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Scaling up our understanding of tipping points

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Abstract

Anthropogenic activities are increasingly affecting ecosystems across the globe. Meanwhile, empirical and theoretical evidence suggest that natural systems can exhibit abrupt collapses in response to incremental increases in the stressors, sometimes with dramatic ecological and economic consequences. These catastrophic shifts are faster and larger than expected from the changes in the stressors and happen once a tipping point is crossed. The primary mechanisms that drive ecosystem responses to perturbations lie in their architecture of relationships, i.e. how species interact with each other and with the physical environment and the spatial structure of the environment. Nonetheless, existing theoretical work on catastrophic shifts has so far largely focused on relatively simple systems that have either few species and/or no spatial structure. This work has laid a critical foundation for understanding how abrupt responses to incremental stressors are possible, but it remains difficult to predict (much less manage) where or when they are most likely to occur in more complex real-world settings. Here, we discuss how scaling up our investigations of catastrophic shifts from simple to more complex - species rich and spatially structured – systems could contribute to expand our understanding of how nature works and improve our ability to anticipate the effects of global change on ecological systems.

1 Introduction

The Intergovernmental Science-Policy Platform on Biodiversity and Ecosystem Services (IPBES) 31 recently estimated that approximately one million species are currently threatened with extinc-32 tion. Meanwhile, the 2021 Intergovernmental Panel on Climate Change (IPCC) report highlights 33 that, even under the most optimistic emissions scenario, we still have decades (at least) of warming 34 ahead. As these and other stressors gradually increase, there is growing concern that many natural 35 ecosystems may exhibit sudden, abrupt, and persistent collapses. While extreme environmental 36 events, such as fires or hurricanes, can have significant, and sometimes irreversible, consequences 37 for ecosystems, we are interested here in ecosystem responses that are much larger and faster that 38 one would intuit from the driver of change. Evidence for these striking, non-linear ecosystem re-39 sponses to incremental changes in conditions has accumulated from a range of experimental and 40 natural systems, including desertification, eutrophication, the degradation of coral reefs and the 41 collapses of fisheries [1-11]. One of the most pressing scientific challenges of our time is to predict 42 where and when this possibility is likely to be realized [12]. 43

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Theoretically, these phenomena are described as 'catastrophic shifts' or 'regime shifts', which have 45 been well studied using classical ecological models [3, 10, 13–15]. These models show that there 46 are different ways in which a system may respond to gradual changes in external conditions. The 47 response can be gradual, from something almost linear to something that can be highly nonlinear 48 or even abrupt [16]. These latter responses may happen when an ecosystem has more than one 49 possible stable state for the same range of conditions. The ecosystem can then be propelled from 50 its current state to an alternative one, resulting in discrete, surprising changes that can be more 51 or less easily reversed. Such catastrophic shifts between ecosystem states can happen because of a 52 perturbation of the ecosystem state itself or when an environmental condition is changed beyond 53 a threshold value, or 'tipping point'. 54

Our current theory of these phenomena is limited to 'simple' models that either assume a) very few 56 species, or b) unrealistic spatial settings. In other words, theoretical work on catastrophic shifts has 57 so far largely focused on species-poor systems [3, 10, 13–15, 17] or on species-rich systems in which 58 species interact with each other with a single interaction type (e.g. feeding or pollination) [18, 19]. 59 Further, the vast majority of work has focused on isolated systems ignoring the spatial structure 60 of the landscapes in which ecosystems are embedded. More specifically, there is a good under-61 standing of how alternative stable states can emerge in simple settings but it is less understood 62 how these dynamics occur when embedded in systems that have more realistic complex diversity 63 in species, interaction types, and spatial structure. This means that, despite clear evidence that 64 alternative stable states, catastrophic shifts, and ecological tipping points are possible in a range of ecosystems, we have no strong theory to predict which natural ecosystems on earth are most vulnerable to exhibiting rapid collapses in the context of global changes. This poor understanding of the conditions under which complex ecological systems may exhibit alternative stable states, and thereby catastrophic shifts and tipping points, strongly limits our ability to anticipate and manage natural systems.

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Here, we argue that – if we want to predict which ecosystems are vulnerable and prioritize 'at 72 risk' ecosystems before they collapse – current ecological theory on catastrophic shifts and tipping 73 points needs to explicitly address species rich and spatially complex systems. In particular, we 74 need to investigate if, when and how scaling up the current theory from simple – species poor and 75 local – to more complex systems leads to possible emergent alternative ecosystem states as species, 76 interactions or spatial complexity are incorporated. We review the current state of theoretical 77 work and illustrate that these are pieces of a puzzle that still need to be assembled toward a more 78 comprehensive framework of how complex systems can exhibit abrupt transitions to equip our 79 societies to face the challenges to come. 80

2 Scale up to more complex systems: species diversity

The majority of theoretical studies on catastrophic shifts has focused on 'simple' models, that 82 is, typically either considering a subset of the species of a community while ignoring the rest or 83 lumping species into groups [3, 10, 13, 15, 20]. These models have highlighted early on the impor-84 tance of reinforcing feedbacks for the emergence of alternative stable states and therefore possible 85 catastrophic shifts between them [14, 15]. A reinforcing feedback occurs when species have positive 86 effects on themselves. For example, in drylands, where water is often the most limiting resource, 87 vegetation can improve local conditions for itself by improving water availability and thus creating 88 a reinforcing feedback: when vegetation growth is increased, vegetation biomass increases, and 89 more water becomes available for plants to grow. This stabilizes an ecosystem state where veg-90 etation is present. However, the feedback loop can work the other way around: if vegetation is 91 lost, for instance, due to perturbations, so that the amount of vegetation falls below a threshold, 92 local conditions will degrade and prevent new vegetation to settle. The loss of vegetation rein-93 forces itself, thereby stabilizing another possible ecosystem state with low or even no vegetation. 94 In these models, a reinforcing feedback loop, often between a biotic and an abiotic component of 95 the ecosystem, creates the possibility for alternative stable states, and therefore for tipping points 96 and catastrophic shifts of the ecosystem. 97

Understanding the emergence of alternative stable states in relatively simple systems is very useful, 99 especially for ecosystems in which a few strongly interacting species dominate, such as lakes or 100 drylands [10] but what about other ecosystems? Ecological communities in nature are composed 101 of the many species they host and of the network of interactions those species have with each 102 other and with the environment. This raises the question of whether and how increased biotic 103 complexity can allow or prevent the emergence of alternative stable states. Predicting the response 104 of species-richer systems to changes is difficult because the response of the whole system does 105 not only depend on the properties of the entities (the species) but also on the presence, type 106 and strength of the interactions between species and on the way they are arranged, i.e. the 107 structure of the interaction network [21–23]. Addressing the question of how the species interaction 108 network impacts catastrophic shifts therefore requires scaling up our understanding from simple to 109 species-richer systems. Although stability in general, and possible catastrophic shifts in particular, 110 have been understudied in ecological networks [24, 25], some previous works have addressed those 111 questions; We present their insights in what follows. 112

Large random species interaction networks

Gilpin and Case [26] already discussed in the 70s the theoretical possibility of several stable equilibria in multispecies competition models. Going further, using a modified version of a multi-species Lotka-Volterra model, van Nes and Scheffer [27] confirmed that alternative stable states commonly arise in complex communities that are randomly generated, in agreement with later studies [28]. This implies that such complex communities could occasionally exhibit abrupt responses to gradual environmental changes.

Structured species interaction networks

One shortcoming of the above discussed studies is that they do not take the structure of species 121 interaction networks into account. Decades of work has shown that real ecological networks are 122 far from random and that their structure matters for community dynamics, and in particular 123 for stability (e.g. [21, 23, 29]). For example, food webs (i.e., ecological networks composed of 124 feeding interactions) tend to be organized into compartments [29], that is, groups of species that 125 interact preferentially with each other and less with the rest of the network. This organization 126 in compartments contributes to the reduction of the spread of perturbations by containing them 127 within the compartments [29]. These approaches have, however, largely overlooked the possibility of 128 the emergence of alternative stable states at the network scale (but see examples below). Therefore, 129 until recently, little was know about the species interaction network configurations that favor 130

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the emergence of alternative stable states and ecologically significant transitions in species-rich ¹³¹ ecological networks. ¹³²

Hints on underlying mechanisms

As previously mentioned, mathematical models have highlighted early on the importance of reinforcing feedbacks for the emergence of alternative stable states [14, 15]. In a complex network of species interactions, feedbacks are typically composed of a succession of interactions between species as well as between species and their abiotic environment, and the type of the feedback emerges from the interplay between all direct and indirect effects [30].

To investigate the mechanisms leading to alternative stable states in ecological networks, Karatayev et al. [19] studied a multispecies model of consumer-resource interactions with different types of feedbacks: specialized (species-specific feedbacks that occur between a given pair of species) or aggregate (which do not depend on the identity of the species). 142

They show that alternative stable states are more prevalent under aggregate than specialized feedbacks. Indeed, for species-specific feedbacks, variability among species may desynchronize their dynamics and potentially dissipate the feedback. Conversely, if all species experience the same mechanism (with e.g. the same threshold), this can lead to a synchronization of the feedbacks and the emergence of global tipping points.

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In the case of species-specific feedbacks, a model of mutualistic plant-pollinator communities 150 showed that pollinator populations can either collapse partially ('partial collapse') or simulta-151 neously (all pollinator populations collapse simultaneously; 'whole community collapse') as the 152 driver of population decline increases beyond a threshold [18]. The probability of having a single 153 whole community collapse, instead of several partial collapses, is strongly influenced by the struc-154 ture of the mutualistic networks (connectance and/or nestedness). Indeed, pollinators have direct 155 negative effects on each other through competition, and they also can have indirect positive effects 156 on each other when they pollinate the same plant. Pollinators that depend on the same plant 157 species have increasingly strong positive net effects on each other as stress increases. Increased 158 connectance and nestedness increase the fraction of mutualistic partners shared by pollinators and 159 thereby favor the emergence of reinforcing feedback at the network scale. This means that in highly 160 connected and/or nested networks, pollinators eventually collapse simultaneously as the stress level 161 goes beyond a threshold. 162

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In conclusion, these studies have provided key elements toward understanding the type of ecological 164

mechanisms, and more specifically the structure of the species interaction networks, that can transform pairwise interactions into whole-system feedback that can promote the emergence of alternative stable states. However, how those results expand to other systems, other interaction types as well as to multi-interaction ecological networks is unknown. Therefore, we still lack a more general understanding of the network configurations that can lead to alternative stable states at the system scale, and whether transitions between these alternative stable states correspond to partial or whole collapses.

3 Scale up to more complex systems: space

Most studies discussed up to here have considered ecological systems in isolation, that is, only at 173 one given location in space. However, ecosystems are typically embedded in complex landscapes 174 within which exchanges of energy, materials, and organisms occur [31, 32]. Work on human-made 175 networks (e.g., electricity and internet networks) has shown that the spread of failures can have 176 drastically different, and even opposite, effects in isolated rather than in interdependent networks 177 [33, 34]. In the same vein, a regime shift in one ecosystem of a landscape composed of several, 178 connected ecosystems could trigger shifts in others and possibly lead to a cascade of shifts. If 179 ecosystems exhibiting alternative stable states locally are spatially connected by the movements of 180 species and fluxes of matter, what are the dynamical properties that emerge at the broad spatial 181 scale, or 'meta'-scale? Are global tipping points possible in such spatially connected systems? 182 Recent studies discussed in what follows have started exploring these questions. Note that we are 183 here not interested in the links between spatial structure and tipping points within ecosystems 184 (e.g. [17, 35, 36]) but in the role of the spatial connectivity between ecosystems. 185

Implicit space

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Investigating the consequences of spatial flows between local systems for functioning at the meta-187 scale has been the focus of the metapopulation, metacommunity and metaecosystem theoretical 188 frameworks [31]. Space has been found to be stabilizing, for instance in the context of stochastic 189 metapopulations where times to extinctions increase due to the rescue of extinct patches via dis-190 persal [37]. Similar results have been found in classical metacommunity studies (e.g., the patch 191 dynamics paradigm; [38]) and have been scaled up to entire metaecosystems (e.g. [31, 32]). Re-192 garding alternative stable states more specifically, classical work on metapopulation dynamics of 193 the Glanville fritillary has shown that multiple equilibria can exist in connected ecological systems 194 when the effects of dispersal on local dynamics are taken into account [39, 40]. However, these 195 models are spatially implicit and therefore do not take into account the role of the way patches 196 are located in space and connected to others. How the spatial structure of the patches themselves ¹⁹⁷ affect the possibility of alternative stable states at the meta-scale remains a largely unresolved ¹⁹⁸ question. ¹⁹⁹

Spatially explicit meta-ecosystems

In analogy to species interaction network topology, one can ask how spatial network topology 201 affects the spread of shifts in a setting where ecosystems are connected to each other in space 202 by the movements of species and fluxes of matter. To answer this question, model studies have 203 included spatial structure explicitly under two paradigms: i) models of continuous space are used to 204 describe systems that lack clear spatial patches and where the habitat can be considered relatively 205 homogeneous (see the example of Lake Veluwe in [41]), and ii) models of discrete space that are 206 better suited to discontinuous habitats with clear patches connected by dispersal of individuals 207 and flow of resources. 208

Continuous space

Let's imagine an ecosystem which has two stable states locally. Studies have investigated how 210 local shifts can spread when such ecosystems are connected in continuous space. They have found 211 that the most stable of the two stable states often dominates over the other: a local shift from 212 the dominant state recovers, while shifts to the dominant state spread through space at a constant 213 speed ('travelling wave') [41, 42]. The whole system is generally not expected to exhibit 'spatial 214 bistability' (i.e. the coexistence of the two stable states in space) in a continuous habitat. Because 215 of that, the whole system is expected to exhibit sharp transitions between a fully occupied state 216 and an empty state (i.e. 'regional bistability'). Moreover, hysteresis is expected to be largely 217 reduced at the landscape scale [41]. 218

Discrete space

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Because the spatially continuous view can be difficult to analyze mathematically and is often 220 more appropriate for relatively small spatial scales, an option is to consider landscapes as discrete 221 patches connected to each other by fluxes. Such discrete systems, modeled in the form of lines or 222 grids, have been found to behave in a similar way as continuous ones when dispersal is strong: a 223 local shift to the dominant state spreads from patch to patch in a domino effect [42, 43]. Here 224 again, regional scale hysteresis mostly disappears if dispersal is strong. However, as dispersal [42] 225 or flushing rates in aquatic systems [43] decrease, the speed of the travelling wave decreases and 226 can eventually come to a halt (so-called 'pinning' phenomenon). In that case, alternative stable 227 states can coexist in a landscape for intermediate environmental conditions, which can smooth the 228

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transition at the regional scale.

Both in discrete and continuous space, several mechanisms have been found to smooth the landscapescale transitions between the fully occupied and the empty state. Demographic stochasticity [44], 232 spatial heterogeneity in environmental conditions [44, 45] or in dispersal rates [41] facilitate spatial 233 bistability and smooth large scale transitions. 234

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We thus have a relatively good understanding of the whole-system dynamics of spatially connected ²²⁶ ecosystems with locally bistable dynamics when the connections between the local systems are ²³⁷ simplified to be along a line or a 2-D grid. The whole system will then only have homogeneous ²³⁸ stable states (fully empty and fully occupied) and exhibit sharp transitions between them. However, ²³⁹ regional bistability and hysteresis being largely reduced, these shifts cannot usually be called ²⁴⁰ catastrophic shifts. ²⁴¹

Increasing spatial complexity

As previously mentioned, for practical reasons, studies have so far focused on simplified settings 243 in terms of how ecosystems are connected to each other in space. Real landscapes, however, are 244 discontinuous, heterogeneous and have emergent properties — e.q., terrestrial populations usually 245 show emergent modularity [46] and riverine systems are dendritic [47]. Importantly, the structure 246 of real landscapes have been shown to affect ecological outcomes [48–50]. Ignoring these impacts is 247 an important shortcoming of the current state of the literature as the properties of such habitats 248 — e.g. the heterogeneity in connectivity resulting from local barriers to dispersal — may change 240 how local bistability affects regional scale dynamics and equilibria [41, 45]. 250

Recently, Saade et al. [51] have started to tackle this issue by investigating the landscape-scale 252 stability of more complex discrete landscapes. Comparing the commonly assumed linear (1-D) and 253 grid-like (2-D) landscapes as well as more realistic networks (reflecting riverine and terrestrial sys-254 tems), they find that local shifts can induce a landscape-scale shift through a domino effect across 255 all landscapes structures. However, the position of landscape-scale tipping points and the extent of 256 hysteresis (*i.e.*, the distance between the degradation and restoration tipping points) is very sensi-257 tive to the landscape structure. The reduction of hysteresis documented by Keitt et al. [42] and Hilt 258 et al. [43] is restricted only to landscapes with very low connectivity such as 1-dimensional linear 259 landscapes and dendritic (riverine) networks. Landscapes with a higher connectivity (2-D grids and 260 terrestrial systems) exhibit landscape-scale bistability with a pronounced hysteresis. Moreover, the 261 commonly used linear (1-D) and grid-like (2-D) system do not necessarily reflect the behaviour of 262

more realistic landscapes structures, as 1-D linear systems consistently exhibit smaller hysteresis 263 and 2-D grids consistently exhibit larger hysteresis than more realistic networks. This result may 264 be explained by the fractal dimension of the network. Since the position of landscape-scale tipping 265 points dictates how easy it is to induce a landscape degradation or restoration, we should be wary 266 of overly simplified spatial structure and, when possible, should prefer realistic network structures 267 to study the stability of spatial systems. In the future, one way to link more general theoretical 268 results to real biological systems could be to study explicitly the impact of network properties 269 (e.g., connectivity, modularity, link distribution...) on landscape-scale stability. 270

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In conclusion, ecological systems quasi universally show spatial structure. Existing work has shown 272 that alternative stable states are possible in spatially structured systems. Yet, because most models 273 so far rely on simplified descriptions of space (either implicit, 1-D or grid-like), we don't know how 274 the results may change in the case of more realistic spatial structure. A more explicit inclusion of 275 spatial complexity, meaning the spatial network topology, could help understand when and how 276 this level of complexity contributes to alternative stable states and possible shifts at the meta-277 scale. It is important to note, however, that studying systems at increasingly larger spatial scales 278 raises the question of whether the environmental conditions are still relatively homogeneous at the 279 scale considered. The existence of alternative stable states, indeed, requires that such states (e.g., 280 species configurations) exist under the same set of environmental conditions, an assumption that 281 becomes less and less likely to hold as scale increases. 282

4 Opportunities to advance understanding

Both, in the context of biotic interactions and spatial structure, we have argued so far that the cur-284 rent ecological theory of alternative stable states and catastrophic shifts lacks aspects of complexity 285 which are worth exploring. At the same time, complex models become very quickly intractable as 286 the number of dimensions and parameters grow, making it difficult to adequately describe systems 287 with more than three or four interacting components. Yet, perhaps ironically, more complex mod-288 els can reveal emergent phenomena which constrain system behavior and simplify predictions in 289 ways that are impossible to see with simpler models [52-54]. Therefore, identifying key dimensions 290 of ecological complexity that are lacking in the current theory of tipping points may reveal oppor-291 tunities to more simply predict when and where they are likely to occur in real-world ecosystems. 292 We next identify a few promising directions in which progress could be made to help get a better 293 understanding of the emergence of shifts in complex ecological systems. 294

Incorporating the diversity of interaction types

Ecological network studies have so far typically investigated a single interaction at a time, e.g. food 296 webs composed of feeding interactions [21, 22] or mutualistic networks composed of plant-pollinator 297 interactions [23], for example. Yet, in natural communities, species depend on and influence each 298 other in multiple ways: they eat each other, compete for nutrients, light and space, fight for refuges, 299 habitats or common prey, and provide habitat and protection to others [30, 55]. In the last few 300 years, there has been increasing recognition about the need to investigate the consequences of the 301 coexistence of multiple interaction types in nature, i.e. describe ecological communities as 'multi-302 layer' ecological networks, in which each layer contains information about one type of interactions 303 that link the species of the community [30, 56, 57]. 304

Dynamical models have started investigating such multi-interactions ecological networks [30, 56, 58, 59], showing that the presence, the relative abundance and the structure of the different interaction types can affect our fundamental understanding of how ecological systems work (e.g. [30, 56, 58]). Altogether, these studies suggest that incorporating the diversity of interactions in ecological theory is a worthwhile effort [57]. There is, however, currently no work that we know of about alternative stable states and their implications for the responses of multi-layer ecological systems to environmental changes.

Measuring the resilience of complex systems

The concepts of stability and resilience have received considerable attention in the ecological literature for decades. As originally defined by Holling in 1973, the term resilience was introduced to specifically refer to ecological systems that can exhibit alternative stable states [14]. The concept, however, has faced difficulties in estimating it in real systems [60].

In the same was as the stability of ecological communities has been estimated in a variety of ways 319 using many different metrics [24, 25, 61], this is also true for resilience and more generally for met-320 rics related to catastrophic shifts [62]. This multiplicity of metrics used to quantify the concepts of 321 stability and resilience in ecological studies raises the question of how the different metrics relate 322 to each other [24]. A better understanding of these relationships could help us identify how many 323 and which metrics need to be measured in natural systems [24, 25, 63]. This is especially true for 324 metrics related to abrupt transitions for which we need to understand the relations between each 325 other but also with other stability metrics [62]. 326

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As stressed in Carpenter et al. [60], when thinking about the resilience of complex systems, it is not

only a matter of deciding on a metric to measure it but also of specifying the 'resilience of what to what'. It is indeed important to clarify on what system state resilience is measured (resilience of what) as well as the perturbations that are considered (resilience to what). Furthermore, one needs to decide on a time period and particular spatial scale since resilience can vary depending on the spatial and temporal scale at which it is measured.

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Altogether, despite recent progress, there are still no clear guidelines on how to measure (with which metrics, at what scale and at what level of organization, e.g. species or community) the overall stability and resilience of ecological systems [24, 25]. The concept of resilience has been very useful so far in thinking about ecological systems and their responses to changes, but a more measurable, operational definition of resilience could provide important insights and practical applications in better understanding, quantifying and eventually possibly mapping ecosystem resilience [60].

Predicting the resilience of complex systems: early warning signals

Important changes in stability, especially abrupt ones, are notably difficult to predict. Generic 342 indicators (so-called 'early-warning signals') have been proposed in the literature [64]. They are 343 based on a phenomenon referred to as critical slowing down, which states that a disturbed system 344 needs more time to recover when it is closer to a shift, i.e. it becomes slower and therefore less 345 resilient [65]. Signatures of this slowing down can be detected in the temporal or spatial dynam-346 ics of ecosystems (increase in recovery time, temporal variance, autocorrelation and skewness of 347 a variable used to quantify the ecosystem's state, such as total biomass) [64]. Therefore, using 348 temporal or spatial data of a given ecosystem, we should be able to detect whether the system is 340 losing resilience. 350

The phenomenon of critical slowing down seems to be universal and occurs when any dynamical system approaches a transition. In fact, mathematically, the early warning signals are expected to work when a dynamical system approaches any kind of bifurcation, even when there is a change in stability that is not technically a bifurcation [16]. These signals are therefore generic, as they operate in principle independently of the specific mechanism responsible for the change of resilience, making their potential scope of application very broad. While their generality makes them very promising, it also means that they are not specific to catastrophic shifts [16].

Early-warning signals have been shown to successfully announce a loss of resilience in different models and in controlled laboratory conditions (e.g. [6, 7, 66]). Although more limited in numbers, manipulative field experiments have also demonstrated their ability to test model predictions, 360

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Success in real-world observational time series data has, however, been more mitigated [68, 69]. 365 This highlights that we need to better understand the conditions under which early-warning signals 366 work or fail. In particular, because theoretical work on early-warning signals has mainly been based 367 on local models with one or a few species so far, it is unclear how reliable these indicators are in 368 systems with multiple species, locations in space and with multiple interaction types. It is also 369 necessary to understand at which organizational level these indicators should be measured (species 370 or community) and on which variable (e.g. biomass, species number). Investigating this question 371 formally in model ecological communities, Patterson et al. [70] show that the reliability of early-372 warning signals depends on the species observed as well as on the type of ecological interaction 373 considered. For example, the relevant species to monitor to maximize early-warning signal detection 374 is not the same in a consumer-ressource than in a mutualistic system. Another important limitation 375 of the current ecological theory of catastrophic shifts is that it describes ecosystems at equilibria, 376 which is probably rarely verified in nature [71, 72]. Regime shifts can be caused by non-equilibrium 377 phenomena and early-warning signals are not expected to apply in these cases [72, 73]. 378

Simplifying complexity

Adding new aspects of complexity in current theory comes with practical challenges but there are 380 promising avenues to overcome this challenge. Indeed, recent papers have tried to overcome this 381 issue by describing high-dimensional systems using a few quantities only. For instance, Gao et al. 382 [54] suggested a method of using mean-field approximations to reduce n-dimensional mutualistic 383 networks to a single dimension (the effective average density) and a single control parameter de-384 scribing structure of the network. This allows getting analytical results on these systems, such as 385 predicting the equilibria from an aggregate of three network metrics (link density, heterogeneity 386 and symmetry). In particular they show how to apply this framework to real ecological networks 387 (plant-pollinator) and gene regulatory networks (from E. coli and S. cerevisiae), and how to detect 388 if network modifications will trigger a shift. 389

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More recent papers generalized this approach by using spectral analysis of the interaction network ³⁹¹ to make it applicable to non-random networks [74] and to heterogeneous networks (*i.e.*, where ³⁹² different nodes can have different types of dynamics), providing us with very promising tools for ³⁹³ the study of mutualistic interaction networks. ³⁹⁴

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Barbier et al. [53] proposed another reduction method inspired from statistical physics to deal

with mutualistic networks, but also with predation and competition networks. It uses aggregate 397 parameters of the interaction network (*i.e.*, the average and variance of the species growth rates, 398 carrying capacities and interaction strengths) to summarize a *n*-dimensional system in a single 300 stochastic dynamical equation. This single equation predicts aggregate properties of the system, 400 such as the total biomass, the fraction of surviving species, the Simpson's diversity or the temporal 401 variability of the system. While it is best suited for disordered networks (e.g., networks where402 species associations are seemingly random) with a single interaction type, the authors suggest that 403 extensions of the approach are possible to deal with more structured networks and more than one 404 interaction type. They illustrate this by reducing a strongly compartmentalized bipartite networks 405 (*i.e.*, with competition interactions within and mutualistic interactions between compartments) to 406 two stochastic equations, each corresponding to one compartment. While this method still requires 407 further investigations to determine its applicability, it is extremely promising to simplify the study 408 of complex systems, especially if a more general framework emerges to deal with multi-interaction 409 networks. 410

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Another issue when studying ecological systems is that they are often 'networks of networks': a 412 community of species (interaction network) lives in and connects through space a set of inhab-413 itable locations (spatial network). The interactions between these two layers can give rise to 414 complex dynamics such as oscillations and pattern formation [75, 76] analogous to Turing-patterns 415 in reaction-diffusion. These dynamics are unexpected from the study of the interaction network 416 alone and can lead to alternative stable states with abrupt transitions between them when the 417 spatial network topology is altered [75]. Brechtel et al. [77] proposed a method that uses the mas-418 ter stability function [78] of the spatial network to study the stability of such systems. It allows 419 to determine whether a stable state determined from the interaction network alone can form an 420 homogeneous stable state in a given spatial network or if complex pattern formation will arise. It 421 also allows to determine if a modification of the spatial network can result in pattern formation. 422 This method could prove very useful in studying the stability patterns arising of 'networks of net-423 works', and the author suggest other types of systems where it could be applied such as cell biology 424 (where gene regulatory networks determining the state of each cell is nested in the cell interaction 425 network in a tissue). 426

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Innovative methods are also available to reconstruct the whole landscape of possible states of a 428 complex system [79, 80]. Chemical organization theory relies on discretizing the description of 429 species interactions in a way that resembles chemical reactions. Once this is done, the approach 430 provides a set of analytical tools which allows to reconstruct the whole landscape of possible states 431 of the systems and transitions among these states. Obtaining this landscape is very powerful ⁴³² because it provides an overview of all possible states (and their species composition) but also of ⁴³³ the possible (and the probability) of transitions between the different states. This shows how a ⁴³⁴ given community is most likely to change when subject to perturbations but also the paths it will ⁴³⁵ take to get there. ⁴³⁶

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Integrating different approaches: beyond modeling

We have so far focused this section on modeling, but progress in understanding catastrophic shifts 439 also requires a better integration of models with observational data and experiments. Although pio-440 neering studies have reported very promising results (see previous section 'Predicting the resilience 441 of complex systems: early warning signals'; [6, 7, 66, 68, 69]), experimental work and particularly 442 manipulative field experiments are largely absent from the catastrophic shifts literature. They, 443 however, have a key role to play in testing predictions from the theory, assess early-warning signals 444 in real world conditions and foster our understanding of the mechanisms behind the reinforcing 445 feedbacks. 446

In conclusion, we have identified key points of improvement of current catastrophic shift theory 448 in ecology. These are mainly due to the lack of biotic and spatial complexity considered in clas-449 sical models. We advocate for a multilayer-network perspective, taking into account networks of 450 networks ranging from species interactions to networks of habitats in landscapes. This will allow 451 us to take into account network topology as an important modulator of catastrophic shifts and 452 to understand how disturbances may propagate in realistically complex landscapes of realistically 453 complex ecosystems. While this step forward intrinsically embraces complexity, this does not need 454 to come at a cost of tractability of underlying models. Indeed, highly promising avenues of di-455 mensionality reduction have the potential to facilitate models and theory development. Of course, 456 this list of improvement points is not exhaustive. Taking the influence of multiple stressors and 457 their interactions [81] as well as the importance of stochasticity and transient dynamics [71] into 458 account are examples of further research avenues. 459

5 Conclusion

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Despite the urgency and magnitude of global change threats to humanity's life support system, there is no synthetic, empirically-grounded body of ecological theory to predict which ecosystems and ecological communities are more likely to abruptly shifts in response to upcoming changes. We need to fill that gap. Mathematical models can help us identify and understand the conditions 461

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under which species rich, interaction rich and spatially structured ecosystems can exhibit alternative stable states possibly leading to ecologically important catastrophic shifts between them.

Addressing that question is however not trivial. Indeed, the current state of the literature only 468 provides pieces of the puzzle. On the one hand, alternative stable states seem to be a robust 469 phenomenon in large random ecological networks [27]. On the other hand, studies have showed 470 that the structure of real ecological networks deviate from random ones in ways that tends to 471 make them more robust than expected by chance. These studies have however rarely looked at 472 the emergence of alternative stable states. So far, it therefore remains unclear how the structure 473 of real ecological communities constraints (or enables) the emergence of reinforcing feedback loops 474 and therefore possible alternative stable states and associated shifts in species-rich communities. 475

In a spatial context, where different ecological systems are connected to each other by movement of organisms or matter, alternative stable states at the whole system scale have been shown to be theoretically possible. However, the degree to which biotic complexity amplifies, or dampens, these transitions remains unknown. Moreover, space has usually been described in a very simplified way ignoring important aspects of topology. It is therefore unclear whether and how the spatial coupling between ecological systems may allow for (or prevent) abrupt transitions at the meta-scale.

Making progress along these lines is not only crucial for our basic knowledge of natural systems, but also urgent for the formulation of environmental policies and the prioritization of management efforts on a landscape. We need to understand which of these different aspects of complexity matters for predicting the dynamics of which ecosystems. Expanding the theory of alternative stable states and tipping points toward more complex systems could provide useful tools to map the fragility of ecosystems broadly and to monitor changes in their resilience, which will help manage ecosystems by better anticipating the effects of upcoming perturbations.

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This new understanding and indicators could also find applications in a number of complex systems 492 outside of ecology (see Table 1 of Brummitt et al. [34] for an overview). Because of the current 493 state of globalization, complex interdependent networks, involving multiple interaction types be-494 tween their components as well as several subsystems, are at the core of our modern society as 495 can be seen in the examples of power grids [82], financial systems [83, 84], transportation networks 496 [85] or the internet [86, 87]. A major challenge is the prediction and control of sudden changes 497 (failures) propagating among coupled subsystems, as seen in episodes of cascading electrical black-498 outs [33], systemic financial crises [84], contagious currency crises [88] or political uprising [89–91]. 499

Understanding stability of ecological systems is a first step in that direction and the knowledge	500
accumulated could prove to be valuable to numerous multi-layer networks outside of ecology.	501

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