**Adsorption and photocatalytic oxidation of ibuprofen using nanocomposites of TiO₂
nanofibers combined with BN nanosheets: degradation products and mechanisms**

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Abstract

This study investigated the adsorption and photocatalytic activity of TiO₂-boron nitride (BN) nanocomposites for the removal of contaminants of emerging concern in water using ibuprofen as a model compound. TiO₂ nanofibers wrapped by BN nanosheets were synthesized by electrospinning method. Characterization of the nanocomposite photocatalysts indicated the BN nanosheets improved the light absorbance and reduced the recombination of the photoexcited charge carriers (e⁻ and h⁺). The photocatalytic oxidation products and mechanisms of ibuprofen by the TiO₂-BN catalysts were elucidated using a multiple analysis approach by high performance liquid chromatography, ultraviolet absorbance, dissolved organic carbon, fluorescence excitation-emission matrices, and electrospray ionization–liquid chromatography–tandem mass spectrometry. The experimental results revealed that the photocatalytic oxidation by the TiO₂-BN nanocomposites is a multi-step process and the interactions between ibuprofen molecules and the TiO₂-BN nanocomposites govern the adsorption process. The increasing BN nanosheet content in the TiO₂ nanofibers facilitated the breakdown of ibuprofen degradation intermediates (hydroxyibuprofen, carboxyibuprofen, and oxypropyl ibuprofen). Kinetic modeling indicated both adsorption and photocatalytic oxidation of ibuprofen by the TiO₂-BN nanocomposites followed the first-order kinetic model. The photocatalytic oxidation rate increased with the increasing BN content in the nanocomposite catalysts, which was attributed to the light absorption capacity and the separation efficiency of the photoexcited electron (e⁻)-hole (h⁺) pairs. Multiple photocatalytic cycles were conducted to investigate the reusability and regeneration of the nanofibers for degradation of ibuprofen.

42   **Keywords:** titanium dioxide boron-nitride nanocomposites; photocatalytic oxidation; adsorption;
43   degradation intermediates; photocatalytic degradation mechanisms

1. Introduction

Contaminants of emerging concerns such as pharmaceuticals and personal care products (PPCPs) have been widely detected in water bodies around the world (Vieno et al., 2007). Many PPCPs are of serious concerns even in trace amounts due to their toxicity to aquatic life and potential risks to public health (Vieno et al., 2007; Christen et al., 2010). Conventional water and wastewater treatment processes including coagulation, sedimentation, media filtration, and biological processes are not effective to remove these recalcitrant organic contaminants (Vieno et al., 2007; Xiang et al., 2016). Adsorption by activated carbon is an efficient technique to remove hydrophobic PPCPs through hydrophobic interactions, however the adsorption capacity gradually decreases as the carbon is saturated with the adsorbed chemicals. In addition, it is ineffective to adsorb hydrophilic compounds such as atenolol, acetaminophen, and naproxen (Vieno et al., 2007; Huerta-Fontela et al., 2011; Wang et al., 2015). Oxidation using chlorine, chlorine dioxide, and ozone is typically effective to degrade PPCPs with electron-donating functional groups (e.g., phenolic-and amine-containing compounds) (Lee and von Gunten, 2010). Ultraviolet (UV) irradiation can degrade many PPCPs but treatment cost is inhibitive because of high energy demand (Rosenfeldt and Linden, 2004).

To degrade the persistent organic contaminants, advanced oxidation processes (AOPs) are typically needed (De la Cruz et al., 2012). AOPs generate hydroxyl radicals ($\cdot\text{OH}$) which react non-selectively with most organic compounds (Andreozzi et al., 1999). Commonly used AOPs for PPCPs removal from wastewater include UV/H₂O₂, UV/O₃, (photo)-Fenton, and heterogeneous photocatalysis (De Witte et al., 2009; Homem and Santos, 2011; Prieto-Rodriguez et al., 2012; Carbonaro et al., 2013; Van Doorslaer et al., 2015; Thagard et al., 2016). Complete mineralization is usually not the goal of AOPs because it is energy intensive and not cost-

effective (Doll and Frimmel, 2004). Partial degradation of recalcitrant PPCPs is more economically attractive to deactivate their biological activity or increase their biodegradability (Van Doorslaer et al., 2015). But more concerns are raised regarding the intermediate degradation products and their environmental toxicities.

Heterogeneous photocatalytic oxidation using metal oxides (e.g., TiO_2) has attracted considerable attention particularly due to its potentially lower cost than other AOPs (Jing et al., 2006; Dalrymple et al., 2007; Rubio et al., 2013; Lin et al., 2015; Arlos et al., 2016; Lin et al., 2016; Lin et al., 2017b; Lin et al., 2017f; Li and Hu, 2018). More than 90 organic and 25 inorganic compounds catalogued on the US Environmental Protection Agency's (USEPA) priority list of contaminants have been investigated using heterogeneous photocatalysis (Blake, 2001), such as acetone, aniline, atrazine, benzene, methyl tertiary butyl ether (MTBE), trichloroethane. The process has also been successfully applied for the treatment of organic and inorganic compounds at low concentrations in aqueous solutions (Wold, 1993; Nakajima et al., 2005; Westerhoff et al., 2005; Le-Clech et al., 2006).

During heterogeneous photocatalysis, organic molecules adsorb onto a metal oxide catalyst surface and react with the photoexcited charge carriers (e^- and h^+) or free radicals (e.g., $\cdot\text{OH}$) (Ollis et al., 1984; Kormann et al., 1991). The photocatalytic reaction depends on the species of organic molecules adsorbed onto the photocatalyst surface and in aqueous phase, which results in different kinetics of photodegradation. As a critical step in heterogeneous photocatalysis, very few studies investigated the kinetics of organics adsorption on a photocatalyst (Bauer et al., 2001; Vautier et al., 2001; Hu et al., 2003; Yu et al., 2005; Lee et al., 2011; Zhang et al., 2013; Rioja et al., 2014; Lv et al., 2016).

89 Titanium dioxide (TiO₂) is the most commonly used photocatalyst due to its relatively low cost,
90 high stability, and low toxicity (Fujishima et al., 2000; Kumar and Devi, 2011a; Nakata and
91 Fujishima, 2012). Substantial efforts have been devoted to improve the photocatalytic
92 degradation efficiency of TiO₂ through combination with metal ions, non-metal ions, and some
93 2D nanomaterials (Liqiang et al., 2006; Rubio et al., 2013; Lin et al., 2015; Arlos et al., 2016;
94 Lin et al., 2016; Lin et al., 2017b; Lin et al., 2017f; Li and Hu, 2018; Yi et al., 2018; Merenda et
95 al., 2019). These dopants, such as Ag (Harifi and Montazer, 2014), Fe (Asiltürk et al., 2009; Lin
96 et al., 2016), P (Zhang et al., 2012), SiO₂ (Tawkaew and Supothina, 2008), and graphene (oxide)
97 (Lin et al., 2017c; Lin et al., 2017d; Yi et al., 2018), have been proven to be effective to improve
98 the photocatalytic activity. A recent study reported that the boron nitride (BN) nanosheets exhibit
99 unique properties due to the high surface area and the reactive edge structure (Biscarat et al.,
100 2015). Previous study demonstrated that the introduction of BN nanosheets enhanced the
101 separation of electron–hole pairs of TiO₂ and dramatically improved photocatalytic activity of
102 methyl orange under ultraviolet (UV) irradiation (Nasr et al., 2017b).

103 This study aims to investigate the mechanisms and kinetics of adsorption and photocatalytic
104 oxidation of PPCPs using novel photocatalytic composite TiO₂ nanofibers with boron nitride
105 (BN) nanosheets. Combining TiO₂ with BN nanosheets by electrospinning method is anticipated
106 to enhance the separation of e⁻-h⁺ pairs of TiO₂ thus to improve the photocatalytic activity.
107 Ibuprofen, a nonsteroidal anti-inflammatory drug that has been found extensively in wastewater
108 effluents, was used as the target recalcitrant PPCP. The adsorption and degradation of ibuprofen
109 by photolysis and photocatalysis was characterized by a series of analytical methods.
110 Degradation intermediates of ibuprofen were analyzed by a positive ion electrospray ionization–
111 liquid chromatography–tandem mass spectrometry [(+) ESI LC-MS/MS]. Multiple

photocatalytic cycles were conducted to evaluate the recyclability and regeneration of the TiO₂-BN nanocomposites for the degradation of ibuprofen.

2. Materials and methods

2.1 Materials and characterization

A series of BN nanosheets incorporated TiO₂ nanofibers (0, 3%, 5%, 7%, and 10%, weight percentage of BN to the mass of Ti, referred as TiO₂, TB1, TB2, TB3, and TB4, respectively) were synthesized using electrospinning technique as described by Nasr et al. (2017b). UV-Vis absorbance of the nanofibers was measured by a spectrophotometer (DR6000; Hach Company, CO). X-ray diffraction (XRD) measurements were conducted using a PANalytical Xpert-PRO diffractometer equipped with an X'celerator detector using Ni-filtered Cu-radiation. The specific surface area of the nanocomposites was determined from the nitrogen adsorption-desorption isotherms measured at liquid nitrogen temperature using a Micromeritics ASAP 2010 (outgassing conditions: 200 °C and 12 h). The structure and morphology of the nanocomposite catalysts was characterized by an H-7650 transmission electron microscope (TEM; Hitachi High-Technologies Corp., Pleasanton, CA). The elemental mapping images were taken with a Zeiss EVO HD15 microscope coupled with an Oxford X-MaxN EDX detector.

2.2 Adsorption and photodegradation experiments

Ibuprofen (Analytical grade; Acros Organics Co, NJ) was used as a model compound representative of PPCPs to evaluate the adsorption and photocatalytic activity of the synthesized nanocomposites. The pH of the tested ibuprofen solution was neutral throughout the experiments.

Batch experiments were conducted in 100 mL beakers containing a suspension of 10 mg nanofibers and 50 mL ibuprofen solution (5 mg L⁻¹) under dark (adsorption), UV light irradiation with nanofibers (photocatalysis) or without nanofibers (photolysis) conditions. The irradiance of UV lamp (160W PUV-10, Zoo Med Laboratories, San Luis Obispo, CA) included both UV (365 nm, minor peaks at 290, 315, 335 nm) and visible light (405, 435, and 545 nm) wavelength spectrum. The solution mixtures were magnetically stirred for 2 to 3 hours, and 1 mL water samples were taken at 0, 10, 20, 30, 60, 90, 120, and 180 min. The suspensions were then filtered through 0.45 µm cellulose acetate membrane to separate nanofibers for analysis. All adsorption, photolysis, and photocatalytic degradation experiments under each condition were conducted at least two times to ensure the reproducibility of the results.

Repeated ibuprofen degradation experiments were conducted for 16 cycles to evaluate the recyclability of the TiO₂-BN nanocomposites under 3-hour UV light irradiation. Because the filtration and the centrifugation were not effective to collect all the nanocomposites from the solution, the suspensions were dried at 80 °C overnight after each degradation experiment to evaporate water from the ibuprofen and nanofibers mixture solution. Fresh ibuprofen solution was then added to the dried nanocomposites to repeat the photocatalytic experiment.

2.3 Analytical methods

The adsorption and degradation of ibuprofen were characterized by a high-performance liquid chromatography (HPLC; PerkinElmer Series 200, CT, USA), UV absorbance at 254 nm (UV₂₅₄), and dissolved organic carbon (DOC) quantified through a carbon analyzer (Shimadzu TOC-L, Kyoto, Japan). Specific UV absorbance (SUVA) was quantified by dividing the UV₂₅₄ by the corresponding DOC concentration of ibuprofen samples, which indicates the change of aromatic

structure of ibuprofen molecules. Fluorescence excitation-emission matrix (FEEM) spectroscopy was used to obtain information on characteristics of ibuprofen and potential degradation products in water samples (Aqualog-UV-800-C; Horiba Instruments Inc., NJ).

The intermediate products of the ibuprofen degradation were further investigated using a (+) ESI LC-MS/MS method with an Ultimate 3000 nano-flow LC system coupled to an Orbitrap Fusion mass spectrometer equipped with an Adivon NanoMate nano-electrospray ionization source. The LC solvent system consisted of water with 0.1% formic acid (A) and acetonitrile with 0.1% formic acid (B). The gradient was 2% B at 0 min to 95% B at 60 min with a 10 min 95% B flush and 20 min equilibration between samples. Data-dependent Orbitrap MS2 mass spectra (HCD Energy = 27%, RP = 30000) were collected with a 3-sec cycle time between parent scans (m/z 100-1000, RP = 120000). Easy-IC lock mass calibration was used for parent scans. Background ion signals were excluded from tandem mass spectrometry. Accurate mass measurement for parent ions (e.g. <1 part per million, ppm) was used to constrain tandem mass spectral library matching and impart high confidence to the matched results. Tandem mass spectra were searched against the 2017 NIST tandem MS library and only matches within 1 ppm precursor mass tolerance, 0.5 m/z product ion mass tolerance, a reverse match factor > 700, and C# < 13 were considered robust matches.

2.4 Adsorption and photocatalytic kinetic models

Adsorption kinetics are usually governed by film diffusion and intra-particle diffusion (Lu Lin, 2014; Xuesong Xu, 2015; Lin et al., 2017g). However, the adsorption kinetics in this study are complex due to the desorption of ibuprofen molecules from nanofibers with weak adsorption

bonding energy. Thus, the adsorption capacity and the rate to reach the equilibrium concentration were investigated using a commonly used kinetic model: the pseudo first-order kinetic model.

Typically, the performance of TiO₂ photocatalysis is usually characterized by Langmuir–Hinshelwood kinetic model (Turchi and Ollis, 1990; Ollis et al., 1991; Coleman et al., 2000; Tanaka et al., 2000; Houas et al., 2001; Li et al., 2006; Kumar and Devi, 2011b). It describes a pseudo-first-order kinetics when the initial concentration of ibuprofen is low (millimolar) (Houas et al., 2001; Konstantinou and Albanis, 2004) (Eq. 1):

$$\ln\left(\frac{C_0}{C}\right) = k_{app}t \quad (1)$$

where C_0 and C are the initial and real-time ibuprofen concentrations (mg L⁻¹) at time t (min), k_{app} is the apparent pseudo-first-order rate constant (min⁻¹).

3. Results and discussion

3.1 Characterization of the synthesized catalysts

As shown in Figure 1, the peaks in XRD spectra were referred to primary anatase TiO₂ phase in the synthesized catalysts with presence of trace amount of rutile TiO₂ phase. Energy dispersive X-ray spectra (EDX) indicated the amounts of elements B and N increased with increasing BN dose in the TiO₂-BN nanocomposites (Nasr et al., 2017a). However, the XRD patterns of TiO₂-BN nanocomposites were similar to the pure TiO₂, suggesting the doping of BN had marginal impact on TiO₂ crystallization during electrospinning process.

The elemental mapping images revealed that Ti, O, B and N elements were evenly distributed over the entire area of the synthesized materials (Nasr et al., 2017a). The morphological and structural features of the synthesized materials were examined by TEM as shown in Figure S1. The TiO₂-BN nanocomposites with different BN doses all displayed in a similar fiber form. The crystallite size of the nanofibers can be determined by Scherrer equation (Lin et al., 2017e). The crystallite sizes of the TiO₂-BN nanocomposites increased slightly with the increasing BN dose (24.3-28.7 nm), while the particle size of pure TiO₂ was the smallest of 16.4 nm (Table 1). The increasing trend of grain sizes was also observed by scanning electron microscopy (SEM) images in previous study (Nasr et al., 2017a). This can be attributed to the increase of the solution viscosity induced by inclusion of BN in the electrospinning solution (Nalbandian et al., 2015). This increase revealed the successful incorporation of BN in TiO₂ nanofibers. Besides, previous XPS analysis of TiO₂-BN nanocomposite suggested the formation of chemical B-O-Ti bond between a titanium atom of TiO₂ and a boron atom at the edge of BN in TiO₂-BN nanocomposite (Liu et al., 2017).

Figure 1. XRD patterns of synthesized catalysts. A: anatase; R: rutile.

The specific surface area of the synthesized nanocomposites increased from 19.7 m² g⁻¹ to 49.6 m² g⁻¹ with the increasing dose of BN nanosheets in the composites (Table 1). In general, materials with a large surface area can accelerate adsorption process due to more active adsorption sites available (Lin et al., 2017a). In addition, the UV-Vis absorption spectra of TiO₂, TB1, TB2, TB3, and TB4 measured in previous study (Nasr et al., 2017a) demonstrated that the incorporation of BN nanosheets improved the light absorbance of the TiO₂-BN nanocomposites in the range of 350-550 nm, which corresponds to the irradiance peak of the UV lamp used in the present work. As such, a higher photocatalytic performance could be expected by combining TiO₂ nanofibers with BN nanosheets. The increased light absorption efficiency resulted in an enhanced photocatalytic activity, which was further evidenced by the degradation of ibuprofen using TiO₂-BN nanocomposites in Section 3.3.

Table 1. Characterization of the synthesized nanocomposites

Photocatalysts	TiO ₂	TB1	TB2	TB3	TB4
Specific surface area (m ² g ⁻¹)	19.7	31.8	34.4	48.3	49.6
Crystalline size (nm)	16.4	24.3	25.2	26.3	28.7

3.2 Adsorption kinetics of the synthesized catalysts

The adsorption kinetics for ibuprofen onto different photocatalysts are shown in Figure 2. Ibuprofen molecules adsorbed onto photocatalysts from the aqueous solution increased quickly

over time, and the equilibrium was achieved within 90 min for all photocatalysts. Adsorption onto synthesized nanocomposites showed the same inverted “L” shape, indicating a similar kinetic adsorption process. The adsorption capacity of these catalysts varied in the range of 40-65% after 3-hour adsorption. Ibuprofen had better affinity and higher adsorption with pure TiO_2 than with TiO_2 -BN nanocomposites.

The model parameters and the coefficients of determination (R^2) of the pseudo first-order kinetic equation were calculated based on the experimental data, and the results are summarized in Table 2 (modeling curves are presented in Supporting Information Figure S2). The kinetic model fitted well to the experimental data with the coefficients of determination (R^2) higher than 0.8. The amount of adsorbed ibuprofen at equilibrium (q_e) was approximately 14 mg g^{-1} for all nanocomposites, except for TB4 (10.3 mg g^{-1}). Pure TiO_2 achieved the highest adsorption kinetic rate constant ($k_1 = 0.106 \text{ min}^{-1}$), which was 10 times larger than that of TB1. The adsorption kinetic rate constants followed the order of $\text{TiO}_2 > \text{TB2} > \text{TB3} > \text{TB4} > \text{TB1}$, which was similar to the tendency observed in Figure 2.

Figure 2. Adsorption kinetics of ibuprofen onto different photocatalysts. Error bars represent the standard deviation of duplicate experiments.

Generally, wrapping BN nanosheet onto TiO₂ nanofibers is expected to raise the ibuprofen adsorption onto the photocatalyst due to the larger specific surface area (Table 1). However, the interactions between ibuprofen and the nanocomposite catalysts also govern the adsorption process. Lee *et al.* proposed that the adsorption mechanism between TiO₂ and organics involved the electrostatic (H-bonding) and the covalent (bidentate-bridging mode) bonding (Lee et al., 2011). The surface complex was formed via H-bonding with TiOH/Ti-OH₂ (Ti-O) and COO- (COOH) groups (Lee et al., 2011). Besides, ibuprofen is negatively charged with the acid-base logarithmic ionization constant pK_a of 4.85 at neutral pH (Lin et al., 2017a). TiO₂ nanofibers are nearly neutral at pH 7 with the point of zero charge pH_{pzc} of 6.4, while the BN nanosheets are hydrophilic and negatively charged (Lin et al., 2010). As such, the introduction of BN increased the electrical repulsion between ibuprofen and the catalysts, resulting in lower adsorption. In addition, smaller particle size of pure TiO₂ (Table 1) may contribute to higher ibuprofen adsorption because smaller dimension often facilitates adsorption (Yean et al., 2005).

Table 2. Kinetic parameters for adsorption and photocatalytic activity of the pure TiO₂ and TiO₂-BN nanocomposites

Catalyst	Adsorption			Photocatalysis	
	k ₁ (min ⁻¹)	q _e (mg g ⁻¹)	R ²	k _{app} (min ⁻¹)	R ²

TiO ₂	0.106	14.13	0.836	0.047	0.994
TB1	0.010	13.95	0.974	0.028	0.985
TB2	0.052	14.1	0.996	0.033	0.979
TB3	0.024	13.77	0.964	0.041	0.975
TB4	0.015	10.28	0.974	0.054	0.952

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268 3.3 Photocatalytic kinetics of the synthesized catalysts

269 The photodegradation of ibuprofen under UV irradiation is shown in Figure 3. The degradation
270 efficiency of direct photolysis (no catalyst) was measured under the same conditions for
271 photocatalysis (with catalyst). After 120 min of UV light exposure, there was 27% degradation
272 of ibuprofen without catalyst. It suggests that ibuprofen molecules can absorb UV light in the
273 region of lamp emission. On the other hand, photodegradation of ibuprofen remarkably improved
274 due to the enhancing effect of the nanocomposites under the UV irradiation. After 2 h of UV
275 light exposure, ibuprofen was almost completely degraded in the presence of photocatalysts.

Figure 3. Photocatalytic kinetics of ibuprofen onto different photocatalysts under UV light irradiation. Error bars represent the standard deviation of duplicate experiments.

Langmuir-Hinshelwood kinetic model (Eq. 1) was applied to the photocatalytic data. A plot of $\ln(C_0/C)$ versus time t for ibuprofen degradation is presented in Figure 4, and the kinetic parameters are listed in Table 2. The Langmuir-Hinshelwood kinetic model fitted well the experimental data and the coefficients of determination (R^2) obtained were greater than 0.9. The photocatalytic activity presented a similar trend as BN content, except for pure TiO_2 . As shown in Table 2, the apparent photocatalysis rate of TB4 (0.054 min^{-1}) was higher than pure TiO_2 (0.047 min^{-1}), while TiO_2 had higher adsorption kinetics rate (TiO_2 0.059 vs. TB4 0.039 min^{-1}). Hence, TB4 exhibits excellent photocatalytic activity for ibuprofen removal rather than adsorption. The photocatalytic degradation rates of TiO_2 and TB2 were lower than adsorption rates, inferring that their adsorption rate is faster than the photocatalytic oxidation rate, so oxidation is the control step for the photocatalytic reaction. On the contrary, adsorption is the

control step for TB1, TB3, and TB4 photocatalysis. This was further demonstrated by the multiple photocatalysis cycles discussed in Section 3.6.

Figure 4. Langmuir-Hinshelwood kinetic modeling curves of ibuprofen degradation by the synthesized nanocomposites

Adsorption has both positive and negative impacts on photocatalysis. On the one hand, more adsorbed ibuprofen molecules on the catalyst surface improve the transfer of photogenerated radicals, which explained why the photocatalytic degradation rate of TiO₂ was slightly higher than TB1 and TB2. On the other hand, the photocatalytic activity is mainly affected by light absorption capacity and separation efficiency of e⁻-h⁺ pairs. The adsorbed ibuprofen may hinder the photocatalysis degradation by screening the light access to the catalysis, so photocatalytic degradation rate of TiO₂ was relatively low even with the highest adsorption rate. The photocatalytic degradation rate of TiO₂-BN nanocomposite increased with the increasing BN content probably due to enhanced light absorption intensity (Section 3.1). Meanwhile, due to

electrostatic interactions, the negatively charged BN nanosheets surface can lead to the transfer of h^+ from the TiO_2 particles to the BN nanosheets. As such, the recombination of e^- and h^+ is inhibited with increasing BN dose, which was confirmed by previous photoluminescence study (Nasr et al., 2017a).

3.4 Ibuprofen degradation intermediates and proposed pathway

Aiming at identifying the degradation intermediates, (+) ESI LC-MS/MS was performed to detect the constituents in the ibuprofen solutions after photocatalytic treatment. Figure 5 showed extracted-ion chromatographic peak areas for ibuprofen degradation intermediates (including hydroxyibuprofen, carboxyibuprofen, and oxypropyl ibuprofen) after 2-hour UV irradiation. The proposed photocatalytic degradation pathway of ibuprofen is present in Figure 6. All compounds shown were confirmed by high resolution mass spectrometry, where elemental composition of the parent ions was determined from accurate m/z measurement and tandem mass spectrometry indicates molecular structure. For simplicity, only one isomer among several possible is shown at each step. Some intermediates are proposed to be formed by successive free radical hydroxylation followed by oxidation to yield the carbonyl compounds (aldehyde and ketone). The primary carbon oxidation to carboxylic acid is a facile process and that product is confirmed by our mass spectrometry results. The most reactive site of the ibuprofen molecule in the presence of free radical ($\cdot OH$) is the benzylic carbon, which gives monohydroxylated ibuprofen (confirmed by mass spectrometry). Once the benzylic hydroxyl ibuprofen is formed, it undergoes further oxidation to give the beta-keto acid. Beta keto acids are susceptible for decarboxylation and results the formation of oxypropyl ibuprofen. A similar oxidation decarboxylation route for the degradation of ibuprofen has been reported in the literature (Skoumal et al., 2009).

329 As presented in Figure 3, the sample under direct photolysis (no catalyst) displayed the lowest
330 removal of ibuprofen (27% in 2-hour UV irradiation) and relatively low amounts of degradation
331 intermediates (Figure 5). It suggested only a small fraction of ibuprofen was oxidized and the
332 degradation process was limited to the initial degradation levels. Degradation in the existence of
333 photocatalysts is more complete since both ibuprofen and its degradation products can be
334 adsorbed onto the catalysts. From Figure 3, ibuprofen was almost completely degraded in the
335 presence of photocatalysts after 2-hour UV light exposure, but the amounts of degradation
336 intermediates varied remarkably (Figure 5). Comparing to direct photolysis, the pure TiO_2
337 catalyst showed 27 times, 12 times, and 76 times increase in hydroxyibuprofen,
338 carboxyibuprofen, and oxypropyl ibuprofen, respectively, owing to the degradation of parent
339 compound (ibuprofen). When the BN content increased to 3% (TB1), the amount of degradation
340 intermediates showed 4%, 60%, and 28% reduction relative to the pure TiO_2 . The catalysts with
341 5% BN (TB2) also showed a similar increase in degradation intermediates. However, when the
342 BN content increased to 7% (TB3) and 10% (TB4), the degradation intermediates decreased by
343 over 99%, 99%, and 90% for hydroxyibuprofen, carboxyibuprofen, and oxypropyl ibuprofen,
344 respectively, comparing to the pure TiO_2 . Hence, the increasing BN content of the catalyst
345 facilitated the breakdown of ibuprofen degradation intermediates. However, the further
346 degradation products were not detected with (+) ESI LC-MS/MS probably due to the low
347 concentration of these compounds. These degradation products with smaller carbon chains might
348 be adsorbed onto the catalysts surface and degraded much faster than ibuprofen and its initial
349 degradation intermediates. More analysis of degradation mechanism was discussed in the
350 following Section 3.5.

Figure 5. Extracted-ion chromatographic peak areas for the precursor masses of ibuprofen degradation intermediates derived from (+) ESI LC-MS/MS chromatograms from different catalyst conditions.

Figure 6. Proposed degradation of ibuprofen by TiO₂-BN photocatalysis. Indicated structures were confirmed by mass spectrometry as described.

3.5 Photocatalytic oxidation mechanisms of ibuprofen

The nanocomposite TB4 was chosen to investigate the photocatalytic degradation mechanism of ibuprofen because of its highest photocatalytic activity among the TiO₂-BN nanocomposites. The photodegradation of ibuprofen was analyzed in terms of HPLC, FEEM, DOC, UV₂₅₄, and SUVA measurements (Figure 7). The concentration measured by HPLC was the exact concentration of ibuprofen (parent compound); its concentration reduced rapidly and almost completely disappeared after 2-hour reaction. DOC includes all organic carbons in parent compounds, degradation intermediates and products, serving as an indicator for mineralization degree. Compared with the ibuprofen removal results from the HPLC analysis, the DOC reduction was much slower, with only 25% after 2-h reaction. Although the TB4 was effective to oxidize ibuprofen at a fast rate, the organic intermediates from the ibuprofen oxidation need longer time to be mineralized.

Figure 7. Reduction of DOC, FEEM peak values, and SUVA values during photocatalytic oxidation of ibuprofen using TB4 under UV irradiation

374

375 The UV₂₅₄ and SUVA are used to determine the removal of aromatic fractions of ibuprofen
376 (Stoll et al., 2015). As shown in Figure 7, the SUVA and UV₂₅₄ results exhibited a similar
377 removal trend suggesting aromatic fraction of ibuprofen was degraded. Even though ibuprofen
378 concentration reduced by 79% for the first 30 min, there were 14% and 3% increases in terms of
379 SUVA and UV₂₅₄, respectively. The degradation of ibuprofen was further characterized by the
380 FEEM spectroscopy. The fluorescence peak disappeared after 2-hour of treatment (Figure S3).
381 The excitation-emission peak observed at 220 nm/285 nm corresponds to aromatic compounds
382 (Chen et al., 2003). The fluorescence peak volume was used to compare the aromatic fractions in
383 the tested ibuprofen solution at different reaction time (Figure 7). The peak volume reduced by
384 94% after 2-hour reaction, supporting the decrease in aromatic compounds, consistent with the
385 results of UV₂₅₄ and SUVA that aromatic organic carbon was removed from the solution.

386 The results implied that the intermediates become more aromatic than the parent compound after
387 a partial oxidation. As discussed in Section 3.4, intermediates with more complex structure were
388 detected after photocatalytic treatment. Interestingly, there was only 13% reduction of ibuprofen
389 concentration in the last 30 min of reaction (from 90 to 120 min), but SUVA, UV₂₅₄, and FEEM
390 decreased by 61%, 50%, and 38%, respectively. It suggested that decomposition from parent
391 compounds to intermediates happened mainly during the initial 90 min, while degradation of
392 intermediates into shorter carbon chains occurred mainly after decomposition. Besides, the DOC
393 reduction in the first 30 min probably attributed to the adsorption of ibuprofen onto TB4 (Figure
394 3), the mineralization of ibuprofen was a slow process in the following 90 min due to the tardy
395 reduction of DOC. These results demonstrated that photocatalytic oxidation of ibuprofen

molecules by the synthesized nanofibers is a complex multi-step process, degradation of intermediates is probably a control step for the ibuprofen photocatalytic oxidation.

3.6 Recyclability of the synthesized catalysts

The recyclability of the photocatalysts by reusing the catalysts in multiple treatment cycles is crucial in accessing the practical application of a catalyst, because the catalyst may be poisoned by some species (e.g., reaction intermediates) during the reaction process, or may otherwise decompose. In the present work, the recyclability of the pure TiO_2 and TB4 was investigated by performing 16 UV irradiation cycles. In each cycle, UV light was irradiated for 3-hour at room temperature. The adsorption of ibuprofen by TiO_2 and TB4 achieved 65% and 41% after 3-hour mixing (Figure 2), but the photocatalytic activity of TiO_2 and TB4 still remained 80% and 79% after eight cycles of photocatalytic reaction (Figure 8).

After eight cycles, the used nanofibers were irradiated by the UV lamp for 3-hour in deionized water. The treatment efficiency was recovered to 98% and 95% for TB4 and TiO_2 after 3-hour regeneration by UV irradiation. However, the degradation efficiency of ibuprofen receded gradually in the following eight cycles, although the TB4 remarkably outperformed TiO_2 . According to Section 3.4, more ibuprofen and intermediates were detected for TiO_2 than TB4, hence, weaker photocatalytic activity is probably attributed to the generation and accumulation of intermediates during the catalytic process, which remained on the catalyst and may restrain the further adsorption and degradation of ibuprofen. In addition to the UV irradiation as an effective method for the regeneration of exhausted nanofibers, chemical regeneration should also be studied to fully restore the photocatalysts.

Figure 8. Ibuprofen removal efficiency of TB4 and TiO₂ catalysts during multiple photocatalytic cycles under the irradiation of UV light

4. Conclusions

The adsorption, photolysis, and photocatalytic oxidation of ibuprofen by TiO₂-BN nanocomposites were studied using a multiple analysis approach. The degradation kinetics, mechanisms, and intermediate products were elucidated. The primary conclusions of the study are summarized as follows.

- Wrapping BN nanosheets onto TiO₂ nanofibers improved light absorption efficiency and specific surface area of TiO₂ nanofibers, and enhanced separation effectiveness of photogenerated e⁻-h⁺ pairs.
- Both adsorption and photocatalytic oxidation of ibuprofen by the TiO₂-BN nanocomposites followed the first-order kinetic models. The amount of adsorbed ibuprofen at

equilibrium reached approximately 14 mg g⁻¹ for the TiO₂-BN photocatalysts. The photocatalytic degradation rate increased with the increasing BN content.

- Analysis of degradation intermediates suggested that the increasing BN content of the catalyst facilitated the breakdown of ibuprofen degradation intermediates (hydroxyibuprofen, carboxyibuprofen, and oxypropyl ibuprofen). No further degradation intermediates were detected by the (+) ESI LC-MS/MS chromatograms likely due to lower concentrations of these compounds.

- Multiple photocatalytic cycles were conducted to investigate the reusability of the photocatalysts. UV irradiation of the catalysts in clean water could recover the degradation efficiency to 98% and 95% for TB4 and TiO₂. However, photocatalytic activity receded gradually, especially TiO₂, as a result of the accumulation of intermediates during the catalytic process. Further chemical regeneration should be investigated to fully recover the photocatalysts.

- Combining TiO₂ nanofibers with BN nanosheets provides an innovative method to improve the photocatalytic performance for the degradation of organic contaminants. Further studies on photocatalytic oxidation of organic contaminants of emerging concerns should be conducted using solar light in environmental conditions (e.g., wastewater).

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List of the Figure Captions

Figure 1. XRD patterns of synthesized catalysts. A: anatase; R: rutile.

Figure 2. Adsorption kinetics of ibuprofen onto different photocatalysts. Error bars represent the standard deviation of duplicate experiments.

Figure 3. Photocatalytic kinetics of ibuprofen onto different photocatalysts under UV light irradiation. Error bars represent the standard deviation of duplicate experiments.

Figure 4. Langmuir-Hinshelwood kinetic modeling curves of ibuprofen degradation by the synthesized nanocomposites

Figure 5. Extracted-ion chromatographic peak areas for the precursor masses of ibuprofen degradation intermediates derived from (+) ESI LC-MS/MS chromatograms from different catalyst conditions.

Figure 6. Proposed degradation of ibuprofen by TiO₂-BN photocatalysis. Indicated structures were confirmed by mass spectrometry as described.

Figure 7. Reduction of DOC, FEEM peak values, and SUVA values during photocatalytic oxidation of ibuprofen using TB4 under UV irradiation

Figure 8. Ibuprofen removal efficiency of TB4 and TiO₂ catalysts during multiple photocatalytic cycles under the irradiation of UV light