015-0264-x.
- 391 [19] McWilliams C, Lurgi M, Montoya JM, Sauve A, Montoya D. The stability of multitrophic communities under habitat loss.
392 *Nature Communications*. 2019;10(1). doi:10.1038/s41467-019-10370-2.
- 393 [20] Hale KRS, Valdovinos FS, Martinez ND. Mutualism increases diversity, stability, and function of multiplex networks that
394 integrate pollinators into food webs. *Nature Communications*. 2020;11(1). doi:10.1038/s41467-020-15688-w.
- 395 [21] Poccock MJO, Evans DM, Memmott J. The Robustness and Restoration of a Network of Ecological Networks. *Science*.
396 2012;335(6071):973–977. doi:10.1126/science.1214915.
- 397 [22] Sauve AMC, Thébault E, Poccock MJO, Fontaine C. How plants connect pollination and herbivory networks and their
398 contribution to community stability. *Ecology*. 2016;97(4):908–917.
- 399 [23] Genrich CM, Mello MAR, Silverira FAO, Bronstein JL, Paglia AP. Duality of interaction outcomes in a plant-frugivore
400 multilayer network. *Oikos*. 2016;126:361–368. doi:doi: 10.1111/oik.03825.
- 401 [24] Dáttilo W, Lara-Rodríguez N, Jordano P, Guimarães PR, Thompson JN, Marquis RJ, et al. Unravelling Darwin’s entangled
402 bank: architecture and robustness of mutualistic networks with multiple interaction types. *Proceedings of the Royal Society*
403 *B: Biological Sciences*. 2016;283(1843):20161564. doi:10.1098/rspb.2016.1564.
- 404 [25] Astegiano J, Altermatt F, Massol F. Disentangling the co-structure of multilayer interaction networks: degree distribution
405 and module composition in two-layer bipartite networks. *Scientific Reports*. 2017;7(1). doi:10.1038/s41598-017-15811-w.
- 406 [26] Mello MAR, Felix GM, Pinheiro RBP, Muylaert RL, Geiselman C, Santana SE, et al. Insights into the assembly rules of a
407 continent-wide multilayer network. *Nature Ecology & Evolution*. 2019;3(11):1525–1532. doi:10.1038/s41559-019-1002-3.
- 408 [27] Morrison BML, Brosi BJ, Dirzo R. Agricultural intensification drives changes in hybrid network robustness by modifying
409 network structure. *Ecology Letters*. 2019;23(2):359–369. doi:10.1111/ele.13440.
- 410 [28] Evans DM, Poccock MJO, Memmott J. The robustness of a network of ecological networks to habitat loss. *Ecology Letters*.
411 2013;16(7):844–852. doi:10.1111/ele.12117.
- 412 [29] Kivela M, Arenas A, Barthelemy M, Gleeson JP, Moreno Y, Porter MA. Multilayer networks. *Journal of Complex Networks*.
413 2014;2(3):203–271. doi:10.1093/comnet/cnu016.
- 414 [30] Bianconi G. *Multilayer Networks: Structure and Function*. Paperback ed. Oxford University Press; 2023. Available from:
415 <https://lead.to/amazon/com/?op=bt&la=en&cu=usd&key=0192865544>.
- 416 [31] Thébault E, Fontaine C. Stability of Ecological Communities and the Architecture of Mutualistic and Trophic Networks.
417 *Science*. 2010;329(5993):853–856. doi:10.1126/science.1188321.

- 418 [32] Dunne JA, Williams RJ, Martinez ND. Food-web structure and network theory: The role of connectance and size. *Proceedings*
419 *of the National Academy of Sciences*. 2002;99(20):12917–12922. doi:10.1073/pnas.192407699.
- 420 [33] Staniczenko PPA, Lewis OT, Jones NS, Reed-Tsochas F. Structural dynamics and robustness of food webs. *Ecology Letters*.
421 2010;13(7):891–899. doi:10.1111/j.1461-0248.2010.01485.x.
- 422 [34] Bascompte J, Stouffer DB. The assembly and disassembly of ecological networks. *Philosophical Transactions of the Royal*
423 *Society B: Biological Sciences*. 2009;364(1524):1781–1787. doi:10.1098/rstb.2008.0226.
- 424 [35] Dallas T, Cornelius E. Co-extinction in a host-parasite network: identifying key hosts for network stability. *Scientific*
425 *Reports*. 2015;5(1). doi:10.1038/srep13185.
- 426 [36] Strona G, Lafferty KD. Environmental change makes robust ecological networks fragile. *Nature Communications*. 2016;7(1).
427 doi:10.1038/ncomms12462.
- 428 [37] Solé RV, Montoya M. Complexity and fragility in ecological networks. *Proceedings of the Royal Society of London Series*
429 *B: Biological Sciences*. 2001;268(1480):2039–2045. doi:10.1098/rspb.2001.1767.
- 430 [38] Dutton EM, Luo EY, Cembrowski AR, Shore JS, Frederickson ME. Three’s a Crowd: Trade-Offs between Attract-
431 ing Pollinators and Ant Bodyguards with Nectar Rewards in *Turnera*. *The American Naturalist*. 2016;188(1):38–51.
432 doi:10.1086/686766.
- 433 [39] Jacobsen DJ, Raguso RA. Lingering Effects of Herbivory and Plant Defenses on Pollinators. *Current Biology*.
434 2018;28(19):R1164–R1169. doi:10.1016/j.cub.2018.08.010.
- 435 [40] Burgos E, Ceva H, Perazzo RPJ, Devoto M, Medan D, Zimmermann M, et al. Why nestedness in mutualistic networks?
436 *Journal of Theoretical Biology*. 2007;249(2):307–313. doi:10.1016/j.jtbi.2007.07.030.
- 437 [41] Jonhson S, Domínguez-García V, Muñoz MA. Factors Determining Nestedness in Complex Networks. *PLoS ONE*.
438 2013;8(9):e74025. doi:10.1371/journal.pone.0074025.
- 439 [42] Vieira MC, Almeida-Neto M. A simple stochastic model for complex coextinctions in mutualistic networks: robustness
440 decreases with connectance. *Ecology Letters*. 2014;18(2):144–152. doi:10.1111/ele.12394.
- 441 [43] Shinohara N, Uchida K, Yoshida T. Contrasting effects of land-use changes on herbivory and pollination networks. *Ecology*
442 *and Evolution*. 2019;9(23):13585–13595. doi:10.1002/ece3.5814.
- 443 [44] Hackett TD, Sauve AMC, Davies N, Montoya D, Tylianakis JM, Memmott J. Reshaping our understanding of species’ roles
444 in landscape-scale networks. *Ecology Letters*. 2019;22(9):1367–1377. doi:10.1111/ele.13292.
- 445 [45] Macfadyen S, Gibson R, Polaszek A, Morris RJ, Craze PG, Planqué R, et al. Do differences in food web structure between
446 organic and conventional farms affect the ecosystem service of pest control? *Ecology Letters*. 2009;12(3):229–238.
447 doi:10.1111/j.1461-0248.2008.01279.x.
- 448 [46] Guimerà R, Amaral LAN. Cartography of complex networks: modules and universal roles. *Journal of Statistical Mechanics:*
449 *Theory and Experiment*. 2005;2005(02):P02001. doi:10.1088/1742-5468/2005/02/p02001.
- 450 [47] Battiston F, Nicosia V, Latora V. Structural measures for multiplex networks. *Physical Review E*. 2014;89(3).
451 doi:10.1103/physreve.89.032804.
- 452 [48] Scherber C, Eisenhauer N, Weisser WW, Schmid B, Voigt W, Fischer M, et al. Bottom-up effects of plant diversity on
453 multitrophic interactions in a biodiversity experiment. *Nature*. 2010;468(7323):553–556. doi:10.1038/nature09492.
- 454 [49] Allesina S, Pascual M. Googling Food Webs: Can an Eigenvector Measure Species’ Importance for Coextinctions? *PLoS*
455 *Computational Biology*. 2009;5(9):e1000494. doi:10.1371/journal.pcbi.1000494.

456 [50] Domínguez-García V, Muñoz MA. Ranking species in mutualistic networks. Scientific Reports. 2015;5(1).
457 doi:10.1038/srep08182.

458 **Supplementary figures' captions**

Fig S1: Map showing the location of the 6 studies we used to build the database of networks. The color and shape of the points represents the different types of networks, and the text the name of the site, the country, and the surname of the leading author of the study. The map has been done by the authors using package “rnatuarearth” in R.

Fig S2: The 4 different null models used in this study. Each figure represents what is kept fixed in each null model, going from the less restrictive on the left, to the more restrictive on the right. The color of the nodes represent the different species set, the colour of the link the two different ecological interactions, the size of the node is proportional to its degree (when kept), and connector nodes are highlighted in red.

Fig S3: Boxplots of the basic structural features in the empirical tripartite networks of our data-set grouped (and color coded) by the sign of the interactions involved. A: Degree Heterogeneity ($\sigma_k / \langle k \rangle$) B: Degree-degree correlations (r). C: Z-score of the degree heterogeneity ($\sigma_k / \langle k \rangle$) in null model 1. D: Z-score of the degree-degrees correlations (r) in null model “4” in the tripartite networks. In B and D horizontal grey lines mark the limits of the confidence interval of 1.96 (95%) and 2.33 (98%). Differences between groups are measured by independent t-test (** $p < 0.01$, *** $p < 0.001$, ‘ns’ not significant).

Fig S4: A: Degree heterogeneity of the empirical tripartite ecological networks ($\sigma_k / \langle k \rangle$) at the species set level B: Degree-degree correlations (r) at the interaction layer scale. C: Boxplot of the Z-score of the degree heterogeneity ($\sigma_k / \langle k \rangle$) by species set in null model 1. D: Boxplot of the Z-score of the degree-degree correlations by interaction layer in null model 4. Horizontal grey lines mark the limits of the confidence interval of 1.96 (95%) and 2.33 (98%) for rejecting the null hypothesis.

Fig S5: Cumulative degree distribution in a tripartite ‘mutualistic-antagonistic’ network of pollination-herbivory compared with its randomization in null model 1. Figures show the degree distribution for the three different sets of species, from left to right: insect herbivores, insect pollinators, plants and all nodes combined (i.e. the merged network). The continuous black line is the empirical cumulative degree distribution, with the best power law distribution fit in a dashed blue line. The continuous red line is the cumulative degree distribution in the constant “NL” ensemble (100 randomizations), and the dashed red line the best exponential fit to that distribution. The insets show the degree distribution of the empirical network (in grey) and that of 100 randomizations (in red).

Fig S6: value and Zscore of several structural properties (columns) in the 4 different null models studied (rows). Each row is marked with a letter, A for the boxplots of the empirical values (name above each panel) and B to C the Zscore in the different null models, from the less rigid (“1”) to the most conservative (“4”)

Fig S6B: Effect of the threshold considered in determining the proportion of shared set hubs that are connector nodes (H_C), i.e. the proportion of connector nodes inside the 5% (A), 10% (B) or 20% (C) of most connected shared set nodes.

Fig S7B: A: Interdependence (I) in the tripartite networks in our data set when plants are randomly driven to extinction, grouped and color coded by the signs of their interaction layers. B to E: Z-score of interdependence (I) in the different null models. The horizontal lines represent the Z-score values associated with a 95% and 98% confidence interval, $z=1.96$ and $z=2.33$ respectively.

Fig S8B: Correlation of interdependence (I) vs the structural features studied: A) degree heterogeneity ($\sigma_k / \langle k \rangle$), B) degree heterogeneity of the shared set species ($\sigma_k / \langle k \rangle_{LS}$), C) degree-degree correlations (r), D) proportion of connector nodes inside the shared set (C), E) Proportion of shared set hubs that are connectors (H_C), and F) average participation ratio of the connector nodes (PR_C) on the tripartite networks in our data set. The colors of the points represent the different types of tripartite networks according to the sign of their interaction layers (see legend). The values in the lower right side are the Pearson correlation coefficients considering only MM and MA networks (blue), only AA networks (violet).

Fig S9B: Robustness of the tripartite networks in our data set in three different extinction scenarios: A) random plant extinction (RND), B) extinction of plants by decreasing degree (DD), i.e. generalist plants are attacked first, and C) extinction of plants by increasing degree (ID), i.e. specialist plants are attacked first.

Fig S10B: A: Robustness (R) of the tripartite networks in our data set when plants are randomly driven to extinction, grouped and color coded by the sign of the interaction layers. B to E: Z-score of R in the different null models. The horizontal lines represent the Z-score values associated with a 95% and 98% confidence interval, $z=1.96$ and $z=2.33$ respectively.

Fig S11B: Upper panel: correlation of the Robustness of the whole merged network (R) with A) the Robustness of the larger interaction layer (R_L), B) the Robustness of the smaller interaction layer (R_S), and C) with the best estimation of Robustness ($R_{(est)}$) as a composition of the Robustness of the two interaction layers. Lower panel: Comparison of the Robustness of the tripartite networks (in blue) with that of the two interaction layers composing them (R_L in green and R_S in red) and with the best estimated Robustness as a composition of the Robustness of the two interaction layers (formula shown).

Fig S12B: Effect of degree heterogeneity ($\sigma_k / \langle k \rangle$) and degree-degree correlations (r) on the Robustness (R) of the tripartite networks in our data set. The colors of the point represent the different types of tripartite networks according to the sign of their interaction layers (violet AA, green MA, and blue MM). The values in the upper right side are the Pearson correlation coefficients considering only MM and MA networks (blue), and only AA networks (violet). The lines represents the best linear regression for each of the correlations.

Fig S13B: Effect of structural features of the shared set on the Robustness (R) of the tripartite networks in our data set. The color of the points represents the different types of tripartite networks according to the sign of their interaction layers (violet AA, green MA, and blue MM). The values on the side are the Pearson correlation coefficients considering only MM and MA networks (blue), and only AA networks (violet). The lines represent the best linear regression for each of the correlations.

Fig S14B: Comparative of results between the stochastic co-extinction algorithm in weighted networks (when available) with the results of the classic co-extinction algorithm used in the main text, for Robustness (left) and Interdependence (right).

Fig S15B: Comparative of the estimated robustness of the mutualistic-antagonistic tripartite networks as a combination of the robustness of the larger and smaller bipartite networks using the basic algorithm (left, as in the main text) and using the stochastic version of the algorithm and weighted networks (right).

Fig S16B: Upper row: Distribution of plant importance values in the network used as example in Fig. 4. The histograms depict the amount of plants with a given importance in the whole community (left), with a given importance for pollinators (center) and for herbivores (right). Lower row: Ranking of importance (based on importance) vs importance. Inset values show the spearman rank correlation coefficient (ρ) and pearson correlation coefficient (r).

Fig S17B: Distribution of plant importance values considering all the communities together. The histograms depict the amount of plants with a given importance in the whole community (left), with a given importance for animals connected through positive interactions -pollination or seed dispersion- (center) and for animals connected through negative interactions -herbivory- (right).

459 **Supplementary tables' captions**

Table S1: Multipartite networks included in our analyses, indicating the sign of the interactions (i.e. if the tripartite network has both mutualistic and antagonistic interactions (MA), only antagonistic interactions (AA), or only mutualistic interactions (MM)), the two ecological interactions composing the tripartite network, the number of network of each type, and the reference.

Table S2: Tripartite empirical networks composing the data-set. Columns indicate from left to right: the name of the network (based on the publication it comes from), the sign of the two interactions composing the network (mutualism-mutualism: MM, antagonism-antagonism: AA, mutualism-antagonism: MA), the name of the two interactions layers composing the network (ant-mutualism: A, herbivory: H, parasitoidism: Pa, pollination: P, and seed-dispersal: SD), the degree heterogeneity ($\frac{\sigma_k}{\langle k \rangle}$), the degree-degree correlations (r), Z-score of degree heterogeneity in the "NL" null model ($Z_{NL}(\frac{\sigma_k}{\langle k \rangle})$), Z-score of degree-degree correlations in null model "4" ($Z_4(r)$), and the reference to the study from which the data was gathered.