



HAL
open science

Temperature and nutrients as drivers of microbially mediated arsenic oxidation and removal from acid mine drainage

Vincent Tardy, Corinne Casiot, Lidia Fernandez-Rojo, Eleonore Resongles, Angélique Desoeuvre, Catherine Joulian, Fabienne Battaglia-Brunet, Marina Hery

► To cite this version:

Vincent Tardy, Corinne Casiot, Lidia Fernandez-Rojo, Eleonore Resongles, Angélique Desoeuvre, et al.. Temperature and nutrients as drivers of microbially mediated arsenic oxidation and removal from acid mine drainage. *Applied Microbiology and Biotechnology*, 2018, 102 (5), pp.2413-2424. 10.1007/s00253-017-8716-4 . hal-02110146

HAL Id: hal-02110146

<https://hal.umontpellier.fr/hal-02110146>

Submitted on 28 May 2021

HAL is a multi-disciplinary open access archive for the deposit and dissemination of scientific research documents, whether they are published or not. The documents may come from teaching and research institutions in France or abroad, or from public or private research centers.

L'archive ouverte pluridisciplinaire **HAL**, est destinée au dépôt et à la diffusion de documents scientifiques de niveau recherche, publiés ou non, émanant des établissements d'enseignement et de recherche français ou étrangers, des laboratoires publics ou privés.

1 **Temperature and nutrients as drivers of microbially mediated**
2 **arsenic oxidation and removal from Acid Mine Drainage**

3 Vincent Tardy¹, Corinne Casiot¹, Lidia Fernandez-Rojo¹, Eléonore Resongles¹, Angélique Desoeuvre¹, Catherine
4 Jouliau², Fabienne Battaglia-Brunet² and Marina Héry^{1†}

5
6 ¹*Laboratoire HydroSciences Montpellier, UMR 5569, Montpellier, France*

7 ²*BRGM, Water, Environment and Ecotechnology Division, Environmental Biogeochemistry and Water Quality*
8 *Unit, Orléans, France*

9
10
11
12
13
14
15 [†]*Correspondence:* Marina Héry, laboratoire HydroSciences Montpellier, UMR 5569, CC 57, 163 rue Auguste
16 Broussonet, 34990, Montpellier, France
17 Email: marina.hery@umontpellier.fr

29 **Abstract**

30 Microbial oxidation of iron (Fe) and arsenic (As) followed by their co-precipitation lead to the natural attenuation
31 of these elements in As-rich Acid Mine Drainage (AMD). The parameters driving the activity and diversity of
32 bacterial communities responsible for this mitigation remain poorly understood. We conducted batch experiments
33 to investigate the effect of temperature (20 vs 35°C) and nutrient supply on the rate of Fe and As oxidation and
34 precipitation, the bacterial diversity (high-throughput sequencing of 16S rRNA gene) and the As oxidation
35 potential (quantification of *aioA* gene) in AMD from the Carnoulès mine (France). In batch incubated at 20°C, the
36 dominance of iron-oxidizing bacteria related to *Gallionella* spp. was associated with almost complete iron
37 oxidation (98%). However, negligible As oxidation led to the formation of As(III)-rich precipitates. Incubation at
38 35°C and nutrient supply both stimulated As oxidation (71-75%), linked to a higher abundance of *aioA* gene and
39 the dominance of As-oxidizing bacteria related to *Thiomonas* spp. As a consequence, As(V)-rich precipitates (70-
40 98% of total As) were produced. Our results highlight strong links between indigenous bacterial community
41 composition and iron and arsenic removal efficiency within AMD, and provide new insights for the future
42 development of a biological treatment of As-rich AMD.

43 **Keywords:** Acid Mine Drainage, arsenic and iron oxidation, bacterial community, temperature, nutrient.

44 Introduction

45 Arsenic (As) is one of the most toxic pollutants commonly associated with mine tailings and Acid Mine
46 Drainage (AMD) with concentration in mine waters ranging from $< 1 \mu\text{g l}^{-1}$ to hundreds of mg l^{-1} (Casiot et al.
47 2003a; Cheng et al. 2009). Because of the severe toxicological effects of As, contaminated waters represent a
48 serious threat for ecosystems located downstream from mining sites and for public health. Numerous studies have
49 reported natural attenuation of arsenic pollution in different AMD across the world (Fukushi et al. 2003; Asta et
50 al. 2010; Egal et al. 2010). The exploitation of the microbially mediated processes involved in this attenuation
51 represents a promising strategy for the development of treatment of As-rich AMD (Johnson and Hallberg 2005).

52 Natural attenuation involves biological oxidation of ferrous iron (Fe(II)) to ferric iron (Fe(III)) and the
53 subsequent adsorption of As onto the newly formed Fe(III) precipitates or its co-precipitation (Paikaray 2015;
54 Ahoranta et al. 2016). Efficiency of arsenic removal from AMD depends on its redox speciation. Indeed, under
55 acid pH, arsenate (As(V)) is more efficiently trapped onto iron phases than arsenite (As(III)) (Hug and Leupin
56 2003). In the environment, As(III) oxidation to As(V) is mainly catalyzed by microbial activity, chemical oxidation
57 being generally very slow (Campbell and Nordstrom, 2014). Therefore, the ability of indigenous bacterial
58 populations to oxidize As(III) largely contributes, together with the activity of iron-oxidizing bacteria (FeOB), to
59 a sustainable arsenic pollution mitigation in AMD. FeOB and As(III)-oxidizing bacteria (AsOB) have been
60 isolated from AMD and their metabolic capacities were investigated (Battaglia-Brunet et al. 2002; Duquesne et al.
61 2003; Bruneel et al. 2003; Casiot et al. 2003b; Egal et al. 2010). However, the factors driving the activity and
62 diversity of indigenous complex populations of FeOB and AsOB involved in arsenic mitigation remain poorly
63 understood.

64 *Acidithiobacillus ferroxydans*, a FeOB associated with attenuation process in AMD, is a strict
65 chemolithoautotroph bacterium (Duquesne et al. 2003; Johnson and Hallberg 2005; Egal et al. 2009). Conversely,
66 AsOB can grow heterotrophically or autotrophically (Santini et al. 2000; Battaglia-Brunet et al. 2006; Garcia-
67 Dominguez et al. 2008). In particular, *Thiomonas* spp. are facultative chemolithoautotrophs that grow optimally
68 in mixotrophic media containing reduced inorganic sulfur compounds and organic supplements (Kelly et al. 2007;
69 Bryan et al. 2009; Slyemi et al. 2011). Thus, nutrient supply in AMD is expected to have a contrasted incidence
70 on pollution mitigation depending on the metabolic feature of the bacterial populations involved. Temperature is
71 another primary factor governing activity and diversity of bacterial community inhabiting AMD (Méndez-García
72 et al. 2015). Previous work on FeOB and AsOB bacterial strains isolated from diverse polluted environments
73 showed that their growth and their oxidation activities were temperature dependent (Battaglia-Brunet et al. 2002;
74 Dopson et al. 2006; Kim et al. 2008; Ito et al. 2012). Furthermore, temperature variation was suggested as a driving
75 factor shaping bacterial communities structure in AMD (Volant et al. 2014). Recently, Debiec and colleagues
76 (2017) showed that both nutrient concentration and temperature were key factors controlling the growth and the
77 oxidation rate of *Sinorhizobium* sp. M14, an AsOB isolated from neutral gold mine waters. Under batch conditions,
78 a temperature increase (from 10 to 23 or 30°C) resulted in the stimulation of bacterial growth associated with a
79 faster As(III) oxidation rate. Under continuous conditions, a supply of yeast extract stimulated both the growth
80 and the As(III) oxidation activity of *Sinorhizobium* sp. M14 (Debiec et al. 2017). Whether or not such studies
81 based on a single strain may be extrapolated to a metabolically and taxonomically diverse indigenous community
82 has not been explored so far. In this context, the aim of the present study was to assess the influence of temperature

83 and nutrients on the diversity of an indigenous AMD bacterial community and on its efficiency for iron and arsenic
84 oxidation and removal from water.

85 For this purpose, we conducted batch experiments with As-rich AMD water of Carnoulès mine (Southern
86 France) incubated either at 20°C or 35°C and supplied or not with yeast extract. Iron and arsenic speciation was
87 monitored in the dissolved phase and in the biogenic precipitates that formed during batch incubation. Diversity
88 of bacterial community was characterized by high-throughput sequencing of 16S rRNA genes and the genetic
89 potential for arsenic oxidation was evaluated by the quantification of *aioA* genes.

90 **Materials and methods**

91 **Water sampling and batch experiment setup**

92 Water was collected in June 2015 from the source of Reigous creek at the abandoned Carnoulès mine
93 (Southern France: 44°7'2.14"N; 4°0'6.00"E). Its physicochemical characteristics were as follows: pH 4.7, 2.69
94 mg O₂ l⁻¹, 14.9°C, 2.56 mS cm⁻¹, 590.6 mV, 511 mg Fe l⁻¹ (100% Fe(II)), 53.8 mg As l⁻¹ (14% As(V)).

95 Batch experiment was setup within six hours of water collection by transferring 450 ml of water in 1 liter
96 Schott Duran® bottles, previously acid-cleaned and autoclaved. Two conditions were tested: biotic (water with
97 indigenous microbial communities) and abiotic (sterile-filtered water with 0.22 µm cellulose acetate filter). For
98 each condition, four treatments ("T") were applied: (i) 20°C without nutrient supply ("T20"), (ii) 35°C without
99 nutrient supply ("T35"), (iii) 20°C with nutrient supply ("T20Y"), and (iv) 35°C with nutrient supply ("T35Y").
100 Nutrient supply consisted of spiking water with yeast extract at a final concentration of 0.2 g l⁻¹. This concentration
101 is typically used in culture media for the heterotrophic growth of AsOB like *Thiomonas* spp. (Battaglia-Brunet et
102 al. 2002; 2006). A total of 12 biotic batch experiments (four treatments, three replicates) and eight abiotic batch
103 experiments (four treatments, two replicates) were set up. All batch were closed with cellulose stoppers to prevent
104 bacterial contamination from outside and to allow oxygen diffusion inside the batch. Batch were placed under
105 orbital agitation at 150 rpm in thermo-regulated chambers to assure water aeration and a constant temperature (set
106 up at 20°C or 35°C) throughout the experiment duration.

107 **Batch experiment monitoring**

108 The experiment was conducted for eight days. Water samples (~3 ml) were collected from the bottles for
109 chemical analysis at days 0, 2, 3, 4, 5, 6, 7 and 8. At day 3, 5 and 8, pH and dissolved oxygen were measured with
110 a multiparameter analyser (Ultrameter™ Model 6P). Dissolved oxygen remained stable, with an average of 7.8 ±
111 0.5 mg O₂ l⁻¹ in all batch (data not shown). pH decreased during incubation from 4.7 to 2.7 ± 0.2 and 3.6 ± 0.1 in
112 the biotic and abiotic experiments respectively (data not shown), in relation with Fe(II) oxidation and subsequent
113 Fe(III) hydrolysis (Nordstrom and Alpers 1999). After sampling, water was immediately filtered through
114 disposable filters (cellulose acetate, pore size 0.22 µm) and the filtrate was analyzed for Fe(II), total Fe, As(III)
115 and As(V) concentration according to routine procedures described in Fernandez-Rojo et al. (2017).

116 At the end of the experiment, batch were sacrificed. After homogenization of liquid and solid phases
117 (including biogenic precipitates that formed during the incubation in biotic batch), a subsample (100 ml) was
118 filtered on sterile 0.22 µm cellulose acetate filters. The filters were stored at -80°C before DNA extraction. Another
119 subsample (3 ml) was collected for biomass quantification by flow cytometry. At last, the remaining batch content
120 (~300-350 ml) was filtrated using 0.22 µm cellulose acetate filter for quantification of total iron and arsenic species
121 (As(III) and (V)) in the particulate fractions. The filter was placed in a vacuum desiccator, dried until constant
122 weight and a chemical extraction with orthophosphoric acid was performed as described in Resongles *et al.* (2016).

123 **Chemical analyses**

124 Total dissolved Fe and Fe(II) concentrations were determined with a spectrophotometer (SECOMAN
125 S250, detection limit = 88 µg l⁻¹, uncertainty = ± 5 %) at 510 nm wavelength (Rodier 1996). For As speciation
126 analysis in the dissolved phase and precipitate extracts, samples were analyzed with HPLC (High Performance

127 Liquid Chromatography) using an anion exchange column (25 cm x 4.1 mm i.d. Hamilton PRP-X100) coupled to
128 ICP-MS (PQ2+, X Series, Thermo; Detection limit = 0.2 µg l⁻¹ for As(III), 0.4 µg l⁻¹ for As(V), uncertainty = ± 5
129 %) (Héry et al. 2014; Resongles et al. 2016). The Certified Reference Water NIST1643e was used to check the
130 analytical accuracy for total As concentration and the RSD was always lower than 5 % with respect to the certified
131 value.

132 The proportion of arsenic oxidized after the eight-day incubation period was calculated using the
133 following equation:

$$134 \text{As(III)}_{\text{oxidized}} = \frac{(\text{As(III)}_d t^0 + \text{As(III)}_p t^0) - (\text{As(III)}_d t^{\text{final}} + \text{As(III)}_p t^{\text{final}})}{(\text{As(III)}_d t^0 + \text{As(III)}_p t^0)} \times 100$$

135 where As(III)_d and As(III)_p were the concentration of dissolved and particulate As(III) (in mg l⁻¹), respectively, at
136 the beginning of experiment (t⁰) or after 8 days of incubation (t^{final}).

137 **Biomass quantification**

138 Quantification of bacterial biomass from samples (water and precipitate) was performed by flow
139 cytometric counting method. Prior analysis, bacterial cells were detached from mineral biogenic precipitates
140 according to the procedure of Lunau et al. (2005), with some modifications. Briefly, methanol 100% was added to
141 water sample to reach a final concentration of 10% and incubated 15 min in an ultrasonic bath (42 kHz). Detrital
142 and inorganic particles were removed by a low centrifugation (1 min at 190 × g) and supernatant was recovered in
143 nine volumes of sterile Milli-Q® water. Then, bacterial cells were pooled by centrifugation at 6500 × g for 10 min,
144 washed in Milli-Q® water and stained 15 min in the dark with the LIVE/DEAD BacLight bacterial viability kit
145 (Invitogen, Carlsbad, CA, USA). Bacterial cells abundance was measured with a Gallios™ flow cytometer
146 (Beckman Coulter, Brea, CA). Live and dead bacterial cells were detected at 520 nm and 630 nm, respectively.
147 More than 20,000 analytical events were counted in triplicate for each sample and data analysis was processed
148 with Kaluza software (Beckman Coulter). Since the numbers of live and dead bacteria were in the same order of
149 magnitude in all the samples, only the total bacteria cells (*i.e.* live + dead bacteria) are presented.

150 **DNA extraction and quantification**

151 For each biotic treatment, DNA was extracted from 100 ml of homogenised and filtered batch content (as
152 described in 2.2) using the Powerwater DNA Isolation Kit according to the manufacturer's recommendations
153 (MoBio Laboratories Inc., Carlsbad, CA, USA). DNA was quantified with a fluorometer (Qubit®, Invitrogen) and
154 stored at -20°C until further analysis.

155 **Quantification of *aioA* genes**

156 The abundance *aioA* genes encoding the catalytic subunit of the As(III)-oxidase was determined by
157 quantitative real-time PCR (qPCR). The reverse primers aoxBM2-1R and forward primer aoxBM4-1F were used
158 to target a 110 bp fragment of the *aioA* gene (Quéméneur et al. 2010). For each sample, 2 ng of DNA were used
159 in a 20 µl PCR reaction with 0.3 µM of each primer, 100 ng of T4GP32 (MP Biomedicals) and 1X IQ SYBR
160 Green Supermix (BioRad). The program was run in a CFX Connect (BioRad) and consisted in an initial
161 denaturation at 95 °C for 3 min, followed by 40 cycles of 95 °C for 10 s, 54 °C for 20 s, 72 °C for 10 s, and a data

162 acquisition step at 80 °C for 10 s. At the end, a melting curve analysis was performed through measurement of the
163 SYBR Green I signal intensities during a 0.5 °C temperature increment every 10 s from 65 °C to 95 °C.

164 **Sequencing of 16S rRNA gene**

165 Bacterial diversity was determined by Illumina high-throughput sequencing of bacterial 16S rRNA genes.
166 V4-V5 region (about 450 bases) was amplified by PCR using primers PCR1_515F (Barret et al. 2015) and
167 PCR1_928R (Wang and Qian 2009). For each sample, 10 ng of DNA were used in a 50 µl PCR reaction conducted
168 under the following conditions: 94°C for 2 min, 30 cycles of 1 min at 94°C, 65°C for 40 s and 72°C for 30 sec,
169 followed by 10 min at 72°C. The PCR products were checked by gel electrophoresis and quantified using a
170 fluorometer (Qubit®, Invitrogen) and sent to GeT-PlaGe platform (Toulouse, France) for Illumina MiSeq analysis
171 using a 2×300 bp protocol.

172 **Bioinformatic analyses of 16S rRNA gene sequences**

173 Raw sequence reads were merged into full-length sequences by FLASH v1.2.11 (Magoc and Salzberg
174 2011). Reads were further processed using the software program MOTHUR version 1.31 (Schloss et al. 2009).
175 Firstly, raw sequences were selected based on the following criteria: (i) length (between 350 and 460 bp), (ii)
176 homopolymer lengths (< 7) and (iii) the absence of ambiguous bases. Then, sequences were aligned against the
177 SILVA reference database (Release 123) and removed when they did not align correctly. Chimeric sequences were
178 detected and removed using the implementation of Chimera UCHIME (Edgar et al. 2011). A further screening
179 step (pre-cluster) was applied to reduce sequencing noise by clustering reads differing by only one base every 100
180 bases. Taxonomic affiliation of 16S rRNA genes was performed with a Bayesian classifier (Wang et al. 2007)
181 (80% bootstrap confidence score) against the SILVA reference database. In order to efficiently compare the
182 datasets and avoid biased community comparisons, the sample reads were reduced to the lowest datasets by random
183 selection (4353 reads). The remaining high quality sequences were used to generate a distance matrix and clustered
184 into Operational Taxonomic Units (OTUs) defined at 97% cutoff using the average neighbor algorithm. OTU-
185 based diversity indices, rarefaction curves and Unifrac distance were calculated with MOTHUR at a level of 97%
186 sequence similarity. The raw datasets are available on the European Nucleotide Archive system under project
187 accession number PRJEB21683.

188 **Statistical analysis**

189 The statistical significance of the bacterial biomass increase during batch incubations was assessed with
190 the nonparametric Kruskal-Wallis test. Physicochemical parameters, oxidation rates, diversity metric and *aioA*
191 genes quantification obtained for the biotic treatments were compared by two-way ANOVA and the differences
192 between them analyzed with a Fisher test ($P < 0.05$). Differences in bacterial community structure between biotic
193 treatments were characterized using UniFrac distance (Lozupone and Knight 2005). Non Metric Multidimensional
194 Scaling (NMDS) was used to graphically depict differences between the bacterial communities. The significance
195 of the observed clustering of samples on the ordination plot was assessed by an ANalysis Of SIMilarity (ANOSIM,
196 999 permutations). All these statistical analyses were performed with the R free software ([http://www.r-](http://www.r-project.org/)
197 [project.org/](http://www.r-project.org/)).

198

199 **Results**

200 **Bacterial biomass**

201 Bacterial biomass initially present in the water collected at the Reigous spring was $1.4 \pm 0.8 \times 10^4$ bacterial
202 cells ml^{-1} . After 8 days of incubation in biotic batch, bacterial cell concentration increased from 1 to 3 order of
203 magnitude depending on the treatment applied ($T20 \approx T35 < T35Y < T20Y$) (Fig. 1). In abiotic batch, small
204 particles that most probably corresponded to Fe colloids induced a background noise ranging from 2.6×10^3 to 1.8
205 $\times 10^4$ particles ml^{-1} .

206 **Evolution of dissolved Fe(II), dissolved As(III) and precipitated Fe concentrations**

207 In the biotic batch experiments, the concentration of dissolved Fe(II) and As(III) decreased substantially
208 over time (Fig. 2A and 2B). In batch incubated at 20°C without nutrient (T20), Fe(II) concentration decreased
209 gradually from 516 mg l^{-1} to 20 mg l^{-1} . At 35°C without nutrients (T35), Fe(II) concentration decreased more
210 rapidly during the first 5 days of incubation, and then remained stable (136 mg l^{-1}). At 20°C with nutrients (T20Y),
211 there was only a slight decrease of Fe(II) throughout time, with a final concentration of 393 mg l^{-1} . At 35°C with
212 nutrient supply (T35Y), the decrease of Fe(II) concentration was delayed; however, a drastic decrease occurred
213 between day 5 and day 7, reaching a final concentration of 7 mg l^{-1} of Fe(II) (Fig. 2A).

214 Nutrient supply and higher temperature (T20Y, T35 and T35Y) induced a stronger decrease of As(III)
215 concentration compared to the 20°C treatment (T20) in the biotic batch experiments (Fig. 2B). The combination
216 of higher temperature and nutrient supply (T35Y) resulted in a fast and complete removal of As(III) after only
217 three days. Complete removal of As(III) was also achieved after 7 days in biotic batch incubated at 35°C without
218 nutrients (T35) and in biotic batch incubated at 20°C with nutrients (T20Y). In the T20 batch, only 67% of the
219 As(III) was removed from the dissolved phase in 8 days (Fig. 2B).

220 Under biotic conditions, the decrease of the dissolved Fe(II) concentration was closely related to Fe
221 precipitation. Consistent trends were observed for the two curves in the different batch experiments (Fig. 2A and
222 C). Slight differences between dissolved Fe(II) loss and precipitated Fe can be explained by the presence of few
223 dissolved Fe(III).

224 In abiotic batch experiments, the concentration of dissolved Fe(II) and As(III) decreased no more than
225 0.6 % and 19 % respectively (Fig. 2A and B).

226 **Iron and arsenic species in the particulate phase**

227 The amount of precipitates formed after eight days under biotic conditions was 4- to 11- fold higher than
228 under abiotic conditions (Table 1). In these biogenic precipitates, the As/Fe and As(III)/As(V) ratios showed wide
229 variations depending on the treatments applied compared to narrow variations observed in the particles formed
230 abiotically (Table 1). Nutrient-amended biotic batch incubated at 20°C (T20Y) exhibited a significantly higher
231 As/Fe ratio (0.60) than the other treatments ($\text{As/Fe} \leq 0.23$), which is consistent with the limited iron oxidation and
232 precipitation observed in these batch (Fig. 2A and 2C). In T20 batch, the particulate phase was As(III)-rich
233 (As(III)/As(V) ratio = 6.48) suggesting that the decrease of dissolved As(III) (Fig. 2B) was not due to its oxidation.
234 Conversely in T20Y, T35, T35Y batch experiments, the biogenic precipitates were As(V)-rich (average

235 As(III)/As(V) ratio \leq 0.43). For these treatments, the decrease of dissolved As(III) observed (Fig. 2B) may then
236 be linked to a possible As(III) oxidation. The T35Y treatment resulted in the formation of the biogenic precipitate
237 the more enriched in As(V) (As(III)/As(V) ratio= 0.02).

238 **Iron and arsenic oxidation**

239 The proportion of iron and arsenic oxidized during the incubation was calculated based on the
240 concentration of dissolved and particulate arsenic and iron species at the beginning and at the end of experiment
241 (Fig. 3). Under abiotic conditions, iron oxidation was negligible and As(III) oxidation did not exceed 6 %. Under
242 biotic conditions, the lowest proportion of Fe(II) oxidized was observed for the T20Y treatment (24%). No
243 biological As(III) oxidation was observed in T20 batch experiments. Conversely, the proportion of As(III) oxidized
244 exceeded 70% in the others treatments (T20Y, T35 and T35Y) (Fig. 3). The highest proportion of As(III) oxidized
245 was obtained in nutrient supplied batch incubated at 35°C (98%).

246 **Diversity of bacterial communities**

247 High-throughput sequencing yielded a total of 119,065 sequences of 16S-rRNA gene corresponding to
248 4353 quality sequences per sample which adequately covered the bacterial diversity in all the experiments (Table
249 2 and Fig. S1). For all the treatments, the bacterial diversity indices (Richness, Evenness and Shannon) decreased
250 significantly after incubation compared to the initial water collected at the Reigous spring. The batch incubated at
251 20°C without nutrients (T20) exhibited the lower level of bacterial diversity. The higher richness was observed in
252 the batch incubated at 35°C (Table 2).

253 NMDS analysis of the full bacterial-sequences datasets (Fig. 4) highlighted the establishment of distinct
254 bacterial community structures at the end of the batch experiments. ANOSIM test confirmed that the genetic
255 structures of the communities were significantly different ($R = 0.988$, $P = 0.001$). These differences were
256 associated with different taxonomic compositions (Fig. 5). In agreement with the diversity indices, bacterial
257 communities for all the treatments (T20, T20Y, T35 and T35Y) were characterized by the dominance of a small
258 number of OTUs. In the batch incubated at 20°C (T20), bacterial community was dominated by a single OTU
259 affiliated to the *Gallionella* genus representing an average of 82% of total sequences. Incubation at 35°C led to the
260 emergence of a dominant OTU affiliated to *Thiomonas* genus (49 to 79%), and to a lesser extent, to *Ferritrophicum*
261 genus (5 to 9%). The supply of yeast extract in batch incubated at 20°C mainly favored the development of bacteria
262 related to genera *Acidocella* (47 to 55 %), *Thiomonas* (7 to 14%) and *Gallionella* (8 to 10%). Finally, bacterial
263 community in the nutrient amended-batch incubated at 35°C was dominated by OTUs affiliated to *Acidicapsa* (13
264 to 55%), *Gallionella* (17 to 23%), *Thiomonas* (11 to 14%), *Acidocella* (0 to 24%) and *Ferritrophicum* (3 to 6%).
265 The bacterial groups that dominated at the end of the batch experiments represented no more than 10 to 20% of
266 the initial community of the Reigous spring water used in these experiments.

267 Batch incubations resulted in an enrichment in bacteria with the genetic potential for As(III) oxidation as
268 revealed by the quantification of *aioA* genes (Fig. 6). Nutrient supply or incubation at 35°C resulted in a one or
269 two order of magnitude higher abundance of *aioA* genes (representing on average $5 \pm 1 \times 10^5$, $2 \pm 1 \times 10^6$, $4 \pm 5 \times$
270 10^5 genes copies per ng of DNA for T20Y, T35 and T35Y, respectively) compared to the batch incubated at 20°C

- 271 (1.4 ± 0.4 × 10⁴ genes copies per ng of DNA) and to the initial water (4 ± 2 × 10³ genes copies per ng of DNA⁻¹).
- 272 The highest number of *aioA* gene copies was obtained as a consequence of the incubation at 35°C.

273 **Discussion**

274 We investigated the role of temperature and nutrients as drivers of the microbially mediated removal of
275 iron and arsenic in an As-rich AMD. We confirmed that indigenous microbial communities through their capacity
276 to oxidize iron and arsenic are the actors of the mitigation of the pollution in AMD (Casiot et al. 2003b; Egal et al.
277 2010; Mitsunobu et al. 2013). Abiotic oxidation and removal of iron and arsenic remained very limited without
278 microbial catalysis and were not influenced by nutrient supply or temperature increase.

279 **Bacterial diversity in AMD water and its evolution in batch experiments**

280 The initial bacterial diversity in the Carnoulès AMD water was similar to those previously described by
281 Volant and colleagues (2014). Batch incubations resulted in the decrease of diversity associated with the
282 preferential development of specific bacterial taxa. This can be explained by the inability of some microorganisms
283 to thrive under laboratory conditions (Koskella and Vos 2015), and by the strong competitiveness of other
284 microorganisms (Hibbing et al., 2010; Puspita et al., 2012). The reduction of diversity was moderated in batch
285 supplied with nutrients and in batch incubated at 35°C. We can hypothesize that a nutrient supply or a temperature
286 increase led to a diversification of bacterial niches (Hibbing et al. 2010; Koskella and Vos 2015; Okie et al. 2015),
287 promoting the co-existence of a greater number of taxa.

288 The incubation of AMD water under contrasted conditions of temperature and nutrient status led to the
289 establishment of distinct bacterial communities (in terms of diversity, taxonomic composition and functional
290 potential for As-oxidizing activity). These results confirm the influence of temperature and nutrients (C, N, etc.)
291 on the diversity and activity of microbial communities in diverse environments (Miller et al. 2009; Lawes et al.
292 2016), including AMD (Kuang et al. 2013, Volant et al. 2014). The development of these distinct bacterial
293 communities resulted in difference in terms of pollution removal efficiency and of composition of the biogenic
294 precipitates formed during incubation.

295 In batch incubated at 20°C, bacterial community was largely dominated by OTUs related to the iron-
296 oxidizing *Gallionella* (92% of total sequences), widely represented in iron-rich environments including AMD
297 (Bruneel et al. 2006; Volant et al. 2014). Dominance of this bacterial group was associated with complete Fe(II)
298 oxidation, and a low As(III) oxidation activity. This is in accordance with the formation of iron- and As(III)-rich
299 precipitates in the T20 batch.

300 **Effect of nutrient supply on bacterial communities and on pollution attenuation**

301 In agreement with other studies (Sipura et al. 2005; Leflaive et al. 2008), bacterial cell concentration
302 increased in nutrient amended batch. The supply of yeast extract stimulated the development of mixotrophic
303 *Thiomonas*. As a result, As(III) oxidation was stimulated and arsenic removal from the dissolved phase was
304 complete, both at 20 and 35°C. Members of the *Thiomonas* genus include *aioA*-carrying AsOB commonly found
305 in AMD-impacted environments (Bruneel et al. 2003; Battaglia-Brunet et al. 2006; Bryan et al. 2009). Conversely,
306 the rate of iron oxidation was partially inhibited in nutrient amended batch incubated at 20°C despite the presence
307 of FeOB like-*Gallionella* (10 % of total sequences). *Gallionella* spp. can grow autotrophically (Emerson et al.
308 2013) or mixotrophically (Hallbeck and Pedersen 1991). We can hypothesize that supply of yeast extract
309 (representing a diverse source of carbon and nitrogen) resulted in the inhibition of growth or activity of FeOB

310 *Gallionella* to the advantage of heterotrophic and mixotrophic bacteria such as *Acidocella*, *Thiomonas*,
311 *Rhodanobacter* and *Arthrobacter*. Among them, *Acidocella*, which is commonly found in iron rich environments
312 including AMD (Sheng et al. 2016), was predominant (52% of total sequences). *Acidocella* is an iron-reducing
313 bacteria (FeRB), able to use Fe(III) as the sole electron acceptor under anaerobic or oxygen limiting conditions
314 (Coupland and Johnson 2008; Lu et al. 2010). Interestingly, the inhibitive effect of yeast extract on iron oxidation
315 was counterbalanced when the batch were incubated at 35°C. In that case, Fe(II) oxidation was delayed but
316 complete. This efficient Fe removal in T35Y batch was possibly due the increasing proportion of bacteria related
317 to the *Ferritrophicum* genus which include iron-oxidizing bacteria (Gonzalez-Toril et al. 2011; Hedrich et al.
318 2011). The possible involvement of other group favored in T35Y including the moderately acidophilic and obligate
319 heterotroph *Acidicapsa* genus (Kulichevskaya et al. 2012) may not be excluded and would require further
320 investigation.

321 **Effect of temperature on bacterial communities and on pollution attenuation**

322 Enhanced arsenic oxidation and removal at 35°C compared to 20°C was not associated with an increase
323 of biomass. This suggests that bacterial activity was boosted rather than bacterial growth. Another possible
324 explanation is the preferential development at 35°C of populations efficient for arsenic removal. The stimulation
325 of As(III) oxidation at 35°C was associated with the large dominance of *Thiomonas* genus (61% of total sequences)
326 and a higher proportion of *aioA* genes. Several *Thiomonas* strains exhibit an optimum growth temperature of 30-
327 37°C (Kelly et al. 2007; Panda et al. 2009). These findings highlight the importance of temperature on both the
328 abundance and activity of AsOB like-*Thiomonas*, as shown for other AsOB (Ito et al. 2012; Debiec et al. 2017).
329 The second more abundant OTU was affiliated to the *Ferritrophicum* genus (8% of total sequences). The co-
330 occurrence of these two bacterial groups in batch incubated at 35°C can be linked to the formation of iron- and
331 As(V)-rich precipitates.

332 **Environmental significance**

333 The biogenic precipitates formed in batch without nutrient exhibited an As/Fe ratio ranging between 0.1-
334 0.2. Similar ranges (0.15-0.2) were obtained in a continuous flow reactor treating Carnoulès AMD water
335 (Fernandez-Rojo et al. 2017). These As/Fe ratios are lower than those observed *in situ* in the Reigous streambed
336 (0.4-0.7, Morin et al. 2003; Egal et al. 2010; Maillot et al. 2013). In the present batch experiments, conditions that
337 favored As oxidation while limiting Fe oxidation (T20Y) led to As/Fe ratio in the precipitate similar to field values.
338 This suggests that As oxidation is probably stimulated in the field, and Fe oxidation slowed down, compared to
339 laboratory conditions. The reason for such difference might be related to the inability of some microorganisms to
340 develop or maintain their activity under laboratory conditions.

341 Seasonal variations of As speciation were observed in Carnoulès AMD (Morin et al. 2003; Egal et al.
342 2010) with the preferential formation of As(III)-rich precipitates during the coldest season, mainly in the form of
343 tooeleite. Conversely, during summer, As(V) dominates in the amorphous ferric arsenate form. These field
344 observations are in agreement with our results that clearly revealed a positive effect of higher temperature on
345 microbially mediated As(III) oxidation. As a consequence, in case of an *in situ* biological treatment, contrasted
346 performance in term of efficiency of arsenic removal and sludge composition is expected depending on the climate

347 or the season. The formation of stable As(V)-rich biogenic precipitates might be favored at temperatures higher
348 than 20°C.

349 Nutrient supply favored the formation of As(V) rich precipitates, which are preferred to As(III) solid
350 phases in AMD treatment process due to their stability upon storage (Palfy et al. 1999). However, the stimulation
351 of iron-reducing bacteria *Acidocella* may result in the remobilization of arsenic if conditions became reducing or
352 in anoxic micro niches (Héry et al. 2014). For this reason, possible use of organic matter amendement in aerobic
353 AMD treatment has to be carefully considered.

354 Our results evidenced clear links between taxonomic composition of bacterial community, abundance of
355 *aioA* gene, and iron and arsenic oxidation and removal from AMD water. They give new insights into the
356 regulation by temperature and nutrients of microbially mediated processes involved in natural pollution
357 attenuation. Studies focusing on single strains did not take into account the metabolic and functional diversity
358 present in natural ecosystems. On the contrary our study based on a complex indigenous community integrates all
359 the differential effects the applied treatment may have on the different bacterial populations co-existing among the
360 community. Then, our approach gives a more representative picture of what could occur *in situ* under temperature
361 or nutrient status changing conditions than previous studies on single Fe(II)- or As(III)-oxidizing bacterial strains.

362

363 **Acknowledgements**

364 This work was supported by the Agence Nationale de Recherche (ANR) as part of ANR IngECOST-DMA project
365 (ANR-13-ECOT-0009) and the OSU OREME. It benefited from the technical facilities of Get-PlaGe platform
366 (<http://get.genotoul.fr/>), Montpellier RIO Imaging microscopy platform (<https://www.mri.cnrs.fr/>) and the AETE-
367 ISO Platform (OSU REME, Université de Montpellier).

368 **Compliance with Ethical Standards**

369 Conflict of Interest: The authors declare they have no conflict of interest.

370 Ethical approval: This article does not contain any studies with human participants or animals performed by any
371 of the authors.

372

373 **References**

- 374 Ahoranta SH, Kokko ME, Papirio S, Özkaya B, Puhakka JA (2016) Arsenic removal from acidic solutions with
375 biogenic ferric precipitates. *J Hazard Mater* 306:124–132. doi: 10.1016/j.jhazmat.2015.12.012
- 376 Asta MP, Ayora C, Román-Ross G, Cama J, Acero P, Gault AG, Charnock JM, Bardelli F (2010) Natural
377 attenuation of arsenic in the Tinto Santa Rosa acid stream (Iberian Pyritic Belt, SW Spain): The role of iron
378 precipitates. *Chem Geol* 271:1–12. doi: 10.1016/j.chemgeo.2009.12.005
- 379 Barret M, Briand M, Bonneau S, Prévieux A, Valière S, Bouchez O, Hunault G, Simoneau P, Jacquesa MA (2015)
380 Emergence shapes the structure of the seed microbiota. *Appl Environ Microbiol* 81:1257–1266. doi:
381 10.1128/AEM.03722-14
- 382 Battaglia-Brunet F, Dictor MC, Garrido F, Crouzet C, Morin D, Dekeyser K, Clarens M, Baranger P (2002) An
383 arsenic (III)-oxidizing bacterial population: selection, characterization, and performance in reactors. *J Appl*
384 *Microbiol* 93:656–667.
- 385 Battaglia-Brunet F, Joulian C, Garrido F, Dictor MC, Morin D, Coupland K, Barrie Johnson D, Hallberg KB,
386 Baranger P (2006) Oxidation of arsenite by *Thiomonas* strains and characterization of *Thiomonas*
387 *arsenivorans* sp. nov. *Antonie van Leeuwenhoek*, 89:99–108. doi: 10.1007/s10482-005-9013-2
- 388 Bruneel O, Duran R, Casiot C, Elbaz-Poulichet F, Personné JC (2006) Diversity of microorganisms in Fe-As-rich
389 acid mine drainage waters of Carnoulès, France. *Appl Environ Microbiol* 72:551–556. doi:
390 10.1128/AEM.72.1.551-556.2006
- 391 Bruneel O, Personné JC, Casiot C, Leblanc M, Elbaz-Poulichet F, Mahler BJ, Le Flèche A, Grimont PAD (2003)
392 Mediation of arsenic oxidation by *Thiomonas* sp. in acid-mine drainage (Carnoulès, France). *J Appl*
393 *Microbiol* 95:492–499. doi: 10.1046/j.1365-2672.2003.02004.x
- 394 Bryan CG, Marchal M, Battaglia-Brunet F, Kugler V, Lemaitre-Guillier C, Lièremont D, Bertin PN, Arsène-
395 Ploetze F (2009) Carbon and arsenic metabolism in *Thiomonas* strains: differences revealed diverse
396 adaptation processes. *BMC Microbiol* 9:127. doi: 10.1186/1471-2180-9-127
- 397 Campbell KM, Kirk Nordstrom D (2014) Arsenic speciation and sorption in natural environments. *Rev Mineral*
398 *Geochem* 79:185–216. doi: 10.1128/9781555817510.ch5
- 399 Casiot C, Leblanc M, Bruneel O (2003a) Geochemical Processes Controlling the Formation of As-rich waters
400 within a Tailings Impoundment (Carnoulès, France). *Aquat Geochem* 9:273–290. doi:
401 10.1023/B:AQUA.0000028985.07557.39
- 402 Casiot C, Morin G, Juillot F, Bruneel O, Personné JC, Leblanc M, Duquesne K, Bonnefoy V, Elbaz-Poulichet F
403 (2003b) Bacterial immobilization and oxidation of arsenic in acid mine drainage (Carnoulès creek, France).
404 *Water Res* 37:2929–2936. doi: 10.1016/S0043-1354(03)00080-0
- 405 Cheng H, Hu Y, Luo J, Xu B, Zhao J (2009) Geochemical processes controlling fate and transport of arsenic in
406 acid mine drainage (AMD) and natural systems. *J Hazard Mater* 165:13–26. doi:
407 10.1016/j.jhazmat.2008.10.070
- 408 Coupland K, Johnson DB (2008) Evidence that the potential for dissimilatory ferric iron reduction is widespread
409 among acidophilic heterotrophic bacteria. *FEMS Microbiol Lett* 279:30–35. doi: 10.1111/j.1574-
410 6968.2007.00998.x
- 411 Debiec K, Krzysztoforski J, Uhrynowski W, Sklodowska A, Drewniak L (2017) Kinetics of arsenite oxidation by
412 *Sinorhizobium* sp. M14 under changing environmental conditions. *Int Biodeterior Biodegradation* 119:476–

413 485. doi: 10.1016/j.ibiod.2016.10.049

414 Dopson M, Halinen A-K, Rahunen N, Ozkaya B, Sahinkaya E, Kaksonen AH, Lindström EB, Puhakka JA (2006)

415 Mineral and Iron Oxidation at Low Temperatures by Pure and Mixed Cultures of Acidophilic

416 Microorganisms. *Biotechnol Bioeng* 97:1205–1215. doi: 10.1002/bit.21312

417 Duquesne K, Lebrun S, Casiot C, Bruneel O, Personné JC, Leblanc M, Morin G, Bonnefoy V (2003)

418 Immobilization of Arsenite and Ferric Iron by *Acidithiobacillus ferrooxidans* and Its Relevance to Acid Mine

419 Drainage. *Appl Environ Microbiol* 69:6165–6173. doi: 10.1128/AEM.69.10.6165

420 Edgar RC, Haas BJ, Clemente JC, Quince C, Knight R (2011) UCHIME improves sensitivity and speed of chimera

421 detection. *Bioinformatics* 27:2194–2200. doi: 10.1093/bioinformatics/btr381

422 Egal M, Casiot C, Morin G, Elbaz-Poulichet F, Cordier MA, Bruneel O (2010) An updated insight into the natural

423 attenuation of As concentrations in Reigous Creek (southern France). *Appl Geochem* 25:1949–1957. doi:

424 10.1016/j.apgeochem.2010.10.012

425 Egal M, Casiot C, Morin G, Parmentier M, Bruneel O, Lebrun S, Elbaz-Poulichet F (2009) Kinetic control on the

426 formation of tooeleite, schwertmannite and jarosite by *Acidithiobacillus ferrooxidans* strains in an As(III)-

427 rich acid mine water. *Chem Geol* 265:432–441. doi: 10.1016/j.chemgeo.2009.05.008

428 Emerson D, Field EK, Chertkov O, Davenport KW, Goodwin L, Munk C, Nolan M, Woyke T (2013) Comparative

429 genomics of freshwater Fe-oxidizing bacteria: Implications for physiology, ecology, and systematics. *Front*

430 *Microbiol* 4:1–17. doi: 10.3389/fmicb.2013.00254

431 Fernandez-Rojo L, Héry M, Le pape P, Braungardt C, Desoeuvre A, Torres E, Tardy V, Resongles E, Laroche E,

432 Delpoux S, Jouliau C, Battaglia-Brunet F, Boisson J, Grapin G, Morin G, Casiot C (2017) Biological

433 attenuation of arsenic and iron in a continuous flow bioreactor treating acid mine drainage (AMD). *Water*

434 *Res* 123:594-606. doi: 10.1016/j.watres.2017.06.059

435 Fukushi K, Sasaki M, Sato T, Yanase N, Amano H, Ikeda H (2003) A natural attenuation of arsenic in drainage

436 from an abandoned arsenic mine dump. *Appl Geochem* 18:1267–1278. doi: 10.1016/S0883-2927(03)00011-

437 8

438 Garcia-Dominguez E, Mumford A, Rhine ED, Paschal A, Young LY (2008) Novel autotrophic arsenite-oxidizing

439 bacteria isolated from soil and sediments. *FEMS Microbiol Ecol* 66:401–410. doi: 10.1111/j.1574-

440 6941.2008.00569.x

441 Gonzalez-Toril E, Aguilera A, Souza-Egipsy V, Pamo EL, Espana JS, Amils R (2011) Geomicrobiology of La

442 Zarza-Perrunal acid mine effluent (Iberian Pyritic Belt, Spain). *Appl Environ Microbiol* 77:2685–2694. doi:

443 10.1128/AEM.02459-10

444 Hallbeck L, Pedersen K (1991) Autotrophic and mixotrophic growth of *Gallionella ferruginea*. *J Gen Microbiol*

445 137:2657–2661. doi: 10.1099/00221287-137-11-2657

446 Hedrich S, Schlomann M, Johnson DB (2011) The iron-oxidizing proteobacteria. *Microbiology* 157:1551–1564.

447 Héry M, Casiot C, Resongles E, Gallice Z, Bruneel O, Desoeuvre A, Delpoux S (2014) Release of arsenite, arsenate

448 and methyl-arsenic species from streambed sediment affected by acid mine drainage: A microcosm study.

449 *Environ Chem* 11:514–524. doi: 10.1071/EN13225

450 Hibbing ME, Fuqua C, Parsek MR, Peterson SB (2010) Bacterial competition: surviving and thriving in the

451 microbial jungle. *Natl Rev Microbiol* 8:15–25. doi: 10.1038/nrmicro2259.Bacterial

452 Hug SJ, Leupin OX (2003) Iron-catalyzed oxidation of arsenic (III) by oxygen and by hydrogen peroxide: pH-

453 dependent formation of oxidants in the Fenton reaction. *Environ Sci Technol* 37:2734–2742.

454 Ito A, Miura JI, Ishikawa N, Umita T (2012) Biological oxidation of arsenite in synthetic groundwater using
455 immobilised bacteria. *Water Res* 46:4825–4831. doi: 10.1016/j.watres.2012.06.013

456 Johnson DB, Hallberg KB (2005) Acid mine drainage remediation options: A review. *Sci Total Environ* 338:3–
457 14. doi: 10.1016/j.scitotenv.2004.09.002

458 Kelly DP, Uchino Y, Huber H, Amils R, Wood AP (2007) Reassessment of the phylogenetic relationships of
459 *Thiomonas cuprina*. *Int J Syst Evol Microbiol* 57:2720–2724. doi: 10.1099/ijs.0.65537-0

460 Kim D-J, Pradhan D, Park K-H, Ahn J-G, Lee S-W (2008) Effect of pH and Temperature on Iron Oxidation by
461 Mesophilic Mixed Iron Oxidizing Microflora. *Mater Trans* 49:2389–2393. doi:
462 10.2320/matertrans.MER2008051

463 Koskella B, Vos M (2015) Adaptation in natural microbial populations. *Annu Rev Ecol Evol Syst* 46:503–522.
464 doi: 10.1146/annurev-ecolsys-112414-054458

465 Kuang J-L, Huang L-N, Chen L-X, Hua Z-S, Li S-J, Hu M, Li J-T, Shu W-S (2013) Contemporary environmental
466 variation determines microbial diversity patterns in acid mine drainage. *ISME J* 7:1038–50. doi:
467 10.1038/ismej.2012.139

468 Kulichevskaya IS, Kostina LA, Valášková V, Rijpstra WIC, Sinninghe Damsté JS, de Boer W, Dedysch SN (2012)
469 *Acidicapsa borealis* gen. nov., sp. nov. and *Acidicapsa ligni* sp. nov., subdivision 1 Acidobacteria from
470 *Sphagnum* peat and decaying wood. *Int J Syst Evol Microbiol* 62, 1512-1520. doi: 10.1099/ijs.0.034819-0

471 Lawes JC, Neilan BA, Brown MV, Clark GF, Johnston EL (2016) Elevated nutrients change bacterial community
472 composition and connectivity: high throughput sequencing of young marine biofilms. *Biofouling* 32:57–69.
473 doi: 10.1080/08927014.2015.1126581

474 Leflaive J, Danger M, Lacroix G, Lyautey E, Oumarou C, Ten-Hage L (2008) Nutrient effects on the genetic and
475 functional diversity of aquatic bacterial communities. *FEMS Microbiol Ecol* 66:379–390. doi:
476 10.1111/j.1574-6941.2008.00593.x

477 Lozupone C, Knight R (2005) UniFrac : a New Phylogenetic Method for Comparing Microbial Communities.
478 *Appl Environ Microbiol* 71:8228–8235. doi: 10.1128/AEM.71.12.8228

479 Lu S, Gischkat S, Reiche M, Akob DM, Hallberg KB, Küsel K (2010) Ecophysiology of Fe-cycling bacteria in
480 acidic sediments. *Appl Environ Microbiol* 76:8174–8183. doi: 10.1128/AEM.01931-10

481 Lunau M, Lemke A, Walther K, Martens-Habbena W, Simon M (2005) An improved method for counting bacteria
482 from sediments and turbid environments by epifluorescence microscopy. *Environ Microbiol* 7:961–968. doi:
483 10.1111/j.1462-2920.2005.00767.x

484 Magoc T, Salzberg SL (2011) FLASH: Fast length adjustment of short reads to improve genome assemblies.
485 *Bioinformatics* 27:2957–2963. doi: 10.1093/bioinformatics/btr507

486 Maillot F, Morin G, Juillot F, Bruneel O, Casiot C, Ona-Nguema G, Wang Y, Lebrun S, Aubry E, Vlais G, Brown
487 GE (2013) Structure and reactivity of As(III)- and As(V)-rich schwertmannites and amorphous ferric
488 arsenate sulfate from the Carnoulès acid mine drainage, France: Comparison with biotic and abiotic model
489 compounds and implications for As remediation. *Geochim Cosmochim Acta* 104:310–329. doi:
490 10.1016/j.gca.2012.11.016

491 Méndez-García C, Peláez AI, Mesa V, Sánchez J, Golyshina OV, Ferrer M (2015) Microbial diversity and
492 metabolic networks in acid mine drainage habitats. *Front Microbiol* 6:475. doi: 10.3389/fmicb.2015.00475

493 Miller SR, Strong AL, Jones KL, Ungerer MC (2009) Bar-coded pyrosequencing reveals shared bacterial
494 community properties along the temperature gradients of two alkaline hot springs in Yellowstone National
495 Park. *Appl Environ Microbiol* 75:4565–4572. doi: 10.1128/AEM.02792-08

496 Mitsunobu S, Hamanura N, Kataoka T, Shiraishi F (2013) Arsenic attenuation in geothermal streamwater coupled
497 with biogenic arsenic(III) oxidation. *Appl Geochemistry* 35:154–160. doi:
498 10.1016/j.apgeochem.2013.04.005

499 Morin G, Juillot F, Casiot C, Bruneel O, Personné JC, Elbaz-Poulichet F, Leblanc M, Ildefonse P, Calas G (2003)
500 Bacterial formation of tooeleite and Mixed Arsenic(III) or Arsenic(V) - Iron(III) gels in the carnoulès acid
501 mine drainage, France. A XANES, XRD, and SEM study. *Environ Sci Technol* 37:1705–1712. doi:
502 10.1021/es025688p

503 Nordstrom DK, Alpers CN (1999) Negative pH, efflorescent mineralogy, and consequences for environmental
504 restoration at the Iron Mountain Superfund site, California. *Proc Natl Acad Sci U S A* 96:3455–3462. doi:
505 10.1073/pnas.96.7.3455

506 Okie JG, Van Horn DJ, Storch D, Barrett JE, Gooseff MN, Kopsova L, Takacs-Vesbach CD (2015) Niche and
507 metabolic principles explain patterns of diversity and distribution: theory and a case study with soil bacterial
508 communities. *Proc R Soc B Biol Sci* 282:20142630. doi: 10.1098/rspb.2014.2630

509 Paikaray S (2015) Arsenic Geochemistry of Acid Mine Drainage. *Mine Water Environ* 34:181–196. doi:
510 10.1007/s10230-014-0286-4

511 Palfy P, Vircikova E, Molnar L (1999) Processing of arsenic waste by precipitation and solidification. *Waste*
512 *Manag* 19:55–59. doi: 10.1016/S0956-053X(99)00014-8

513 Panda SK, Jyoti V, Bhadra B, Nayak KC, Shivaji S, Rainey FA, Das SK (2009) *Thiomonas bhubaneswarensis* sp.
514 nov., an obligately mixotrophic, moderately thermophilic, thiosulfate-oxidizing bacterium. *Int J Syst Evol*
515 *Microbiol* 59:2171–2175. doi: 10.1099/ijs.0.007120-0

516 Puspita ID, Kamagata Y, Tanaka M, Asano K, Nakatsu CH (2012) Are Uncultivated Bacteria Really Uncultivable?
517 *Microbes Environ.* 27: 356–366. doi:10.1264/jsme2.ME12092

518

519 Quéméneur M, Cébron A, Billard P, Battaglia-Brunet F, Garrido F, Leyval C, Joulian C (2010) Population
520 structure and abundance of arsenite-oxidizing bacteria along an arsenic pollution gradient in waters of the
521 upper isle river basin, France. *Appl Environ Microbiol* 76:4566–4570. doi: 10.1128/AEM.03104-09

522 Resongles E, Le Pape P, Fernandez-Rojo L, Morin G, Delpoux S, Brest J, Guo S, Casiot C (2016) Routine
523 determination of inorganic arsenic speciation in precipitates from acid mine drainage using orthophosphoric
524 acid extraction followed by HPLC-ICP-MS. *Anal Methods* 8:7420–7426. doi: 10.1039/c6ay02084d

525 Rodier J (1996) *L'analyse de l'eau, eaux résiduaires, eau de mer*, 8^{ème} Edition. Dénod Paris

526 Santini JM, Sly LI, Schnagl RD, Macy JM (2000) A new chemolithoautotrophic arsenite-oxidizing bacterium
527 isolated from a gold mine: Phylogenetic, physiological, and preliminary biochemical studies. *Appl Environ*
528 *Microbiol* 66:92–97. doi: 10.1128/aem.66.1.92-97.2000

529 Schloss PD, Westcott SL, Ryabin T, Hall JR, Hartmann M, Hollister EB, Lesniewski RA, Oakley BB, Parks DH,
530 Robinson CJ, Sahl JW, Stres B, Thallinger GG, Van Horn DJ, Weber CF (2009) Introducing mothur: Open-
531 source, platform-independent, community-supported software for describing and comparing microbial
532 communities. *Appl Environ Microbiol* 75:7537–7541. doi: 10.1128/AEM.01541-09

533 Sheng Y, Bibby K, Grettenberger C, Kaley B, Macalady JL, Wang G, Burgos WD (2016) Geochemical and
534 temporal influences on the enrichment of acidophilic iron-oxidizing bacterial communities. *Appl Environ*
535 *Microbiol* 82:3611–3621. doi: 10.1128/AEM.00917-16

536 Sipura J, Haukka K, Helminen H, Lagus A, Suomela J, Sivonen K (2005) Effect of nutrient enrichment on
537 bacterioplankton biomass and community composition in mesocosms in the Archipelago Sea, northern
538 Baltic. *J Plankton Res* 27:1261–1272. doi: 10.1093/plankt/fbi092

539 Slyemi D, Moinier D, Brochier-Armanet C, Bonnefoy V, Johnson DB (2011) Characteristics of a phylogenetically
540 ambiguous, arsenic-oxidizing *Thiomonas* sp., *Thiomonas arsenitoxydans* strain 3AsT sp. nov. *Arch*
541 *Microbiol* 193:439–449. doi: 10.1007/s00203-011-0684-y

542 Volant A, Bruneel O, Desoeuvre A, Héry M, Casiot C, Bru N, Delpoux S, Fahy A, Javerliat F, Bouchez O, Duran
543 R, Bertin PN, Elbaz-Poulichet F, Lauga B (2014) Diversity and spatiotemporal dynamics of bacterial
544 communities: physicochemical and others drivers along an acid mine drainage. *FEMS Microbiol Ecol*
545 90:247–263. doi: 10.1111/1574-6941.12394

546 Wang Q, Garrity GM, Tiedje JM, Cole JR (2007) Naïve Bayesian classifier for rapid assignment of rRNA
547 sequences into the new bacterial taxonomy. *Appl Environ Microbiol* 73:5261–5267. doi:
548 10.1128/AEM.00062-07

549 Wang Y, Qian PY (2009) Conservative fragments in bacterial 16S rRNA genes and primer design for 16S
550 ribosomal DNA amplicons in metagenomic studies. *PLoS One* 4 (10):e70401. doi:
551 10.1371/journal.pone.0007401

552

553 List of figure legends

554 **Fig. 1** Bacterial biomass in abiotic (n=5) and biotic (n=8) batch experiment at the beginning (T0) and at the end
555 of the experiment (T8). Symbols inside boxplots represent the treatments applied (T20: batch incubated at 20°C;
556 T35: batch incubated at 35°C; T20Y: batch incubated at 20°C with nutrients; T35Y: batch incubated at 35°C with
557 nutrients). Different letters in brackets indicate statistically significant differences ($P < 0.05$) according to *Kruskal-*
558 *Wallis* test.

559 **Fig. 2** Evolution of dissolved Fe(II) (A), dissolved As(III) (B) and precipitated Fe concentrations (C) during
560 incubation time for abiotic (empty symbols) and biotic (lines with full symbols) batch experiment for all treatments
561 (T20: batch incubated at 20°C; T35: batch incubated at 35°C; T20Y: batch incubated at 20°C with nutrients; T35Y:
562 batch incubated at 35°C with nutrients).

563 **Fig. 3** Proportion of Fe(II) (A) and As(III) (B) oxidized after 8 days of incubation relatively to Fe(II) and As(III)
564 concentration in the Carnoulès AMD at t_0 for both abiotic and biotic treatments (T20: batch incubated at 20°C;
565 T35: batch incubated at 35°C; T20Y: batch incubated at 20°C with nutrients; T35Y: batch incubated at 35°C with
566 nutrients). For biotic treatments, values with different letters differ significantly ($P < 0.05$) according to *Fisher*
567 test.

568 **Fig. 4** Non-metric multi-dimensional scaling (NMDS) ordination plot derived from weighted pairwise Unifrac
569 distances for bacterial communities for each treatment (S: Source; T20: batch incubated at 20°C; T35: batch
570 incubated at 35°C; T20Y: batch incubated at 20°C with nutrients; T35Y: batch incubated at 35°C with nutrients).
571 Stress values for ordination plot were < 0.2 which indicates that these data were well-represented by the two
572 dimensional representation.

573 **Fig. 5** Relative abundance of bacterial genera in the water collected at the Reigous stream and in the batch at the
574 end of the incubations. All the analyses were performed in triplicates (S: Source; T20: batch incubated at 20°C;
575 T35: batch incubated at 35°C; T20Y: batch incubated at 20°C with nutrients; T35Y: batch incubated at 35°C with
576 nutrients). Cluster tree represent phylogenetic community distance based on the OTU composition. Other groups
577 represent the phylogenetic groups (genus) with a relative abundance $< 1\%$ calculated on the whole dataset.
578 Asterisks represent phylogenetic group affiliated to higher taxonomic levels.

579 **Fig. 6** Quantification of *aioA* genes in the Reigous water and in the biotic batch at the end of the incubations (n=3)
580 with S: Source corresponding to water collected at the Reigous stream used in batch experiments; T20: batch
581 incubated at 20°C; T35: batch incubated at 35°C; T20Y: batch incubated at 20°C with nutrients; T35Y: batch
582 incubated at 35°C with nutrients. Letters in brackets indicate significant differences between treatments, according
583 to *Kruskal-Wallis* test ($P < 0.05$).

584