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# ENSO Climate Forcing of the Marine Mercury Cycle in the Peruvian Upwelling Zone Does Not Affect Methylmercury Levels of Marine Avian Top Predators

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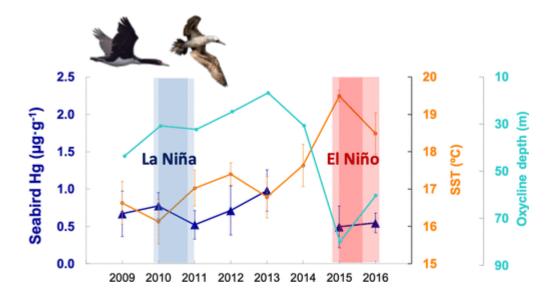
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#### Abstract :

Climate change is expected to affect marine mercury (Hg) biogeochemistry and biomagnification. Recent modeling work suggested that ocean warming increases methylmercury (MeHg) levels in fish. Here, we studied the influence of El Niño Southern Oscillations (ENSO) on Hg concentrations and stable isotopes in time series of seabird blood from the Peruvian upwelling and oxygen minimum zone. Between 2009 and 2016, La Niña (2011) and El Niño conditions (2015–2016) were accompanied by sea surface temperature anomalies up to 3 °C, oxycline depth change (20–100 m), and strong primary production gradients. Seabird Hg levels were stable and did not co-vary significantly with oceanographic parameters, nor with anchovy biomass, the primary dietary source to seabirds (90%). In contrast, seabird  $\Delta$ 199Hg, proxy for marine photochemical MeHg breakdown, and  $\delta$ 15N showed strong interannual variability (up to 0.8 and 3‰, respectively) and sharply decreased during El Niño. We suggest that lower  $\Delta$ 199Hg during El Niño represents reduced MeHg photodegradation due to the deepening of the oxycline. This process was balanced by equally reduced Hg methylation due to reduced productivity, carbon export, and remineralization. The non-dependence of seabird MeHg levels on strong ENSO variability suggests that marine predator MeHg levels may not be as sensitive to climate change as is currently thought.

# **Graphical abstract**



Keywords : Hg stable isotopes, MIF, Northern Humboldt current system, climate change, ecology

#### 62 **1 Introduction**

63 Mercury (Hg) is a pollutant of major concern, efficiently bioassimilated and bioaccumulated in 64 its methylmercury form (MeHg). It poses risks for the health of marine top predators, and humans who are mainly exposed to MeHg via seafood consumption<sup>1</sup>. In the industrial era, 65 66 anthropogenic Hg emissions have exceeded natural volcanic and soil degassing emissions by a factor of seven <sup>2</sup>, resulting in similar enrichment of atmospheric and surface ocean Hg levels 67  $^{2,3}$ . The biogeochemical Hg cycle is highly complex and comprises multiple chemical and 68 69 biological transformations that control marine MeHg formation, breakdown and bioavailability 70 to marine organisms, such as methylation/demethylation, redox reactions or photochemical 71 processes. Mercury methylation is thought to be mainly driven by anaerobic micro-organisms 72 within the water column, with a peak in net MeHg production at depth, in oceanic oxygen minimum zones (OMZ) <sup>4-6</sup> where organic matter remineralisation rates are the highest. 73 74 Consequently, the depth of occurrence of pelagic fish has been identified as a prominent control factor on MeHg concentrations in marine predators <sup>7</sup>. 75

76 The impact of climate change forcing on the net production of marine MeHg represents a yet unknown and critical focal point of Hg research <sup>8-10</sup>. Global estimations predict rising 77 78 seawater temperatures of 1.1 to 2°C by 2100 depending on emission scenarios <sup>11</sup>, an enhanced level of ocean acidification and oxygen loss <sup>12</sup>, reduced ocean primary production <sup>13</sup> and the 79 expansion of OMZ<sup>14,15</sup>. Global warming is also expected to cause the loss of coastal resources 80 81 and reduce the productivity of fisheries and aquaculture, with fisheries models estimating a decrease of 1.5 to 3.4 million tons of annual global marine fisheries catches during the 21st 82 century <sup>16,17</sup>. These changes in oceanographic dynamics, primary production, OMZ extension 83 and ocean-atmosphere exchanges can potentially influence marine MeHg production, 84 bioavailability and transfer in the food web 8,10,18. Experimental studies on estuarine and 85 86 freshwater fish showed that rising water temperatures could increase MeHg bioaccumulation

levels <sup>19,20</sup>. A recent modelling study also predicted increasing trends of MeHg concentrations 87 in fish from the Gulf of Maine due to warmer seawater temperatures <sup>9</sup>. The authors also 88 89 suggested that overfishing could lead to enhanced MeHg accumulation in some marine fish 90 species due to subsequent dietary shifts <sup>9</sup>. However, modelling studies present limitations 91 relative to the most probable warming scenarios, and the complex connections between 92 biogeochemical and ecological processes operating at regional and global scales. Experimental 93 studies and long-term observations in relevant warming marine ecosystems are needed to help 94 capture the complexity and coupling between baseline and trophic processes that govern MeHg 95 concentrations in the food web.

The Northern Humboldt Current System (NHCS) off Peru<sup>21,22</sup> represents a unique 96 97 ecosystem to explore the influence of ocean warming on marine biogeochemistry and ecology in the context of the climate forcing induced by El Niño Southern Oscillations (ENSO)<sup>22</sup>. The 98 NHCS is one of the most productive regions of the global ocean <sup>23</sup>, exhibiting the most 99 extensive and shallow OMZ<sup>22,24</sup> on earth, and accounting for >10% of global fisheries, and 100 particularly anchovy (*Engraulis ringens*)  $^{23}$ . The shallow oxycline (~20–100 m) acts as a 101 102 physical barrier for fish biomass, which is mostly concentrated at the surface. The oxycline is 103 also the relevant interface for the formation of MeHg. A high proportion of dissolved MeHg 104 (relative to total Hg) has been documented in this region in surface waters compared to other parts of the Pacific Ocean <sup>25,26</sup>. Every 2–7 years, ENSO events strongly affect marine primary 105 106 productivity and food web dynamics. During extreme Eastern Pacific (EP) El Niño events, the 107 reduced wind speed along the Peruvian coast leads to a decreased upwelling circulation that 108 brings normally deep cold nutrient rich waters to the surface sustaining the high productivity of 109 this region <sup>21</sup>. Under El Niño conditions, the thermocline, nitracline and oxycline deepen 110 significantly and the water column becomes more oxygenated at the surface, leading to warmer surface temperature anomalies <sup>24</sup>. Under these circumstances, the food web is particularly 111

affected since the phytoplankton growth is strongly reduced. This impacts the whole ecosystem leading to a reduced fish biomass potentially influencing the food web structure, causing a high mortality of marine top predators <sup>27</sup>, but also impacting fishery catch <sup>28,29</sup>.

115 In this work, we investigated how climate forcing induced by ENSO in the NHCS, 116 influences MeHg biogeochemistry and concentrations at the top of the marine food web. We 117 hypothesized that the warming conditions and the deepening of the oxycline during El Niño 118 events would change the depth at which MeHg is produced, likely decreasing its 119 photodegradation, possibly favouring its production, bioavailability and concentration further 120 in the food chain. We also explored and discussed how complementary and compensating 121 processes could mitigate the production of MeHg and its concentration in the food chain during 122 these events due to the reduced vertical export of organic matter toward the oxycline. Our 123 strategy consisted of documenting and discussing the role of ENSO through the change of the 124 environmental, physical and oceanic biogeochemical variables on Hg concentrations and Hg 125 stable isotopic composition in a 8 year time series of blood samples from two key sentinel avian 126 top predators : the Peruvian booby Sula variegata and the Guanay cormorant Phalacrocorax *bougainvillii*, that are known to feed almost exclusively (90% on average) on anchovy <sup>30,31</sup>. 127 128 Mercury concentration in seabird blood has been documented to be a very sensitive proxy of MeHg levels and bioavailability in the local and regional marine ecosystem <sup>32,33</sup>. Blood Hg 129 isotopic composition of seabirds efficiently reflects recent Hg exposure <sup>34</sup> (i.e. 2-5 weeks 130 preceding sampling <sup>35</sup>), and can provide major clues to elucidate the marine Hg cycle <sup>36,37</sup>. 131 132 Mercury stable isotopes has greatly improved knowledge on the sources of exposure, transfer pathways, and metabolism of Hg in marine consumers <sup>36,38,39</sup>. Many abiotic (e.g. 133 photoreduction, volatilization)<sup>40,41</sup> and biotic processes (e.g. methylation, demethylation)<sup>42–44</sup> 134 135 result in mass-dependent isotope fractionation (MDF, reported as  $\delta^{202}$ Hg). By contrast, massindependent fractionation of odd-mass number isotopes (odd-MIF, reported as  $\Delta^{199}$ Hg or 136

 $\Delta^{201}$ Hg) has been primarily observed during aquatic photochemical reactions <sup>40</sup>, and is 137 preserved during biological or trophic processes <sup>45</sup>. The systematic decline in  $\Delta^{199}$ Hg observed 138 139 in marine organisms with the depth at which they feed has been attributed to the decrease of 140 MeHg photodegradation in the water column, allowing to demonstrate that MeHg is mainly produced in the OMZ<sup>38</sup>. This proxy is thus particularly relevant in the context of the present 141 study. Carbon ( $\delta^{13}$ C) and nitrogen ( $\delta^{15}$ N) stable isotopes were also determined to provide 142 143 insights about the foraging ecology of the seabird species, and to detail the influence of isotopic 144 baseline variations during contrasted upwelling regimes <sup>46–49</sup> on Hg concentrations and stable 145 isotopes.

146

#### 147 2 Materials and methods

#### 148 2.1 <u>Sample collection</u>

149 Sampling was conducted from 2009 to 2016 (except 2014) during the months of October-150 November at Pescadores Island, Peru (11°46'S, 77°15'W) which correspond to the breeding 151 period. During that time, seabirds are restricted to their nesting area making Hg concentrations 152 and stable isotope data only representative of their breeding grounds in the NHCS. Individuals 153 of Guanay cormorants and Peruvian boobies were randomly chosen during each sampling 154 campaign. Each bird was measured, weighed and sexed as detailed in a previous work <sup>31</sup>. Birds 155 were breeders (breeding 1-4 weeks old chicks at the time of sampling) or juveniles. Some 156 individuals of boobies and cormorants were equipped with GPS recorders for spatial tracking of their foraging trips and time-depth recorders to track their diving behaviour <sup>31</sup>. The two 157 158 seabird species slightly differ on their foraging strategies in terms of diving depths, with cormorants presenting better diving capacities than boobies <sup>31</sup>. No sex-related differences in 159 foraging habits have been observed between males and females of the two species <sup>31</sup>. Blood 160

161 samples were collected from a wing or tarsal vein and preserved in 70% ethanol <sup>49</sup>. Samples of 162 ingested anchovies (standard length of 12-13 cm) were also collected and preserved similarly 163 in 70% ethanol when the birds regurgitated spontaneously. No other prey was observed. Prior 164 analysis, blood and prey samples were dried, lyophilized and grounded. The number of seabird 165 individuals and prey analysed for Hg concentrations, C, N isotope analysis ranged from 5 to 41 166 individuals per species depending on the year, while Hg isotope analysis was performed in 4 to 167 10 chosen individuals, per species per year, based on the sample mass available.

168

#### 169 2.2 Analytical methods

# 170 Total Hg concentrations

Total Hg concentrations, hereafter expressed in  $\mu g g^{-1}$  of dry weight (dw), were quantified by 171 using a DMA80 analyzer (Milestone, USA). The detection limit was  $0.005 \ \mu g \cdot g^{-1}$  dw. Three 172 173 certified reference materials were analysed for evaluation of accuracy and reproducibility of 174 our methods: tuna fish muscle (ERM-CE-464 and IAEA-436) and lobster hepatopancreas 175 material (TORT-3); providing a respective accuracy of 96±12% (n=15), 96±9% (n=6), and 176 100±9% (n=6) relative to recommended reference values. Internal homemade blood reference 177 material (red blood cells, RBC-KP) was also used for validation of seabird blood analyses and 178 provided an accuracy of 96±6% (n=3) relative to previously published values, as detailed elsewhere <sup>34</sup>. 179

## 180 Total Hg isotopic analyses

Blood samples (0.05–0.10 g) were soaked in 3 or 5 mL of concentrated bi-distilled HNO<sub>3</sub> (~11N) overnight at room temperature then heated on a hotplate at 80 °C during 6 h (4 h in HNO<sub>3</sub> and 2 h more after addition of 1/3 of the total volume of H<sub>2</sub>O<sub>2</sub> 30%, ULTREX quality). We added 100  $\mu$ L of BrCl (0.2 mol L<sup>-1</sup>) to convert all MeHg to inorganic Hg. The digest mixtures were finally diluted in inverse aqua regia (3 HNO<sub>3</sub>: 1 HCl). Final Hg concentrations ranged from 0.5 to 1  $ng \cdot g^{-1}$ .

187 Mercury isotopic composition was measured using a multicollector inductively coupled 188 to plasma mass spectrometry (MC-ICP-MS, Thermo Finnigan Neptune Plus) with continuous-189 flow cold vapor (CV) generation using Sn (II) reduction (CETAC HGX-200). Mercury isotopic 190 values were reported as delta notation, calculated relative to the bracketing standard NIST 191 SRM-3133 reference material to allow interlaboratory comparisons, as described in the SI. 192 Secondary standard NIST RM-8160 (previously UM-Almadén standard) was used for 193 validation of the analytical session (Table S1). Recoveries of extraction (102±17%, mean±SD) 194 were calculated for all samples (n=87) by checking the signal intensity obtained on the MC-195 ICPMS for diluted extracts relative to NIST 3133 standard (with an approximate uncertainty of 196 ±15%). Certified reference materials (ERM-CE-464 and IAEA-436) and the internal blood 197 reference material (RBC-KP) were also measured (Table S1). Uncertainty for delta values was 198 based on 2SD errors for each certified reference material (Table S1).

### 199 Carbon and nitrogen stable isotopic analyses

Blood samples were freeze-dried and powdered, and subsamples were weighed with a 200 201 microbalance and packed in tin containers. Carbon and nitrogen stable isotope ratios were 202 determined in total seabird blood and anchovies with a continuous flow mass spectrometer 203 (Delta V Advantage with a Conflo IV interface, Thermo Scientific, Bremen, Germany) coupled 204 to an elemental analyser (Flash EA1112, Thermo Scientific, Milan, Italy). Results are in delta notation relative to Vienna PeeDee Belemnite and atmospheric N<sub>2</sub> for  $\delta^{13}$ C and  $\delta^{15}$ N, 205 206 respectively. Replicate measurements of internal laboratory standards (acetanilide) indicated measurement errors <0.15‰ for both  $\delta^{13}$ C and  $\delta^{15}$ N values. 207

#### 208 2.3 Environmental parameters

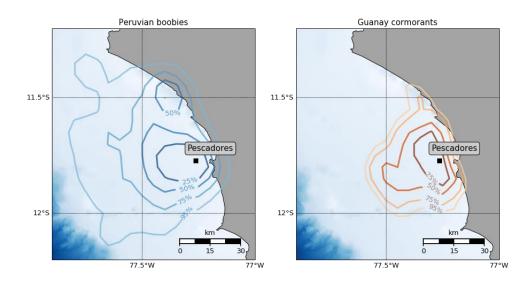
209 The spatio-temporal variability of the main marine physical and biogeochemical parameters of 210 seabird foraging areas was derived from satellite and acoustic observations. Satellite datasets 211 of sea-surface temperature (SST), net primary productivity (NPP), wind speed, diffuse 212 attenuation coefficient Kd490, photosynthetically available radiation (PAR), total kinetic 213 energy (TKE) and upwelling index (Ekman transport), were integrated at three different spatial 214 scales, respectively from the coast to 80, 100 and 150 km offshore: small-scale A1 (lat: 12.5-215 11.0°S); medium-scale A2 (lat: 13.5-10.5°S) and large-scale A3 (lat: 14.0-9.5°S). The smallscale spatial integration (A1) is based on the tracking data of the main seabird trajectories <sup>31</sup> 216 217 and reflects their effective foraging zone (Figure 1). We know that seabirds are restricted to 218 foraging areas near their colonies during their breeding period due to the need to frequently 219 feed their chicks. However, since MeHg accumulated in seabirds is mainly transferred from anchovies that integrate larger areas 50,51, we also treated the environmental data at large scale 220 area (A3) based on anchovy biomass distribution <sup>50,52</sup> for the considered period. Medium-scale 221 222 area A2 was also considered as an intermediate area between A1 and A3 (Figure S1).

223 Seabird blood represents exposure over the last 2 to 5 weeks preceding sampling <sup>34</sup>, i.e. 224 October-November. We therefore calculated mean values (and standard error, SE) of each 225 environmental parameter for the period from the 1<sup>st</sup> October to 30<sup>th</sup> November. All 226 environmental satellite variables were averaged at monthly resolution. Anomalies of the 227 environmental variables were calculated relative to mean seasonal values for the 2009-2016 228 period.

Oxycline depth variations were recorded by acoustic measurements based on the depth of the vertical extension of the epipelagic community as explained in previous work and specific for this area <sup>53</sup>. Acoustic oxycline data were previously compared and validated with *in situ* dissolved oxygen concentration analyses from the period October-December in Callao station 233 (estimation of oxycline at DO=22  $\mu$ mol Kg<sup>-1</sup>, see <sup>22,54</sup> for more details). Acoustic data used in 234 this work were processed at the three integrated areas A1, A2 and A3 with respective spatial 235 resolutions (0.5, 1 and 2 degrees and 50 meters of acoustic data resolution) and integrated 236 during the period of October to November. Acoustic data of 2013 were only available for the 237 period February-April, so we estimated the oxycline depth values for October-November 2013 238 by extrapolation validated with the correlation between oxycline depth and SST anomalies for 239 the three spatial areas.

240 Anchovy total biomass (mean  $\pm$  SE) was determined by acoustic observations and 241 integrated for a large spatial coastal area (8 to 16°S) corresponding to the spring survey (October-November) using the methodology developed by Simmonds and co-workers <sup>55</sup>. Data 242 243 of anchovy fish captures were provided by the FAO (Food and Agriculture Organization of the 244 United Nations <sup>56</sup>) for the Major Fishing Area in the southeast Pacific Ocean and by the Ministry 245 of Production of Peru<sup>57</sup>. The variability of ENSO was determined by the ICEN (Coastal El 246 Niño Index), which is a regional (Eastern Pacific) index of ENSO activity in the region of Peru 247 provided by the Technical Committee of the National Study of the El Niño Phenomenon (ENFEN) 58 248

249



251Figure 1. Seabird foraging area over the 2009-2015 period: kernel of dive density estimations for boobies252(left panel) and cormorants (right panel).

253

#### 254 2.4 <u>Statistical tests</u>

Statistical analyses were performed using the software R 3.3.2<sup>59</sup>. Data were first checked for 255 256 normality of distribution and homogeneity of variances using Shapiro-Wilk and Breusch-Pagan 257 tests, respectively. Statistical significance of the temporal variations of Hg concentrations, C, 258 N and Hg isotopic composition were examined using a linear regression fitted along years. 259 Statistical differences between years were investigated by non-parametrical tests (Kruskal-260 Wallis with Conover-Iman post-hoc test) since data did not meet specificities of normality and 261 homoscedasticity. Interspecies differences were also tested statistically by Wilcoxon tests. 262 Statistical significance was set at p<0.05.

We used generalized additive models (GAM) <sup>60</sup> to assess the influence of environmental 263 264 parameters in explaining the interannual variations of seabird Hg concentrations and isotopic signatures of the NHCS. GAM tests were fitted in R using the mgcv package <sup>61</sup>. Response 265 variables (blood Hg concentrations converted into log(Hg),  $\delta^{202}$ Hg and  $\Delta^{199}$ Hg values were 266 267 assumed to follow a Gaussian distribution. Explanatory variables tested included ecological factors ( $\delta^{15}$ N and  $\delta^{13}$ C values) and oceanographic parameters (oxycline depth, SST, ICEN 268 269 index, chlorophyll-a concentration, NPP, wind speed, diffuse attenuation coefficient Kd490, 270 PAR, coastal upwelling index, TKE and anchovy biomass). Multi-colinearity was checked by 271 calculation of the variance inflation factor (VIF) <sup>62</sup>. Only the parameters whose VIF<2 were 272 considered in the GAM models. ICEN index was found to be collinear to other variables and 273 was removed from the explanatory variables. Chlorophyll-a concentrations, PAR and NPP were 274 highly correlated, so we only tested these variables separately in the GAM models. All the 275 explanatory variables were fitted in the GAM with a low spline complexity (k = 3) to reduce over-fitting. The best-fit GAM models were selected by using the Akaike's Information
Criterion (AICc) corrected for small samples sizes <sup>60</sup>. Additionally, we examined the fitted
GAM models with diagnostic plots of residuals. To determine the amount of variation explained
by each explanatory variable, we fitted a separate model for individual variable. The deviance
explained (% DE) for each model was compared to assess the capacity of prediction of each
model. Results of GAM tests are included in the SI (Table S2).

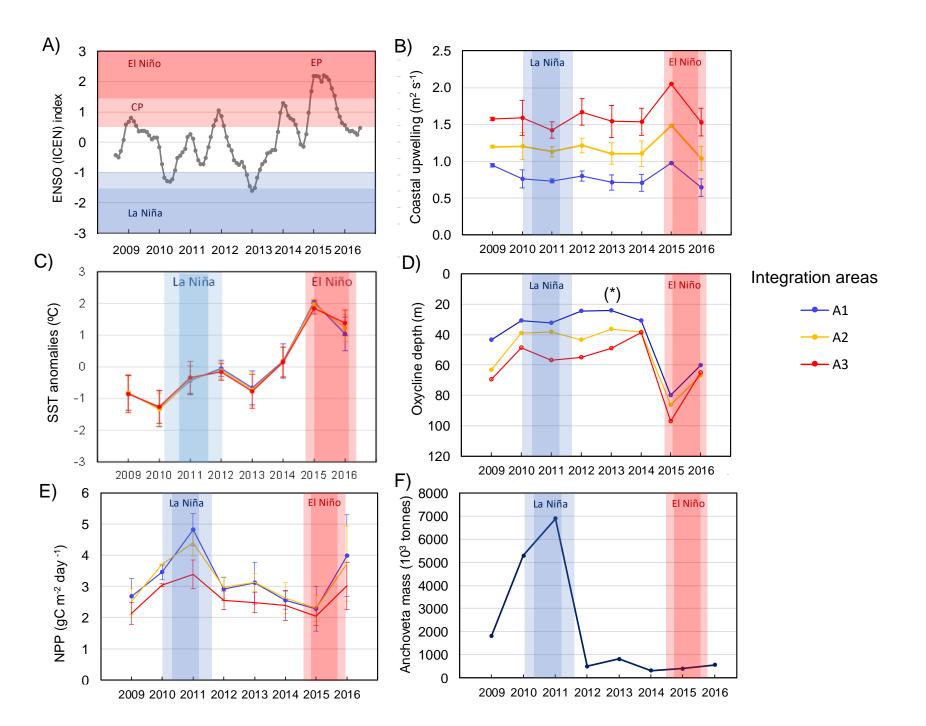
#### 282 **3 Results & discussion**

#### 283 3.1 <u>Humboldt ENSO events: physical and biogeochemical dynamics</u>

284 Monthly-averaged ICEN index shows the ENSO variability for the period 2009-2016 (Figure 285 2A). The intensity of the events (weak, moderate, strong) was also based on the ICEN index diagnostic of the ENFEN <sup>58,63</sup>. The impact of a given ENSO event on the Peruvian upwelling 286 287 system depends on its magnitude and the spatial structure. El Niño events that develop in the 288 Central-Pacific (called CP El Niño or Modoki) do not induce significant warming of the surface 289 waters off the shelf of Peru, contrary to extreme and Eastern Pacific El Niño events (EP El Niño)<sup>64,65</sup>. The definition used for the ENFEN to classify extreme ENSO events requires the 290 291 registration during more than three consecutive months of an ICEN index > 1 °C for El Niño and < -1.2 °C for La Niña events <sup>58,63</sup>. The period of 2009-2010 was classified as a weak CP El 292 293 Niño (ICEN index of 0.7 °C) exhibiting a very limited influence on the Peruvian coast <sup>29</sup>, so 294 was not considered in this study. However, an extreme EP El Niño occurred during the 2015-295 2016 period (ICEN index of 2.2 °C) affecting the NHCS. The period from mid-2010 to early 2011 was classified as a moderate La Niña event (ICEN index of -1.3 °C). 296

297 Despite a medium but sustained yearly upwelling intensity during the study period, a 298 relatively high persistent upwelling index was observed in late 2015 (Figure 2B). Positive SST 299 anomalies ( $\approx$ +2°C) were observed both in 2015 and 2016, which corresponds to the signature 300 of extreme EP El Niño conditions (Figure 2C). The warming of surface waters during El Niño 301 is caused by downwelling equatorial Kelvin waves generated by wind anomalies in the western 302 equatorial Pacific <sup>66</sup>. During El Niño events, the relatively warmer and deeper mixed layer 303 decreases water ventilation and upwelling of deep cold-water masses is reduced which 304 translates into a significant deepening of the oxycline in our study area (from 40 to ~90 m depth) 305 in 2015 (Figure 2D), whereas in normal conditions the oxycline depth (OMZ=22  $\mu$ mol Kg<sup>-1</sup>) oscillates between 15 to 25 m depth <sup>24</sup>. This strong vertical deepening of the oxycline is in 306 agreement with previous observations 54,67 and consistent with El Niño peak phase (November-307 308 December 2015).

309 The NPP off the Peruvian upwelling system was relatively stable during the period 2009-2016 and varied within the typically observed ranges for this ecosystem (1.8 to 5.2 g C 310 m<sup>-2</sup> dav<sup>-1</sup>. <sup>68</sup>) (Figure 2E). El Niño 2015-2016 did not strongly impact primary production in 311 312 the eastern Pacific, compared to past extreme EP El Niño events <sup>69</sup>. During La Niña 2010-2011, 313 the average NPP during the spring period exhibited a maximum, particularly in the spatial area near the coast A1 (4.8 gC m<sup>-2</sup> day <sup>-1</sup>). This year also accounted for an exceptionally large 314 315 anchovy biomass (~7 million tonnes). On average, enhanced global biomass of approximately +1.1 million tonnes are observed during strong La Niña years <sup>29</sup>. In contrast, EP and extreme 316 317 El Niño events commonly result in reduced global landings, with anomalies of -0.7 to -3.2million tonnes depending on the intensity of the event (Figure 2F)<sup>29</sup>. Therefore, strong El Niño 318 319 conditions in 2015-2016 and La Niña event occurring in 2010-2011 resulted in strong 320 variability of climatic conditions of the Peruvian ecosystem, with SST anomalies of ±3°C, 321 oxycline variability from 20 to 100 m depth and large range of NPP and biomass gradients.



324 Figure 2. Interannual variations of 2-month means (1<sup>st</sup> October to 30<sup>th</sup> November) of physical and 325 biogeochemical/biological parameters measured in the Humboldt upwelling region. A) ENSO index (as 326 regional Coastal El Niño index, ICEN): El Niño event in red (CP and EP refer respectively to Central Pacific 327 and Eastern Pacific El Niño events) and La Niña event in blue, B) Coastal upwelling index (m<sup>2</sup> s<sup>-1</sup>, Ekman 328 transport), C) Sea surface temperature anomalies (°C), D) Oxycline depth (m), E) Net primary production 329 (mg C.m<sup>-2</sup>.day<sup>-1</sup>), F) Anchoveta total biomass (in 10<sup>3</sup> tons). Values are means ± standard error for the total 330 number of observations over the 2-month period. (\*) Values of oxycline depth for 2013 for the period 331 October-November are estimated by extrapolation from the correlation of recorded SST anomalies and 332 oxycline depth data available for the period February-April 2013.

333

#### 334 3.2 <u>Relation of seabird isotopes to their specific foraging habits.</u>

Similar to previous works, we assumed here that total Hg concentration and Hg isotopic 335 composition in seabird blood represent the dominant (>90%) MeHg fraction <sup>36,37</sup>. Overall blood 336 337 Hg concentrations (mean  $\pm$  SD; min-max) were consistently higher in Guanay cormorants (0.86  $\pm 0.34$ ; 0.23-1.81 µg  $\cdot$ g<sup>-1</sup>, n=118) than in Peruvian boobies (0.57  $\pm 0.21$ ; 0.13-1.29 µg  $\cdot$ g<sup>-1</sup>, 338 n=110) for the period considered (Wilcoxon test, p<0.001). Fieldwork revealed almost no other 339 340 preys than anchovy in seabird regurgitate, confirming that both Guanay cormorants and Peruvian boobies mainly prey upon anchovies with a similar relative importance of 81–96% 341 and 80–93% of their diets, respectively <sup>30,70</sup>. The offset of blood Hg levels between the two 342 species may be due to slight differences in their foraging habits and/or or species-specific 343 metabolic response to Hg <sup>36,71</sup>. 344

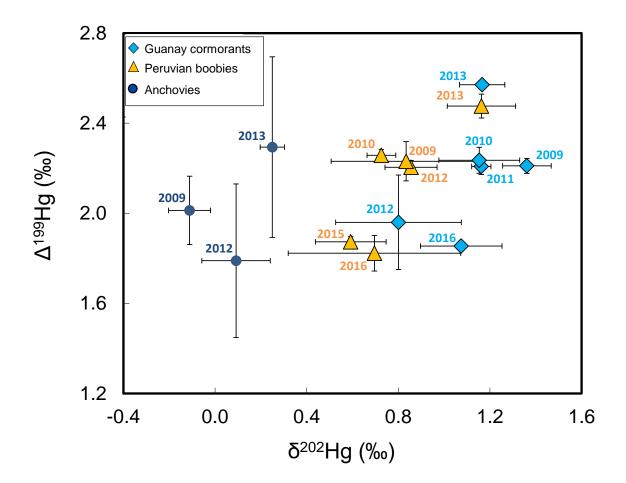
A decreasing trend of  $\delta^{13}$ C values from coastal to oceanic zones is commonly observed 345 in many marine ecosystems, including the Peruvian upwelling system <sup>46</sup>, because  $\delta^{13}$ C values 346 347 of particulate organic matter decrease from inshore to offshore waters due to the reduction of productivity  $^{68,72}$ . Consistently lower  $\delta^{13}$ C values of Guanay cormorants relative to Peruvian 348 349 boobies (p<0.001) over the studied period (Figure S4) are therefore consistent with cormorants using more inshore foraging habitats, as shown in Figure 1. Guanay cormorants are known to 350 have excellent diving capacities up to 32 m depth <sup>31</sup>, then covering potentially the deep vertical 351 352 range of the anchovy aggregations, while Peruvian boobies target the extended horizontal range of anchovy schools  ${}^{31,73}$ . Similarly,  $\delta^{13}$ C values of particulate organic matter are related to water depth  ${}^{74}$  so the vertical foraging habitats of seabirds could also influence their C isotopic values, in line with the distribution of anchovy schools distributed above the oxycline  ${}^{75}$ .

Guanav cormorants presented only slightly higher  $\delta^{15}N$  values (0.2 to 0.5% of mean 356 357 difference) than Peruvian boobies (Wilcoxon test, p<0.001), except in 2016. This suggests that 358 the two seabird species forage at a near similar trophic level and on very close prey items. The limited but significant differences in both  $\delta^{13}$ C and  $\delta^{15}$ N among the two-seabird species could 359 360 reflect slight variations of their foraging habits and specialisation in line with the variability of 361 the aggregation and distribution of anchovy schools between the surface and the oxycline. 362 Differences in seabird metabolism could also explain the slight differences observed, as for Hg. 363 However, these results confirm that both seabird species forage essentially on anchovy schools (Figure S4) confirming the field work observations by our team $^{30,70}$ . As detailed in section 3.4, 364 365 we found that during el Niño events, a reduced denitrification associated with a deeper oxycline  $^{76}$  is mostly responsible for the interannual differences in  $\delta^{15}N$  in seabirds, and for the rest of 366 the food web  $^{48}$ . 367

Individual bird blood  $\delta^{202}$ Hg (MDF) showed moderate variability both for Guanav 368 369 cormorants (0.41-1.49‰) and Peruvian boobies (0.28-1.37‰). Overall, the two species 370 presented statistically different  $\delta^{202}$ Hg values (Kruskal Wallis, H=8.144, p=0.004), especially 371 during 2009 and 2010 in which cormorants displayed 0.5% higher mean values than boobies (Figure 3). Significant correlation between Hg concentration and  $\delta^{202}$ Hg values was observed 372 373 in seabirds (R<sup>2</sup>=0.51, p<0.0001, n=72, Figure S7) although it was not significant for anchovies 374 (p=0.06, n=14). Mass dependent fractionation of Hg isotopes occurs during many physical, chemical or biological processes <sup>41,77–79</sup>. The influence of biological and ecological factors on 375 376  $\delta^{202}$ Hg values shows the limitation of this type of signature to trace which Hg transformations occur during ENSO-related variations. The interpretation of seabird  $\delta^{202}$ Hg values as a proxy 377

378 of changing environmental conditions requires a complete knowledge of all the processes and 379 factors driving Hg MDF, especially when using biological models such as seabirds (i.e., trophic ecology and intrinsic metabolic/physiological processes). For instance, seabird  $\delta^{202}$ Hg values 380 are known to be sensitive to their specific foraging habitats <sup>36</sup>. Higher Hg concentrations and 381  $\delta^{202}$ Hg values of Guanay cormorants support that this species targets larger size and deeper 382 383 anchovy aggregations than Peruvian boobies. Both seabird species exhibited significantly 384 higher  $\delta^{202}$ Hg values than anchovy (mean difference of 0.9‰). This is in good agreement with 385 the Hg MDF commonly observed in predator-prey and is thought to reflect the partial metabolic breakdown of MeHg, which leads to bioaccumulation of isotopically heavier MeHg in birds, 386 sharks and mammals <sup>42,80,81</sup>. The lower blood Hg levels in Peruvian boobies are accompanied 387 by lower  $\delta^{202}$ Hg (i.e. less demethylated MeHg), suggesting that species-specific metabolic 388 389 MeHg breakdown is not a major control factor on blood Hg.

390 We observed similar variability of  $\Delta^{199}$ Hg (Hg odd-MIF) between Guanay cormorants 391 (1.63-2.60‰) and Peruvian boobies (1.70-2.57‰) (Kruskal Wallis, H=2.117, p=0.146). Positive  $\Delta^{199}$ Hg in aquatic food webs are mainly due to the photochemical processes prior to 392 393 MeHg incorporation into food webs, then preserving the  $\Delta^{199}$ Hg baseline signature <sup>40</sup>. This is confirmed by the  $\Delta^{199}$ Hg/ $\Delta^{201}$ Hg of 1.15±0.09 (SE) of anchovy and 1.02±0.05 (SE) of seabirds 394 395 (Figure S5). Overall  $\Delta^{199}$ Hg values of anchovies (2.04±0.37‰, n=14) were similar to  $\Delta^{199}$ Hg 396 values of seabirds (2.08±0.24‰, n=72) (H=4.89, p=0.09), which is coherent with the absence of Hg odd-MIF during trophic transfer<sup>45</sup>, and confirms that the seabirds feed predominantly on 397 anchovy. A consistent decrease of fish  $\Delta^{199}$ Hg with the foraging depth has been observed in the 398 399 open Pacific ocean due to the dilution of surface photodegraded MeHg by in situ methylated Hg at depth <sup>38</sup>. Although the two seabird species slightly differ in their foraging depth <sup>31</sup>, their 400 401 similar  $\Delta^{199}$ Hg values suggest that both accumulate MeHg that has undergone similar degrees 402 of photodemethylation in the water column confirming that both species rely on similar403 anchovy aggregations.



404

405 Figure 3. Annually averaged values of Hg odd-MIF *versus* Hg MDF values ( $\Delta^{199}$ Hg *versus*  $\delta^{202}$ Hg) of seabirds 406 (2009-2016) and anchovies (2009, 2012, 2013). Values are means±SD. Number of individuals: n=4-10 per 407 species per year).

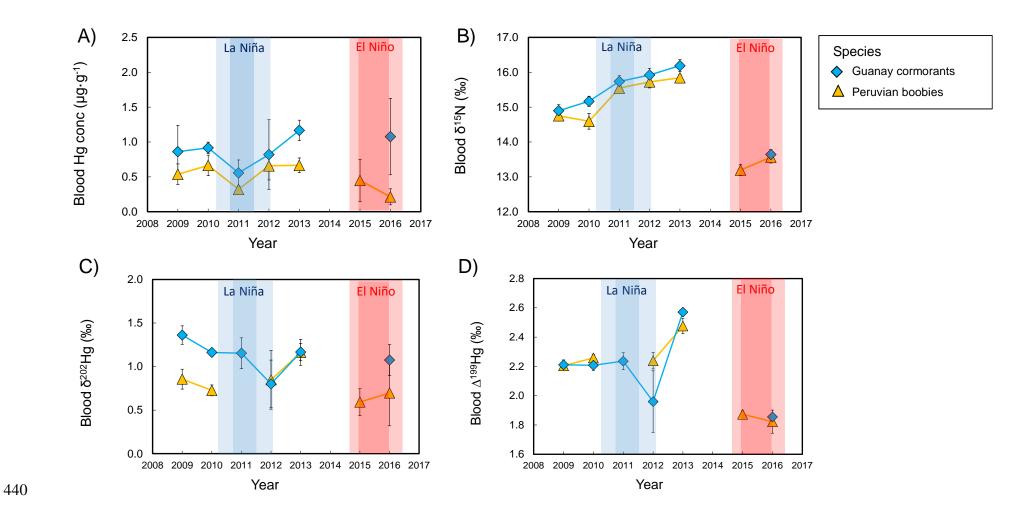
408

# 409 3.3 Limited influence of ENSO forcing on seabird blood Hg concentrations.

Significant interannual blood Hg variability was observed for both Guanay cormorants and Peruvian boobies (Kruskal-Wallis, H=48.3 and 55.4, respectively, both p<0.0001). In 2013, cormorants exhibited significantly higher mean Hg concentrations  $(1.17 \pm 0.15 \,\mu g \cdot g^{-1})$  than the rest of the years. Guanay cormorants and Peruvian boobies showed statistically lower concentrations in 2011 (0.56 ± 0.19 and 0.32 ± 0.05  $\mu g \cdot g^{-1}$ , respectively), and during 2016,

Peruvian boobies exhibited their lowest concentration of 0.21  $\pm$  0.12 µg  $\cdot$ g<sup>-1</sup>. Mercury 415 416 concentrations of anchovies (muscle) were highly variable during the period 2009-2013 (0.08  $\pm$  0.05, 0.01-0.23 µg  $\cdot$ g<sup>-1</sup>, n=66) but did not show significantly different Hg concentration 417 among years (Kruskal-Wallis, H=22.65, p=0.001) (Figure S6). When applying GAMs, the 418 419 optimal model to explain the variability of Hg concentration, expressed as log(Hg), was the model including anchovy biomass, blood  $\delta^{15}$ N and upwelling index together as explanatory 420 421 variables (Table S2). However, this model only explained for 27 to 30% of the interannual Hg 422 concentration trends, depending on the spatial area considered. This suggests a weak impact of 423 the changing biogeochemical parameters on the MeHg levels at the top of food web during 424 contrasted climatic conditions occurring on the period 2009-2016.

425 A recent study in the Galapagos archipelago also reported low interannual variations of 426 feather Hg concentrations in several species of booby Sula spp. during the period 2011-2017 427 <sup>82</sup>. This strengthens our finding, at a much larger regional scale, showing the absence of a 428 significant change in MeHg concentrations of third-level consumers of the marine food web 429 under strong climate forcing conditions. The relatively stable MeHg concentrations observed 430 in seabirds during ENSO events of high amplitude contrast with tentative modelling and 431 experimental studies which suggested higher MeHg bioaccumulation rates in fish with increasing water temperatures <sup>9,19,20</sup>. Schartup et al. <sup>9</sup> predicted that a 1°C increase of seawater 432 433 temperature would lead to a 32 and 70% increase of MeHg concentrations in Atlantic cod and 434 spiny dogfish, respectively. The same authors also suggested that dietary changes in pelagic fish in response to overfishing could lead to changes in MeHg bioaccumulation rates <sup>9</sup>. 435 Contrarily, our results show that an increase of seawater temperature of 2°C during El Niño 436 2015-2016 (which can reach +3°C range in extreme El Niño cases <sup>83</sup>) did not translate into a 437 438 significant impact on MeHg concentrations in top predators of the Peruvian food web, 439 represented here by seabirds.



441 Figure 4. Interannual variations of seabird blood A) Hg concentrations (μg ·g<sup>-1</sup>, dw); B) N isotopes (δ<sup>15</sup>N, ‰); C) Hg MDF (δ<sup>202</sup>Hg, ‰); D) Hg odd-MIF (Δ<sup>199</sup>Hg, ‰).

442 Values are Means ± SD. Number of individuals analysed: blood Hg concentration and δ<sup>15</sup>N values (5-41 individuals per species per year), Hg isotope analysis (4-10

443 individuals, per species per year).

# 444 3.4 ENSO driven variability of nitrogen isotopes

Figure 4B shows the progressive increase of seabird  $\delta^{15}N$  values from 2010 to 2013, which 445 reached the highest mean values in 2013, both for Guanay cormorants and Peruvian boobies 446 (16.19  $\pm$  0.17 and 15.84  $\pm$  0.09‰, respectively). In contrast,  $\delta^{15}$ N values of seabirds decreased 447 448 sharply in 2015 (boobies:  $13.19 \pm 0.16\%$ ) and 2016 (cormorants:  $13.64 \pm 0.14\%$  and boobies: 449  $13.56 \pm 0.17\%$ ) during the El Niño event. We observed significant interannual variations of 450 blood  $\delta^{15}$ N values of cormorants and boobies (Kruskal Wallis, H=91.3 and 97.8 respectively, both p<0.0001). Interestingly, the observation of a limited interspecies difference in  $\delta^{15}N$  (0.2 451 452 to 0.5%) among all years, in the context of following similar and parallel interannual trends 453 capturing the exact same  $\delta^{15}$ N drop of ~2.5‰ in 2015-2016 (which corresponds in theory to a 454 change close of one trophic level) is inconsistent with the idea that it might be caused by a 455 change in diet for the two seabird species. The vertical movements of anchovies to deeper zones during El Niño<sup>75</sup> could lead to higher foraging specialisation of the two seabird species due to 456 457 higher competition for prey, as documented in other booby species <sup>84</sup>. Therefore, we would 458 expect that significant changes in the diet of seabirds could lead to high inter-species differences 459 in  $\delta^{15}$ N values during years of less prey abundance, but we did not observe this trend during our studied period. A pronounced baseline shift in  $\delta^{15}$ N translated into the local food chain, 460 caused by a change in nutrient regimes related to the upwelling dynamics driven by ENSO <sup>76</sup> is 461 462 most likely the driving factor. Indeed, despite  $\delta^{15}N$  represents a unique proxy to document 463 trophic interactions and seabird diet in most ecosystems, the singularity of the upwelling regime 464 on the N cycle operating at the baseline level of the NHCS is known to induce significant  $\delta^{15}N$ 465 baseline shifts translated up to marine predators <sup>46,47</sup>. This finding is consistent with (i) field 466 observations confirming that anchovy was almost the sole species observed in seabird regurgitates for all years, (ii) that the biomass of anchovies has decreased and stayed low as 467 soon as 2012 (Figure 2F), not influencing the  $\delta^{15}N$  shift in seabirds during the same period of 468

time, and (iii) that the interannual differences in  $\delta^{13}$ C and  $\delta^{15}$ N in the two seabird species reflect those in anchovies for the same years (Figure S4). Baseline processes are thus mostly responsible for the interannual differences in  $\delta^{15}$ N values in seabird blood.

472 Marine oxygen consumption strongly determines N-loss processes, such as denitrification and anammox<sup>85</sup>, that induce significant N isotopic fractionation. Anaerobic 473 474 bacteria use nitrate (NO<sub>3</sub><sup>-</sup>) as an oxidant for the degradation of organic matter and its reduction produces isotopically light products ( $N_2$  and  $N_2O$ ) and isotopically heavier residual nitrate <sup>86,87</sup>. 475 476 Also, anaerobic ammonia  $(NH_4^+)$  oxidation (annamox) is likely the predominant pathway for N loss in oceanic OMZs 88. Both heterotrophic denitrification and anammox processes appear 477 to decrease significantly under El Niño conditions<sup>89</sup>, coupled to low primary productivity, the 478 479 deepening of the OMZ distribution and high bottom water oxygenation. Therefore, lower 480 organic matter mineralization and limited nitrate supply to surface waters under El Niño 481 conditions depletes isotopically the particulate organic matter utilized by phytoplankton, leading to lower  $\delta^{15}$ N values in biota <sup>89,90</sup>. Consequently, the significant interannual variations 482 of seabird  $\delta^{15}$ N are more likely related to shifts in the N isotopic baseline and therefore can be 483 484 used as a proxy of oxygenation events/organic matter remineralization, like El Niño or the warm ENSO events of the NHCS <sup>91</sup>. 485

### 486 3.5 ENSO driven variability in Hg isotopes

487 Despite the absence of substantial changes in Hg concentration in biota over the studied period, 488 we detected an enhanced interannual variability of seabird Hg isotopes (Figure 4C-D). We 489 observed no linear relationship of  $\delta^{202}$ Hg with ENSO events proxied by ICEN index (p=0.08) 490 and no significant  $\delta^{202}$ Hg interannual variations neither for cormorants (Kruskal Wallis, 491 H=17.1, p=0.004) nor boobies (H=16.4, p=0.006). The GAM applied for  $\delta^{202}$ Hg showed that 492 the best explanatory variables are  $\delta^{15}$ N, upwelling index and NPP together, but only explained 493 for 36 to 40% of the total  $\delta^{202}$ Hg variability, depending on the integrated area (Table S2).

Interannual variability of  $\Delta^{199}$ Hg of Guanay cormorants and Peruvian boobies was significant 494 (Kruskal Wallis, H=28.8 and 32.21 respectively, both p<0.0001) and lower  $\Delta^{199}$ Hg values were 495 496 observed during El Niño years relative to La Niña and normal years for both species (H=24.02, 497 p<0.0001) (Figure 4D). Both Guanay cormorants and Peruvian boobies exhibited higher  $\Delta^{199}$ Hg 498 values in 2013 ( $2.57 \pm 0.02$  and  $2.48 \pm 0.05\%$ ), with a pronounced significant decrease in 2015 499 (boobies:  $1.87 \pm 0.02\%$ ) and 2016 (cormorants:  $1.85 \pm 0.02$  and boobies:  $1.82 \pm 0.08\%$ ). Significant linear relationships were found between  $\Delta^{199}$ Hg and  $\delta^{15}$ N values of cormorants 500 501  $(R^2=0.59, p<0.001)$  and boobies  $(R^2=0.77, p<0.001)$ , suggesting that both isotopic signatures are biogeochemically related and that interannual variations in  $\Delta^{199}$ Hg and  $\delta^{15}$ N values are not 502 trophic or ecologically driven. Additionally, results of GAM showed that seabird  $\Delta^{199}$ Hg are 503 strongly explained by oxycline depth, SST and NPP, explaining for 77% of the  $\Delta^{199}$ Hg 504 505 internannual variability (Table S2).

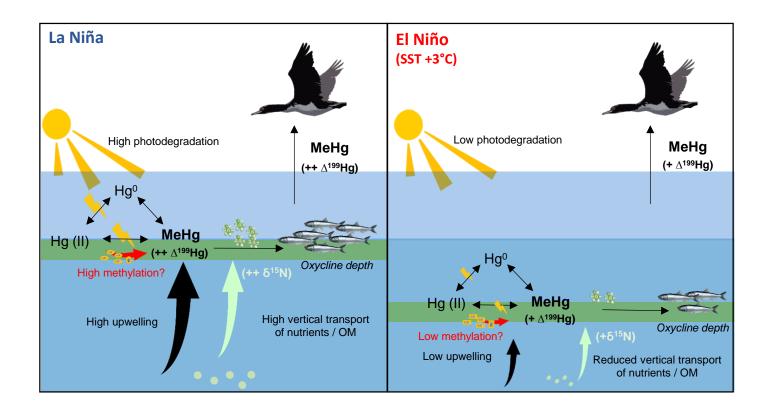
506 Due to the sensitivity of  $\Delta^{199}$ Hg to sea-surface MeHg photochemical reactions, these 507 results suggest that seabird  $\Delta^{199}$ Hg values likely reflect the change at which MeHg is formed 508 and photodegraded, mirroring the oxycline depth oscillations during ENSO events. Indeed, the sharp decrease of seabird  $\Delta^{199}$ Hg values during the El Niño 2015-2016 (down to ~1.8‰) is 509 510 coincident with a deepening of the oxycline from 30-50 m to 100 m depth (Figure 2D), and 511 therefore likely reflects a larger contribution of deeper and less photodegraded MeHg 512 production. Oxycline depth variability has a direct impact on the biogeochemical processes of 513 the northern region of the NHCS because oxygen is a control factor in the distribution of many organisms <sup>53,92</sup>. The bottom depth of the oxycline is the zone where the most intense particulate 514 515 organic matter remineralization occurs, an essential process for the maintenance of the Peru 516 OMZ <sup>93</sup> and Hg methylation. Vertical profiles across a large coastal-oceanic section off Peru 517 have shown maxima of MeHg concentrations within the OMZ and near the subsurface chlorophyll maximum<sup>25</sup>, that are expected to be driven by *in situ* microbial methylation in these 518

519 coastal productive areas. During El Niño episodes, the deepening of the oxycline can favour 520 the expansion of the vertical habitat of most marine organisms into deeper zones of the water 521 column that are less connected to the photic zone. In contrast, the sudden increase of seabird 522  $\Delta^{199}$ Hg values from 2012 to 2013 is coincident with an increase of  $\delta^{15}$ N values, and of shallower 523 oxycline (Figure 2D), likely favouring MeHg formation and photodegradation near the surface 524 in 2013 in the context of a compressed habitat for anchovy schools.

525 Our results showing profound changes in the depth at which MeHg is formed and 526 photodegraded while displaying a limited impact on seabird MeHg concentrations during 527 ENSO can be explained by different factors. We consider two potential explanations for this. 528 First, photochemical MeHg degradation may not be the dominant MeHg loss mechanism above the oxycline, leading to measurable ENSO driven  $\Delta^{199}$ Hg variability but not food web MeHg 529 530 variability. Second, photochemical MeHg degradation is the dominant loss mechanism, 531 enhanced in 2013 due to the shallow oxycline. MeHg photodemethylation is balanced by 532 enhanced inorganic Hg methylation in the OMZ due to higher primary productivity and ensuing 533 particle remineralization and Hg methylation in the OMZ under these conditions (Figure 5). During El Niño events, the deepening of the oxycline leading to reduced MeHg 534 photodegradation, as suggested by the drop in  $\Delta^{199}$ Hg, is associated with lower vertical inputs 535 of organic matter and remineralization traced by lower  $\delta^{15}$ N, leading to lower MeHg formation. 536 537 Finally, we considered the possibility that during La Niña years, a higher advection of nutrients 538 and organic matter towards the surface increases the supply of MeHg produced at depth (with close-to zero  $\Delta^{199}$ Hg values). However, this scenario would lead to a dilution of the highly 539 photodemethylated MeHg in the surface with MeHg from deeper zones, lowering the  $\Delta^{199}$ Hg 540 541 values of MeHg assimilated by biota during La Niña, which is the opposite to our observations. 542 Consequently, our results reveal that variability in  $\Delta^{199}$ Hg values of the Peruvian food web are mostly driven by changes in productivity and in oxycline depth rather than water mass mixing,and that most bioavailable MeHg is produced in the oxycline.

545 In summary, the strong ENSO climate variability of SST, NPP and oxycline depth in 546 the NHCS and OMZ was not accompanied by variability in marine food web top predator 547 MeHg concentrations. These observations contrast with recent model predictions, based on 548 Atlantic Ocean food web data, that ocean warming and overfishing have led to increases in fish MeHg levels <sup>9</sup>.We suggest that marine MeHg dynamics, food web uptake and bioaccumulation 549 550 are more resilient to climate warming than suggested by the model developed in Schartup et al. 2019 <sup>9</sup>. Further research investigating other oceanic regions, and exploring different marine 551 552 ecosystems are needed.

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- 555



558Figure 5. Schematic figure about the dominant Hg processes during contrasted ENSO events. During La Niña (left panel), the conditions of higher supply of nutrient559may favour productivity potentially enhancing in situ microbial methylation/demethylation of Hg in the OMZ. However, since the oxycline is shallower, the extent of560MeHg photodemethylation would be high (leading to higher  $\Delta^{199}$ Hg values). Conversely, during El Niño, lower nutrient supply and remineralization may reduce561microbial Hg methylation rates, and, as the oxycline is much deeper, MeHg photodegradation processes would be inhibited. Both processes (MeHg formation and562MeHg degradation) seem to be compensated, leading to stable MeHg concentrations in the Peruvian food web during extreme ENSO events.

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# 576 5 Supporting information paragraph

577 Further experimental details of methodological approach; Statistical results of generalized 578 additive models (GAM) models; Detailed QA/QC of Hg isotopic analyses; Complete tables of 579 Hg concentrations and C, N and Hg isotopic results of seabirds; Supplementary figures: Figure 580 S1, Spatial resolution of the areas integrated for satellite data; Figure S2, Results of the optimal 581 GAM models; Figure S3, Interannual variations of 2-month averages of additional physical and biogeochemical/biological parameters in the NHCS; Figure S4, Annually averaged  $\delta^{15}$ N and 582  $\delta^{13}$ C values of blood samples of seabirds and prey; Figure S5,  $\Delta^{199}$ Hg versus  $\Delta^{201}$ Hg values of 583 584 seabirds and anchovies; Figure S6, Temporal trend of Hg concentrations of anchovy samples Figure S7, Hg concentrations *versus*  $\delta^{202}$ Hg values of blood samples of seabirds and anchovies. 585

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