



HAL
open science

Community-wide scan identifies fish species associated with coral reef services across the Indo-Pacific

Eva Maire, Sebastien Villeger, Nicholas A. J. Graham, Andrew S. Hoey, Joshua Cinner, Sebastian C. A. Ferse, Catherine Aliaume, David J. Booth, David A. Feary, Michel Kulbicki, et al.

► To cite this version:

Eva Maire, Sebastien Villeger, Nicholas A. J. Graham, Andrew S. Hoey, Joshua Cinner, et al.. Community-wide scan identifies fish species associated with coral reef services across the Indo-Pacific. *Proceedings of the Royal Society B: Biological Sciences*, 2018, 285 (1883), 10.1098/rspb.2018.1167. hal-02002283

HAL Id: hal-02002283

<https://hal.umontpellier.fr/hal-02002283v1>

Submitted on 6 Jan 2025

HAL is a multi-disciplinary open access archive for the deposit and dissemination of scientific research documents, whether they are published or not. The documents may come from teaching and research institutions in France or abroad, or from public or private research centers.

L'archive ouverte pluridisciplinaire **HAL**, est destinée au dépôt et à la diffusion de documents scientifiques de niveau recherche, publiés ou non, émanant des établissements d'enseignement et de recherche français ou étrangers, des laboratoires publics ou privés.

1 **Community-wide scan identifies fish species associated with**
2 **coral reef services across the Indo-Pacific**

3

4 **Authors:** Eva Maire^{1,2,*}, Sébastien Villéger¹, Nicholas A. J.
5 Graham³, Andrew S. Hoey², Joshua Cinner², Sebastian C. A.
6 Ferse⁴, Catherine Aliaume¹, David J. Booth⁵, David A. Feary⁶,
7 Michel Kulbicki⁷, Stuart A. Sandin⁸, Laurent Vigliola⁹ & David
8 Mouillot^{1,2}

9

10 **Author affiliations:**

11 ¹ MARBEC, UMR IRD-CNRS-UM-IFREMER 9190, Université Montpellier,
12 34095 Montpellier Cedex, France.

13 ² Australian Research Council Centre of Excellence for Coral Reef Studies, James
14 Cook University, Townsville, QLD 4811 Australia.

15 ³ Lancaster Environment Centre, Lancaster University, Lancaster, LA1 4YQ, UK

16 ⁴ Leibniz Centre for Tropical Marine Ecology (ZMT), Fahrenheitstrasse 6, D-
17 28359 Bremen, Germany

18 ⁵ School of Life Sciences, University of Technology, Sydney, New South Wales
19 2007, Australia

20 ⁶ Ecology & Evolution Group, School of Life Sciences, University Park,
21 University of Nottingham, Nottingham NG7 2RD, UK

22 ⁷ UMR Entropie, Labex Corail, –IRD, Université de Perpignan, 66000 Perpignan,
23 France

24 ⁸ Scripps Institution of Oceanography, University of California, San Diego, La
25 Jolla, California 92093, USA

26 ⁹ Institut de Recherche pour le Développement, UMR IRD-UR-CNRS
27 ENTROPIE, Laboratoire d'Excellence LABEX CORAIL, BP A5, 98848 Nouméa
28 Cedex, New Caledonia

29

30 ***Corresponding author:** Eva Maire

31 Contact: eva.maire@umontpellier.fr / +33 4 67 14 40 93

32 Address: Laboratoire Marine Biodiversity Exploitation and Conservation,
33 Université Montpellier, Place Eugène Bataillon, 34095 Montpellier, France.

34

35 **Author Contributions.** D.M. conceived the study with support from E.M., S.V.
36 and N.A.J.G.; E.M. developed and ran the analyses; E.M. wrote the first draft of
37 the manuscript and all the authors made substantial contributions.

38

39 **Data accessibility statement.** If the manuscript is accepted, we confirm that the
40 data supporting the results will be archived in an appropriate public repository
41 such as Dryad.

42

43 **Competing interests.** The authors declare no competing interests.

44 **Funding.** Data collection was supported, amongst others, by NOAA's Coral Reef
45 Conservation Program and the US National Science Foundation for the Moorea
46 Coral Reef LTER site (OCE 1637396, 1236905, 1026851, 0417412).

47

48 **Article type:** Research article

49

50 **Keywords:** ecosystem functioning, biodiversity, fish community, key species,
51 ecosystem services

52 Number of words in the Abstract: 157

53 Number of words in main body of the paper: 4357

54 Number of figures: 5

55 Number of references: 61

56

57

58

59

60

61

62

63

64 **ABSTRACT**

65 Determining whether all species or only a subset are necessary to maintain
66 ecosystem functioning and services is still an unresolved issue. Identifying such key
67 species remains challenging, especially in the tropics where many species co-occur.
68 Here, we develop a new community-wide scan (CWS) approach, analogous to the
69 genome-wide scan, to identify fish species that contribute disproportionately to fish
70 biomass and live coral cover across the Indo-Pacific. We found that only a limited
71 set of species (51 out of ~400 or 13%), belonging to various functional groups and
72 evolutionary lineages, are strongly and positively associated to fish biomass and live
73 coral cover. Many of these species have not previously been identified as
74 functionally important and thus, may be involved in unknown, yet important,
75 biological mechanisms that help sustain healthy and productive coral reefs. CWS
76 studies can uncover the species that contribute to ecosystem functioning and
77 services and help guide experiments to decipher underlying ecological mechanisms.

78

79

80

81

82

83

84 **MAIN TEXT**

85 **INTRODUCTION**

86 Within the context of global changes and biodiversity loss, effective ecosystem
87 management relies on a better understanding of the causal pathways between
88 ecological communities and the myriad of services they sustain (Cardinale *et al.*
89 2012; Cheung *et al.* 2016; Ricketts *et al.* 2016; Ratcliffe *et al.* 2017). Experiments
90 that manipulate community compositions have unambiguously demonstrated the
91 positive effect of species diversity on ecosystem functioning over short and long
92 timescales (Cardinale *et al.* 2006; Isbell *et al.* 2011; Isbell *et al.* 2015; Isbell *et al.*
93 2018). Recent studies have also convincingly shown that natural species-rich
94 communities are more productive and can deliver higher rates of ecosystem services
95 than impoverished communities (Grace *et al.* 2016; Duffy *et al.* 2017). Beyond the
96 mere number of species, the diversity of species traits and evolutionary histories
97 have been related to enhanced ecosystem functioning in both controlled experiments
98 and natural communities (Cadotte *et al.* 2009; Flynn *et al.* 2011; Mora *et al.* 2014;
99 Gross *et al.* 2017). In parallel, another line of evidence suggests that particular
100 species are key to ecosystem functioning as they contribute disproportionately to
101 certain processes when present (Bellwood *et al.* 2012; Reich 2012; Bozec *et al.*
102 2016; Tobner *et al.* 2016; Meyer *et al.* 2018). However, identifying key species
103 remains highly challenging in diverse ecosystems, such as tropical reefs or
104 rainforests, where many species co-occur and individual species can have multiple

105 or unique contributions to ecosystem functions and services (Bozec *et al.* 2016;
106 Pigot *et al.* 2016).

107 To tackle this challenge, ecologists can now use the increasing availability of
108 extensive and standardized databases that have compiled environmental, social,
109 biological and ecosystem service information across space and time (Cinner *et al.*
110 2016; Duffy *et al.* 2017). This emergence of large socio-ecological databases
111 parallels what happened 20 years ago in genetics with advances in genome
112 sequencing, generating millions of genetic variants for individual loci. To identify
113 genetic variants among this myriad of sequences that are more frequent in people
114 with a particular disease or traits of biomedical significance, genome-wide scan or
115 genome-wide association studies (GWAS) were developed (Bush & Moore 2012).
116 Such an approach is powerful to relate a given biological feature or trait to its
117 underlying genetics, based on the simple idea that if a genetic variant increases the
118 frequency of a given trait it should be more frequent in individuals with this trait
119 than expected by chance (Visscher *et al.* 2017). Although this approach does not
120 attribute causality, it can uncover previously unsuspected, yet important, potential
121 biological mechanisms and pathways (McCarthy & Hirschhorn 2008). Although
122 similar approaches have not been used in ecology, they hold much promise in
123 empirical community ecology where only a few, among dozens or even hundreds
124 of species (counterparts of genetic variants) can disproportionately drive ecosystem
125 functioning and the delivery of services (counterparts of diseases, traits or

126 phenotypes) (Bellwood *et al.* 2006; Straub & Snyder 2006; Bozec *et al.* 2016;
127 Meyer *et al.* 2018). This approach could also reveal the unknown level of
128 ecological pleiotropy in communities, i.e. the propensity that a single species can be
129 key to many ecological functions and services (Hooper *et al.* 2005; Gascon *et al.*
130 2015). This term was initially coined by Strauss & Irwin (2004) [48] by analogy to
131 genetic pleiotropy when one gene can influence two or more seemingly unrelated
132 phenotypic traits. Under ecological pleiotropy a few species may underpin many
133 different ecosystem functions or services and would deserve particular conservation
134 actions.

135 Identifying functionally important or key species is particularly challenging in
136 biodiverse ecosystems, due largely to the complexity of interactions among species
137 and with their environment including human disturbances. For example, despite the
138 large body of research on coral reefs, the identification of fish species that
139 disproportionately drive ecosystem functioning is still in its infancy (Hoey &
140 Bellwood 2009; Bellwood *et al.* 2012). The functional importance of most coral reef
141 fishes is still poorly understood, and no study has scanned entire fish communities
142 to detect potential links with ecosystem functioning and services at large scale. Here,
143 we develop a new community-wide scan (CWS) approach, analogous to the GWAS
144 approach, to identify key fish species that contribute to the delivery of services on
145 coral reef ecosystems. Here ‘key’ has a different meaning than ‘keystone’ which
146 corresponds to a “species whose effect is large, and disproportionately large relative

147 to its abundance” (Paine 1966; Power *et al.* 1996). We define key species as those
148 consistently and significantly associated, so above a certain statistical threshold, to
149 a certain level of ecosystem functioning or services.

150 More precisely, we propose a statistical framework and use empirical data from
151 1,824 Indo-Pacific coral reefs hosting ~400 fish species to determine species whose
152 presence disproportionately influences fish biomass and live coral cover. We then
153 place those key species in a functional trait space (Villegger *et al.* 2008), and on a
154 reef fish phylogeny to show the extent of species traits and evolutionary lineages
155 that are necessary to sustain these two services on coral reefs. Identifying key
156 species can provide new research priorities to elucidate ecological processes by
157 which such candidate species positively affect coral reefs and to motivate a
158 diversification of management options to maintain fish communities and their
159 associated services in the face of a highly uncertain future.

160

161 **MATERIAL AND METHOD**

162 **General framework**

163 The Community-Wide Scan (CWS) framework to identify species that are
164 associated with higher levels of ecosystem services involves three steps (Figure
165 1): i) collecting environmental, socio-economic, species abundance and/or
166 biomass, and indicators of ecosystem services data across many sites; ii) modeling

167 ecosystem services as functions of this large set of predictor variables (socio-
168 economic, environmental conditions, and species richness). The relevance of this
169 comprehensive initial, or reference, model (M_0) is validated according to its R^2
170 and its Akaike Information Criterion (AIC_{M_0}); iii) testing the effect of each species
171 separately on each ecosystem service beyond the effect of previous variables
172 including species richness. For this, the presence of a given candidate species
173 (binary variable) is added as an explanatory variable to M_0 . The resulting model
174 M_1 is evaluated according to its AIC ($AIC_{M_{1k}}$). A species is declared as a potential
175 key contributor to the ecosystem service if ΔAIC ($AIC_{M_0} - AIC_{M_{1k}}$) > 4 and if its
176 partial effect is positive (Figure 1).

177

178 **Coral reef data**

179 *Coral reef services.* The proxies for coral reef services we considered are fish
180 biomass and live coral cover which support, among many others, food security,
181 shoreline protection and recreational value (Burke *et al.* 2011; Kittinger *et al.*
182 2012; Harris *et al.* 2018). Coral cover and fish biomass are already monitored at
183 the global scale using visual censuses as well as underwater video surveys (e.g.
184 Reef Life Survey, Catlin Sea Survey) and are highly sensitive to local human
185 activities (e.g. fishing, habitat destruction, pollution) and global climate change
186 (Hughes *et al.* 2018) and thus, can be considered as key variables for the health
187 and productivity of coral reefs (Bozec *et al.* 2016; Cinner *et al.* 2016; Hughes *et al.*

188 2017). We used data from 1,824 coral reefs in 26 nations, states, or territories
189 located across the Indo-Pacific which include fish biomass and live coral cover
190 estimates.

191

192 *Fish Biomass.* Reef fish biomass estimates were based on instantaneous visual
193 counts from 4,694 surveys collected from 1,824 reefs (Fig. S1). All surveys used
194 two census methods (belt-transects or distance sampling) and were conducted
195 between 2004 and 2013. On average 2.4 transects (sd: 1.32; 1-10) were performed
196 on each reef. Within each survey area, reef associated fishes were identified to
197 species level, abundance counted, and total length (TL) estimated.

198 To make estimates of biomass from these transect-level data comparable among
199 studies, we:

200 i) Considered only Indo-Pacific reefs and retained families that were
201 consistently included in surveys and were above a minimum size cut-
202 off. Thus, we retained counts of non-cryptic reef fish species >10cm in
203 total length, that are reef-associated (30 families, 748 species) (Table
204 S1). We did not include sharks as they were often excluded from visual
205 surveys. We calculated total biomass of fishes on each reef using
206 published species-level length-weight relationship parameters or those
207 available on FishBase (Froese & Pauly 2014). When length-weight

208 relationship parameters were not available for a species, we used the
209 parameters for a closely related species or genus.

210 ii) Directly accounted for depth and habitat as covariates in the model (see
211 “environmental drivers” section in Supplementary Material).

212 iii) Accounted for any potential bias among census method by including
213 census method and sampling area as covariates in the model.

214 Biomass values were calculated at the reef scale and demonstrates a wide
215 distribution (mean: 1,055 kg.ha⁻¹; 2-25,910).

216 *Coral cover.* Percent cover of live coral was based on 1,715 point-intercept
217 transects or quadrats collected from 741 reefs (Fig. S2). All surveys were
218 conducted between 2008 and 2013. On average 2.3 transects (sd: 1.03; 1-4) were
219 performed on each reef. To make estimates of coral cover from these transect-
220 level data comparable among studies, we included depth, habitat and census
221 method as covariates in the model (see “drivers” section in Supplementary
222 Material). Coral cover values were calculated at the reef scale and demonstrates a
223 wide distribution (mean: 27%; 3-94).

224

225

226 *Initial models and species candidates.* For each of the 1,824 reefs located in the
227 Indo-Pacific we collected and used 12 relevant social and environmental variables
228 (listed below), together with the occurrence, abundance and size of 748 reef fish

229 species (Cinner *et al.* 2016). To provide the initial model (M_0) and reference
230 Akaike Information Criterion (AIC_{M_0}) we modeled fish biomass and live coral
231 cover using linear mixed models (LMM) with the complete set of socio-economic
232 and environmental conditions plus species richness as predictor variables. For
233 each of the 748-fish species present in this dataset, we calculated its occurrence
234 on each reef. Then, we computed the number of reefs where a fish species occurs
235 and excluded those present on less than 1% (threshold commonly used for rare
236 species) of the reefs so 18 and 7 reefs for fish biomass and coral cover,
237 respectively. Thus, 381 fish species grouped into 116 genera and 30 families were
238 considered as potential candidate species (binary variable as presence/absence in
239 the model M_1).

240

241 **Identifying potential key contributors to ecosystem services.** Each of the 381
242 species were tested as candidates for improving prediction of reef fish biomass
243 and live coral cover given the socioeconomic and environmental conditions. More
244 precisely, we tested presence of each candidate species as an additional
245 explanatory binary variable to M_0 to compute model M_1 and its associated AIC
246 (AIC_{MIK}). Finally, a species was identified as a key contributor to the ecosystem
247 service if, when included, $\Delta AIC > 4$ and if its partial effect was positive (positive
248 coefficient in the model). The binary variable describing the presence/absence of
249 species was computed according to its occurrence (i.e. presence of at least 1

250 individual) but could be also determined using any relative abundance threshold
251 (Fig. 1 and supplemental materials).

252

253 **Environmental and socio-economic variables.** The variables included in the
254 models were environmental: 1) oceanic productivity, 2) habitat type, 3) depth and
255 socio-economic: 4) Management 5) local human population growth rate, 6)
256 gravity of local population, 7) gravity of markets, 8) levels of human development
257 (Human Development Index), 9) human population size, 10) levels of tourism,
258 11) degree of voice and accountability of citizens, and 12) reef fish landings
259 (tons)/km² of reef (definitions and detail on data collection for these variables are
260 provided in the Supplementary Material).

261

262 **Statistical analyses.** We first computed two linear mixed models (LMM), which
263 predicted fish biomass and live coral cover respectively, while accounting for the
264 different scales at which the data were collected as random effects (reef location,
265 site, and nation/state, see Supplementary Material), 11 key environmental and
266 socio-economic variables expected to influence reef conditions (Cinner *et al.*
267 2016) and fish species richness as fixed effects (Supplementary Material). To
268 evaluate the fit of the two linear mixed models, we checked the relationship
269 between observed and predicted values. Model validation and quality control
270 procedures are described in the Supplementary Material.

271 In order to quantify the potential net benefit of each identified key species, we
272 extracted the net effect of each key species for biomass and live coral cover using
273 a partial plot from linear mixed models while the other variables were held
274 constant.

275 We next investigated whether reefs with several key species show high levels of
276 fish biomass and live coral cover. To control for the effects of species richness we
277 compared modeled estimates of fish biomass and live coral between reefs while
278 increasing the number of key species. We defined the number of key species for
279 every reef and chose the richest quartile as a threshold (i.e. 4 and 6 key species
280 for biomass and live coral cover respectively). We next created 3 categories of
281 reefs: those with no key species, those with at least one key species but below the
282 richness threshold (4 and 6 for biomass and live coral cover respectively) and
283 those with more key species than the threshold.

284

285 **Functional space and entities.** The 381-fish candidates species were functionally
286 described using six traits: (1) size, coded using 3 ordered categories: 10-30cm,
287 30.1-50cm, >50cm ; (2) mobility, coded using 3 ordered categories: sedentary,
288 mobile within a reef and mobile between reefs; (3) period of activity, coded using
289 3 ordered categories: diurnal, both diurnal and nocturnal, and nocturnal ; (4)
290 schooling, coded using 5 ordered categories: solitary, pairing, or living in small
291 (3-20 individuals), medium (20-50 individuals) or large groups (>50 groups) ; (5)

292 vertical position in the water column, coded using 3 ordered categories: benthic,
293 benthopelagic and pelagic; (6) diet, coded using 7 trophic categories:
294 herbivorous-detritivorous, macro-algal browser, invertivorous targeting sessile
295 invertebrates, invertivorous targeting mobile invertebrates, planktivorous,
296 piscivorous, and omnivorous, i.e. fish for which both plant and animal material
297 are important in their diet. Values for these six traits were taken from the global
298 trait database on tropical reef fishes from Mouillot *et al.* (2014) (Mouillot *et al.*
299 2014). Since all traits were categorical, species with identical traits were grouped
300 into functional entities. Most functional entities comprise species from different
301 genera (Mouillot *et al.* 2014).

302 We assessed the functional richness (FRic), e.g. the functional space occupied by
303 the key fish species for biomass and coral cover respectively, using the convex
304 hull volume index proposed by Cornwell *et al.* (2006) (Cornwell *et al.* 2006). This
305 volume corresponds to the amount of functional space filled by key species, where
306 axes are defined by species traits.

307

308 **Fish phylogeny.** We used a time-calibrated phylogeny of Acanthomorph fishes
309 (Near *et al.* 2013) which covers the 33 major reef fish families. (i.e. Labridae,
310 Chaetodontidae, Gobiidae, Blenniidae, Apogonidae, Epinephelinae, Serranidae,
311 Pomacentridae, Acanthuridae, Balistidae, Lutjanidae, Callionymidae,
312 Carangidae, Cirrhitidae, Haemulidae, Hemiramphidae, Holocentridae,

313 Lethrinidae, Monacanthidae, Mullidae, Nemipteridae, Ostraciidae, Pempheridae,
314 Pinguipedidae, Platycephalidae, Pomacanthidae, Pseudochromidae,
315 Scorpaenidae, Siganidae, Sparidae, Synanceiidae, Tetraodontidae,
316 Tripterygiidae). Some fish genera (e.g. *Elagatis* and *Parupeneus* for example)
317 recorded on reefs were missing in this phylogeny.

318

319 **RESULTS**

320 **Predictability of fish biomass and coral cover**

321 The two initial (M0) models explained 79% and 61% of the variance in fish
322 biomass and live coral cover, respectively (Fig. S3 and Supplementary Material).
323 The residuals of the two models were normally distributed (Fig S3). In total, 8
324 and 6 variables had the highest importance (Akaike weight = 1) to predict fish
325 biomass and live coral cover respectively (see Tables S2 & S3). Fish species
326 richness, oceanic productivity, population size, tourism and census method were
327 the main predictors of both fish biomass and coral cover. Depth, management,
328 and sampling area were also important predictors of fish biomass while habitat
329 type was important to predict coral cover (see Tables S2 & S3).

330

331 **Key species associated to reef fish biomass**

332 Among the 381 fish species considered as candidates, only 26 species (7%) were
333 significantly related to fish biomass beyond the initial set of variables ($\Delta AIC > 4$
334 and positive effect) considering their presence (at least 1 individual) (Table S4).
335 Those 26 key species covered a wide breadth of phylogenetic lineages (Fig. 2),
336 representing 16 out of 116 genera and 8 out of 31 families (i.e. Acanthuridae,
337 Carangidae, Labridae, Lethrinidae, Lutjanidae, Mullidae, Scombridae,
338 Serranidae).

339 When considering functional traits, we found that those 26 key species represented
340 24 different functional entities (Table S6) demonstrating a very low functional
341 redundancy with 1.1 species per functional entity (range 1-2). In addition, key
342 species had contrasting functional traits with all body sizes (from 10cm to > 50cm)
343 and all diets (7 trophic categories) represented (Table S6). Together these 26 key
344 species filled 20% of the whole functional space defined by the 240 functional
345 entities corresponding to the 381 candidate species ($FRic = 0.20$; Fig. 3).

346

347 **Key species associated to live coral cover**

348 We found that 28 reef fish species out of 381 (7%), were significantly and
349 positively related to coral cover ($\Delta AIC > 4$) considering their presence (at least 1
350 individual) (Table S5). Those 28 key species also encompassed a wide breadth of
351 phylogenetic lineages (Fig. 2), representing 15 out of 116 genera and 8 out of 31

352 families (i.e. Acanthuridae, Chaetodontidae, Cirrhitidae, Haemulidae, Labridae,
353 Lutjanidae, Monacanthidae, Serranidae).

354 When considering functional traits, we found that the 28 key fish species were
355 distributed among 17 different functional entities (Table S6). Key fish species for
356 coral cover showed some degree of functional redundancy with, on average, 1.6
357 key species per functional entity (range 1-6). This higher functional redundancy
358 translated into a more restricted functional space filled by these key species (only
359 5% with FRic = 0.05; Fig. 3). Species of all sizes (from 10cm to > 50cm) and
360 almost all diets (6 diet categories out of 7) were significantly associated to live
361 coral cover. However, large mobile predators and large herbivorous fishes were
362 not considered as key for live coral cover (Table S6).

363

364 **Low overlap between species key to fish biomass and coral cover**

365 The two sets of key fish species associated to total fish biomass and live coral
366 cover (26 and 28 key species, respectively) represented all together less than 10%
367 of the 381 fish species tested as candidates. Only three species (*Acanthurus*
368 *albipectoralis*, *Lutjanus bohar*, *Lutjanus gibbus*), four genera (*Acanthurus*,
369 *Chlorurus*, *Lutjanus*, and *Scarus*) and four families (Acanthuridae, Labridae,
370 Lutjanidae and Serranidae) were common to both sets and thus, significantly
371 associated to the two proxies of coral reef services (Fig. 2, Table S6).

372 Only six functional entities were common and significantly associated to both
373 biomass and live coral cover (Fig. 3), namely small and medium herbivores, small
374 planktivores, medium and large fishes targeting mobile invertebrates and meso-
375 predators (Table S6).

376

377 **The net benefit of key species for fish biomass and live coral cover**

378 When present, each key species belonged to a community with a median level of
379 fish biomass higher (560 kg/ha, range 439-773 kg/ha) than the median biomass
380 observed when absent (370 kg/ha, range 337-385). Similarly, live coral cover was
381 estimated at a median value of 50% (range 36-82) when each key species was
382 present against 34% (range 26-38) when absent (Fig. 4). For clarity, we only
383 presented the net effect of the four most significant key species (lowest AIC
384 compared to AIC(M₀), see Tables S4 & S5) associated to biomass and live coral
385 cover (Fig. 4). It is important to note that these four most significant key species
386 (lowest AIC) did not necessarily provide the highest level of biomass and coral
387 cover. Highest levels of biomass were obtained when either *Elagatis bipinnulata*,
388 *Naso tuberosus*, *Bolbometopon muricatum* or *Plectropomus laevis* were present
389 in the model, while highest levels of coral cover were reached when either
390 *Plectorhinchus picus*, *Acanthurus leucosternon*, *Chaetodon guttatissimus* or
391 *Chaetodon meyeri* were present (Fig S4).

392 It is not only individual key species, but also the accumulation of key species that
393 promoted high levels of ecosystem services. For instance, reefs with more than
394 four key species reached a median level of biomass of 1,150 kg/ha (range 362-
395 3715), i.e. three times the median biomass observed (370 kg/ha, range 86-1380)
396 in reefs with an intermediate number of key species (from 1 to 3 key species) and
397 more than seven times higher than the median level of fish biomass reached in
398 reefs having no key species (156 kg/ha, range 12-812). Although less pronounced,
399 reefs with at least 6 key fish species showed a median live coral cover of 40%
400 (range 20-68) while reefs with no key species had a median level of 31% (range
401 18-54) live coral cover (Fig. 5).

402

403 **DISCUSSION**

404 **Sustaining healthy and productive coral reefs**

405 In the present study, many different fish species (~400 species candidates) were
406 scanned and only 26 and 28 species were identified as positively related to fish
407 biomass and live coral cover respectively, with only three species being common
408 to both. In total, these 51 species (i.e. ~13% of the species pool tested), represent
409 35 distinct functional entities (out of 240 so 15%) and are widespread in the
410 functional space. While large-bodied species may be expected to
411 disproportionately contribute to fish biomass, our results indicate that only 25%
412 (7 out of 26, see Table S6) of key species for fish biomass were large-bodied

20

413 (>50cm), which is directly comparable to the percentage of large-bodied species
414 among the initial candidate species (20% or 49 functional entities out of 240). We
415 also show that 35% of key species for fish biomass are smaller than 30cm (9 out
416 of 26, see Table S6). The positive association with fish biomass is thus
417 independent of body size.

418 It comes as no surprise that some key fish species identified in this study have
419 already aroused considerable interest in coral reef ecology. Herbivorous fish
420 support coral reef resilience by controlling algal growth, influencing competitive
421 interactions between corals and macroalgae, and preventing coral-algal phase
422 shifts (Bellwood *et al.* 2004; Mumby *et al.* 2006; Hughes *et al.* 2007; Rasher *et*
423 *al.* 2013; Graham *et al.* 2015; Bozec *et al.* 2016), and therefore may contribute to
424 the maintenance of high coral cover and fish biomass. In particular, scarinine
425 parrotfishes (i.e., *Bolbometopon*, *Chlorurus*, *Hipposcarus* and *Scarus*, see Tables
426 S4,S5&S6) play critical roles as grazers and bioeroders of the reef substratum
427 (Bellwood *et al.* 2003; Mumby 2006), and their abundances have strong positive
428 effects on cover of corals and hence accretion rates of the reef (Cramer *et al.*
429 2017). Further, grazing and detritivorous acanthurids (i.e., *Acanthurus* and
430 *Ctenochaetus*, see Tables S4,S5&S6) intensely graze epilithic algal turfs
431 (Marshall & Mumby 2012; Rasher *et al.* 2013), while benthic-feeding
432 unicornfishes (i.e., *Naso*, see Tables S4&S6) play a significant role in macroalgal
433 removal (Fox & Bellwood 2008; Hoey & Bellwood 2009).

434 By contrast, some key species identified in the present study have not previously
435 been identified as playing significant roles. While predation is a key process
436 shaping prey behavior and populations (Rasher *et al.* 2017), structuring ecological
437 communities (Boaden & Kingsford 2015), and promoting nutrient capacity and
438 primary production (Allgeier *et al.* 2016; Allgeier *et al.* 2017), no individual
439 predator species have been expressly identified as beneficial for total fish biomass
440 and coral cover. Here, we show that the predatory species of *Aprion*, *Caranx*,
441 *Cephalopholis*, *Elagatis*, *Gymnosarda*, *Lethrinus*, *Lujtanus*, *Oxycheilinus*, and
442 *Plectropomus* (Fig 4, Tables S4&S5, Fig S4) may play a critical role for fish
443 biomass and live coral cover, although the exact pathways through which they act
444 remain to be elucidated.

445

446 **Low ecological pleiotropy on coral reefs**

447 The finding that a limited number of functionally and evolutionary different
448 species are positively related to high levels of fish biomass and coral cover (Fig.
449 2,3&S4) supports the idea that sustaining ecosystem services may require a large
450 breadth of particular attributes beyond the number of species (Soliveres *et al.*
451 2016; Gross *et al.* 2017). The limited overlap between the two sets of species
452 significantly associated to two proxies of key services (3 species, 4 genera and 6
453 functional entities) suggests a low level of ecological pleiotropy (Strauss & Irwin
454 2004), i.e. that a single species, genus or functional entity cannot be key to many

455 independent ecosystem functions and services. Extended to the community level,
456 we show that ecological pleiotropy, the exact opposite of functional redundancy,
457 is not the norm on coral reefs. This finding explains why the multi-functionality
458 of ecosystems relies more strongly on biodiversity than do single functions (Isbell
459 *et al.* 2011; Lefcheck *et al.* 2015; Soliveres *et al.* 2016; Gross *et al.* 2017), since
460 some species play unique and thus irreplaceable roles in ecosystems (Bellwood *et*
461 *al.* 2006; Petchey *et al.* 2008). We suggest that this ecological pleiotropy
462 reconciles two opposing views in Biodiversity and Ecosystem Functioning (BEF)
463 research since many complementary species groups and lineages, and hence a
464 large amount of biodiversity, are necessary to sustain ecosystem multi-
465 functionality and associated services. Once combined, those key species appear
466 to provide high benefits in terms of fish biomass and live coral cover (Fig 5).
467 Maintaining habitat heterogeneity and associated processes is thus a major
468 component of management and conservation. Our results call for more species-
469 focused management strategies such as the banning of fishing species considered
470 as key for the ecosystem (Bozec *et al.* 2016). Alternatively, sustaining multi-
471 functionality also requires a broader portfolio approach which may reduce local
472 extinction risk by securing the biodiversity level in an increasingly uncertain
473 future (Webster *et al.* 2017).

474

475 **CWS as a flexible framework to link biodiversity to ecosystem functioning**
476 **and services**

477 The community-wide scan (CWS) approach can be adapted for a wide range of
478 ecosystems, combinations of taxa or interactions and services. Here we only tested
479 the presence of key species, while it would be possible to look for key species
480 groups (pairs or more), key evolutionary lineages or even key biotic interactions.
481 For instance, interactions among local neighbors promote overyielding in species
482 mixtures of tree communities (Fichtner *et al.* 2018). Since those interactions are
483 potentially multiple in species-rich communities they cannot be experimentally
484 tested, but they can emerge from empirical data using the WCS approach. In the
485 same vein, positive effects of some species may only be revealed beyond some
486 thresholds. For instance, herbivorous fish are both important for standing biomass
487 and ecosystem resilience, and it has been shown that maintaining high fish
488 abundance may increase behavioral coupling and thus, promote the consumption
489 of algae by fish (Gil & Hein 2017) while maintaining high parrotfish biomass and
490 implementing a minimum harvest size may be critical in sustaining the health and
491 the productivity of coral reefs (Bozec *et al.* 2016). The way candidates are tested
492 can be modulated while respecting independence between predicted and
493 explanatory variables. For example, presence data can be determined by any
494 abundance threshold such as a minimum number of individuals, cover rate,
495 biomass or level of interactions.

496 On coral reefs, defining species presence based on intraspecific biomass
497 distribution (using upper percentiles or deciles) can promote small-bodied species
498 but can also disadvantage species that are not commonly encountered or have
499 skewed biomass distributions since species need to occur sufficiently to be tested
500 (see supplemental materials and Table S7-10). Rather, defining species presence
501 as a proportion of the intracommunity biomass can be applied independently of
502 the species biomass distribution, but large species may be more likely to reach the
503 biomass threshold. However, we found consistent results between these two
504 procedures since the majority of species detected as key species using the
505 intracommunity approach are also significant using the intraspecific approach
506 reinforcing the robustness of our findings.

507 The CWS framework thus offers a new and flexible way to analyze empirical data
508 relating biodiversity to ecosystem functioning and services. In genetics, the
509 greatest challenge is no longer the identification of the association signals
510 themselves, but the discovery of molecular mechanisms through which genes
511 influence disease risk and phenotypic expression. In ecology, CWS studies can
512 only be considered as initial forays into a better understanding of the complex
513 relationships between particular species, species groups or interactions and
514 ecosystem functioning and associated services. A large number of false-positives,
515 species being detected as key while they are not, may be revealed and no causality
516 is determined in this approach; the main merit is to identify unsuspected and

517 statistically significant positive associations. The logical progression would be to
518 conduct experiments focusing on potential key species or interactions with the
519 ultimate aim of highlighting the underlying ecological or biological processes and
520 pathways that potentially sustain healthy and productive ecosystems.

521

522 **CONCLUSION**

523 The community-wide scan (CWS) approach has the potential to reveal
524 unsuspected contributions to ecosystem functioning and its associated services,
525 especially in complex and biodiverse ecosystems where the detection of such
526 contributions remains challenging. Using coral reefs as our model system, we
527 have identified a phylogenetically and functionally wide set of fish species
528 strongly associated with high levels of fish biomass and live coral cover. While
529 these findings cannot be interpreted as causative, they highlight a limited pool of
530 species (58 fish species), many of which appear to be involved in unsuspected,
531 yet important, biological mechanisms and pathways that could sustain healthy and
532 productive reefs upon which human welfare depends. The CWS approach holds
533 much promise in empirical BEF studies where only a few species, functional or
534 phylogenetic groups, can disproportionately drive ecosystem functioning and the
535 delivery of services. Ultimately, the key species identified form tractable
536 conservation targets, for example the implementation of bans on fishing species
537 considered as key for the ecosystem (Bozec *et al.* 2016). Given the growing

538 interest in the assessment and consequences of the ongoing extinction crisis on
539 ecosystem functioning, such a framework is extremely timely and widely
540 applicable. Our framework offers a new and flexible way to analyze the ongoing
541 massive empirical data relating biodiversity to ecosystem functioning and
542 services with the potential to reconcile two opposing views: species identity vs.
543 diversity.

544

545

546

547

548 **Acknowledgments**

549 We thank M. Beger, E. Brokovich, A. J. Brooks, P. Chabanet, A. Green, C.
550 Gough, C. Mora, M. H. Tupper and I. Williams for compiling and providing
551 essential data for this study. We are grateful to M. Troussellier for his great
552 support and A-S. Tribot for providing many fish silhouettes used in the figures.

553

554

555

556

557

558

559

560

561

562

563

564

565

566

567

568 **References**

569 1.

570 Allgeier, J.E., Burkepille, D.E. & Layman, C.A. (2017). Animal pee in the sea: consumer-
571 mediated nutrient dynamics in the world's changing oceans. *Global Change*
572 *Biology*, 23, 2166-2178.

573 2.

574 Allgeier, J.E., Valdivia, A., Cox, C. & Layman, C.A. (2016). Fishing down nutrients on coral
575 reefs. *Nat Commun*, 7, 12461.

576 3.

577 Bellwood, D.R., Hoey, A.S. & Choat, J.H. (2003). Limited functional redundancy in high
578 diversity systems: resilience and ecosystem function on coral reefs. *Ecology*
579 *Letters*, 6, 281-285.

580 4.

581 Bellwood, D.R., Hoey, A.S. & Hughes, T.P. (2012). Human activity selectively impacts the
582 ecosystem roles of parrotfishes on coral reefs. *Proceedings of the Royal Society*
583 *B-Biological Sciences*, 279, 1621-1629.

- 584 5.
- 585 Bellwood, D.R., Hughes, T.P., Folke, C. & Nystrom, M. (2004). Confronting the coral reef
586 crisis. *Nature*, 429, 827-833.
- 587 6.
- 588 Bellwood, D.R., Hughes, T.P. & Hoey, A.S. (2006). Sleeping functional group drives coral-
589 reef recovery. *Current Biology*, 16, 2434-2439.
- 590 7.
- 591 Boaden, A.E. & Kingsford, M.J. (2015). Predators drive community structure in coral reef fish
592 assemblages. *Ecosphere*, 6, 1-33.
- 593 8.
- 594 Bozec, Y.-M., O'Farrell, S., Bruggemann, J.H., Luckhurst, B.E. & Mumby, P.J. (2016).
595 Tradeoffs between fisheries harvest and the resilience of coral reefs.
596 *Proceedings of the National Academy of Sciences of the United States of*
597 *America*, 113, 4536-4541.
- 598 9.
- 599 Burke, L., Reyntar, K., Spalding, M. & Perry, A. (2011). *Reefs at risk revisited*.
- 600 10.
- 601 Bush, W.S. & Moore, J.H. (2012). Chapter 11: Genome-Wide Association Studies. *PLOS*
602 *Computational Biology*, 8, e1002822.
- 603 11.
- 604 Cadotte, M.W., Cavender-Bares, J., Tilman, D. & Oakley, T.H. (2009). Using Phylogenetic,
605 Functional and Trait Diversity to Understand Patterns of Plant Community
606 Productivity. *Plos One*, 4.
- 607 12.
- 608 Cardinale, B.J., Duffy, J.E., Gonzalez, A., Hooper, D.U., Perrings, C., Venail, P. *et al.* (2012).
609 Biodiversity loss and its impact on humanity. *Nature*, 486, 59-67.
- 610 13.
- 611 Cardinale, B.J., Srivastava, D.S., Duffy, J.E., Wright, J.P., Downing, A.L., Sankaran, M. *et al.*
612 (2006). Effects of biodiversity on the functioning of trophic groups and
613 ecosystems. *Nature*, 443, 989-992.
- 614 14.
- 615 Cheung, W.W.L., Reygondeau, G. & Frölicher, T.L. (2016). Large benefits to marine fisheries
616 of meeting the 1.5°C global warming target. *Science*, 354, 1591.
- 617 15.
- 29

- 618 Cinner, J.E., Huchery, C., MacNeil, M.A., Graham, N.A.J., McClanahan, T.R., Maina, J. *et al.*
619 (2016). Bright spots among the world's coral reefs. *Nature*, 535, 416-419.
- 620 16.
- 621 Cornwell, W.K., Schwilk, D.W. & Ackerly, D.D. (2006). A trait-based test for habitat filtering:
622 Convex hull volume. *Ecology*, 87, 1465-1471.
- 623 17.
- 624 Cramer, K.L., O'Dea, A., Clark, T.R., Zhao, J.-x. & Norris, R.D. (2017). Prehistorical and
625 historical declines in Caribbean coral reef accretion rates driven by loss of
626 parrotfish. *Nat Commun*, 8, 14160.
- 627 18.
- 628 Duffy, J.E., Godwin, C.M. & Cardinale, B.J. (2017). Biodiversity effects in the wild are
629 common and as strong as key drivers of productivity. *Nature*, 549, 261.
- 630 19.
- 631 Fichtner, A., Härdtle, W., Bruelheide, H., Kunz, M., Li, Y. & von Oheimb, G. (2018).
632 Neighbourhood interactions drive overyielding in mixed-species tree
633 communities. *Nat Commun*, 9, 1144.
- 634 20.
- 635 Flynn, D.F.B., Mirotnick, N., Jain, M., Palmer, M.I. & Naeem, S. (2011). Functional and
636 phylogenetic diversity as predictors of biodiversity–ecosystem-function
637 relationships. *Ecology*, 92, 1573-1581.
- 638 21.
- 639 Fox, R.J. & Bellwood, D.R. (2008). Remote video bioassays reveal the potential feeding impact
640 of the rabbitfish *Siganus canaliculatus* (f: Siganidae) on an inner-shelf reef of
641 the Great Barrier Reef. *Coral Reefs*, 27, 605-615.
- 642 22.
- 643 Froese, R. & Pauly, D. (2014). FishBase. version (2/2011). *World Wide Web electronic*
644 *publication. Available at: <http://www.fishbase.org> (accessed 30 May 2014).*
- 645 23.
- 646 Gascon, C., Brooks, Thomas M., Contreras-MacBeath, T., Heard, N., Konstant, W., Lamoreux,
647 J. *et al.* (2015). The Importance and Benefits of Species. *Current Biology*, 25,
648 R431-R438.
- 649 24.
- 650 Gil, M.A. & Hein, A.M. (2017). Social interactions among grazing reef fish drive material flux
651 in a coral reef ecosystem. *Proceedings of the National Academy of Sciences*,
652 114, 4703.

- 653 25.
- 654 Grace, J.B., Anderson, T.M., Seabloom, E.W., Borer, E.T., Adler, P.B., Harpole, W.S. *et al.*
655 (2016). Integrative modelling reveals mechanisms linking productivity and plant
656 species richness. *Nature*, 529, 390-+.
- 657 26.
- 658 Graham, N.A.J., Jennings, S., MacNeil, M.A., Mouillot, D. & Wilson, S.K. (2015). Predicting
659 climate-driven regime shifts versus rebound potential in coral reefs. *Nature*, 518,
660 94-97.
- 661 27.
- 662 Gross, N., Bagousse-Pinguet, Y.L., Liancourt, P., Berdugo, M., Gotelli, N.J. & Maestre, F.T.
663 (2017). Functional trait diversity maximizes ecosystem multifunctionality.
664 *Nature Ecology & Evolution*, 1.
- 665 28.
- 666 Harris, D.L., Rovere, A., Casella, E., Power, H., Canavesio, R., Collin, A. *et al.* (2018). Coral
667 reef structural complexity provides important coastal protection from waves
668 under rising sea levels. *Science Advances*, 4.
- 669 29.
- 670 Hoey, A.S. & Bellwood, D.R. (2009). Limited Functional Redundancy in a High Diversity
671 System: Single Species Dominates Key Ecological Process on Coral Reefs.
672 *Ecosystems*, 12, 1316-1328.
- 673 30.
- 674 Hooper, D.U., Chapin, F.S., Ewel, J.J., Hector, A., Inchausti, P., Lavorel, S. *et al.* (2005).
675 Effects of biodiversity on ecosystem functioning: A consensus of current
676 knowledge. *Ecological Monographs*, 75, 3-35.
- 677 31.
- 678 Hughes, T.P., Anderson, K.D., Connolly, S.R., Heron, S.F., Kerry, J.T., Lough, J.M. *et al.*
679 (2018). Spatial and temporal patterns of mass bleaching of corals in the
680 Anthropocene. *Science*, 359, 80.
- 681 32.
- 682 Hughes, T.P., Kerry, J.T., Alvarez-Noriega, M., Alvarez-Romero, J.G., Anderson, K.D., Baird,
683 A.H. *et al.* (2017). Global warming and recurrent mass bleaching of corals.
684 *Nature*, 543, 373-377.
- 685 33.
- 686 Hughes, T.P., Rodrigues, M.J., Bellwood, D.R., Ceccarelli, D., Hoegh-Guldberg, O., McCook,
687 L. *et al.* (2007). Phase shifts, herbivory, and the resilience of coral reefs to
688 climate change. *Current Biology*, 17, 360-365.

- 689 34.
- 690 Isbell, F., Calcagno, V., Hector, A., Connolly, J., Harpole, W.S., Reich, P.B. *et al.* (2011). High
691 plant diversity is needed to maintain ecosystem services. *Nature*, 477, 199-
692 U196.
- 693 35.
- 694 Isbell, F., Cowles, J., Dee Laura, E., Loreau, M., Reich Peter, B., Gonzalez, A. *et al.* (2018).
695 Quantifying effects of biodiversity on ecosystem functioning across times and
696 places. *Ecology Letters*, 0.
- 697 36.
- 698 Isbell, F., Craven, D., Connolly, J., Loreau, M., Schmid, B., Beierkuhnlein, C. *et al.* (2015).
699 Biodiversity increases the resistance of ecosystem productivity to climate
700 extremes. *Nature*, 526, 574-577.
- 701 37.
- 702 Kittinger, J., Finkbeiner, E., Glazier, E. & Crowder, L. (2012). Human dimensions of coral reef
703 social-ecological systems. *Ecology and Society*, 17.
- 704 38.
- 705 Lefcheck, J.S., Byrnes, J.E.K., Isbell, F., Gamfeldt, L., Griffin, J.N., Eisenhauer, N. *et al.*
706 (2015). Biodiversity enhances ecosystem multifunctionality across trophic
707 levels and habitats. *Nat Commun*, 6, 6936.
- 708 39.
- 709 Marshall, A. & Mumby, P.J. (2012). Revisiting the functional roles of the surgeonfish
710 *Acanthurus nigrofuscus* and *Ctenochaetus striatus*. *Coral Reefs*, 31, 1093-1101.
- 711 40.
- 712 McCarthy, M.I. & Hirschhorn, J.N. (2008). Genome-wide association studies: potential next
713 steps on a genetic journey.
- 714 41.
- 715 Meyer, S.T., Ptacnik, R., Hillebrand, H., Bessler, H., Buchmann, N., Ebeling, A. *et al.* (2018).
716 Biodiversity–multifunctionality relationships depend on identity and number of
717 measured functions. *Nature Ecology & Evolution*, 2, 44-49.
- 718 42.
- 719 Mora, C., Danovaro, R. & Loreau, M. (2014). Alternative hypotheses to explain why
720 biodiversity–ecosystem functioning relationships are concave-up in some
721 natural ecosystems but concave-down in manipulative experiments. *Scientific*
722 *Reports*, 4, 5427.

723 43.

- 724 Mouillot, D., Vileger, S., Parravicini, V., Kulbicki, M., Ernesto Arias-Gonzalez, J., Bender,
725 M. *et al.* (2014). Functional over-redundancy and high functional vulnerability
726 in global fish faunas on tropical reefs. *Proceedings of the National Academy of*
727 *Sciences of the United States of America*, 111, 13757-13762.
- 728 44.
- 729 Mumby, P.J. (2006). The Impact Of Exploiting Grazers (Scaridae) On The Dynamics Of
730 Caribbean Coral Reefs. *Ecological Applications*, 16, 747-769.
- 731 45.
- 732 Mumby, P.J., Dahlgren, C.P., Harborne, A.R., Kappel, C.V., Micheli, F., Brumbaugh, D.R. *et*
733 *al.* (2006). Fishing, trophic cascades, and the process of grazing on coral reefs.
734 *Science*, 311, 98-101.
- 735 46.
- 736 Near, T.J., Dornburg, A., Eytan, R.I., Keck, B.P., Smith, W.L., Kuhn, K.L. *et al.* (2013).
737 Phylogeny and tempo of diversification in the superradiation of spiny-rayed
738 fishes. *Proceedings of the National Academy of Sciences of the United States of*
739 *America*, 110, 12738-12743.
- 740 47.
- 741 Paine, R.T. (1966). Food Web Complexity and Species Diversity. *Am Nat*, 100, 65-+.
- 742 48.
- 743 Petchey, O.L., Eklof, A., Borrvall, C. & Ebenman, B. (2008). Trophically unique species are
744 vulnerable to cascading extinction. *Am Nat*, 171, 568-579.
- 745 49.
- 746 Pigot, A.L., Bregman, T., Sheard, C., Daly, B., Etienne, R.S. & Tobias, J.A. (2016). Quantifying
747 species contributions to ecosystem processes: a global assessment of functional
748 trait and phylogenetic metrics across avian seed-dispersal networks.
749 *Proceedings of the Royal Society B: Biological Sciences*, 283.
- 750 50.
- 751 Power, M.E., Tilman, D., Estes, J.A., Menge, B.A., Bond, W.J., Mills, L.S. *et al.* (1996).
752 Challenges in the quest for keystones. *Bioscience*, 46, 609-620.
- 753 51.
- 754 Rasher, D.B., Hoey, A.S. & Hay, M.E. (2013). Consumer diversity interacts with prey defenses
755 to drive ecosystem function. *Ecology*, 94, 1347-1358.
- 756 52.
- 757 Rasher, D.B., Hoey, A.S. & Hay, M.E. (2017). Cascading predator effects in a Fijian coral reef
758 ecosystem. *Scientific Reports*, 7, 15684.

- 759 53.
- 760 Ratcliffe, S., Wirth, C., Jucker, T., van der Plas, F., Scherer-Lorenzen, M., Verheyen, K. *et al.*
761 (2017). Biodiversity and ecosystem functioning relations in European forests
762 depend on environmental context. *Ecol Lett*, 20, 1414-1426.
- 763 54.
- 764 Reich, P.B. (2012). Key canopy traits drive forest productivity. *Proceedings of the Royal*
765 *Society B: Biological Sciences*, 279, 2128.
- 766 55.
- 767 Ricketts, T.H., Watson, K.B., Koh, I., Ellis, A.M., Nicholson, C.C., Posner, S. *et al.* (2016).
768 Disaggregating the evidence linking biodiversity and ecosystem services. *Nat*
769 *Commun*, 7.
- 770 56.
- 771 Soliveres, S., van der Plas, F., Manning, P., Prati, D., Gossner, M.M., Renner, S.C. *et al.* (2016).
772 Biodiversity at multiple trophic levels is needed for ecosystem
773 multifunctionality. *Nature*, 536, 456-459.
- 774 57.
- 775 Straub, C.S. & Snyder, W.E. (2006). Species identity dominates the relationship between
776 predator biodiversity and herbivore suppression. *Ecology*, 87, 277-282.
- 777 58.
- 778 Strauss, S.Y. & Irwin, R.E. (2004). Ecological and Evolutionary Consequences of Multispecies
779 Plant-Animal Interactions. *Annual Review of Ecology, Evolution, and*
780 *Systematics*, 35, 435-466.
- 781 59.
- 782 Tobner, C.M., Paquette, A., Gravel, D., Reich, P.B., Williams, L.J. & Messier, C. (2016).
783 Functional identity is the main driver of diversity effects in young tree
784 communities. *Ecology Letters*, 19, 638-647.
- 785 60.
- 786 Villeger, S., Mason, N.W.H. & Mouillot, D. (2008). New multidimensional functional diversity
787 indices for a multifaceted framework in functional ecology. *Ecology*, 89, 2290-
788 2301.
- 789 61.
- 790 Visscher, P.M., Wray, N.R., Zhang, Q., Sklar, P., McCarthy, M.I., Brown, M.A. *et al.* (2017).
791 10 Years of GWAS Discovery: Biology, Function, and Translation.
- 792 62.

793 Webster, M.S., Colton, M.A., Darling, E.S., Armstrong, J., Pinsky, M.L., Knowlton, N. *et al.*
 794 (2017). Who Should Pick the Winners of Climate Change? *Trends in Ecology*
 795 *& Evolution*, 32, 167-173.

796

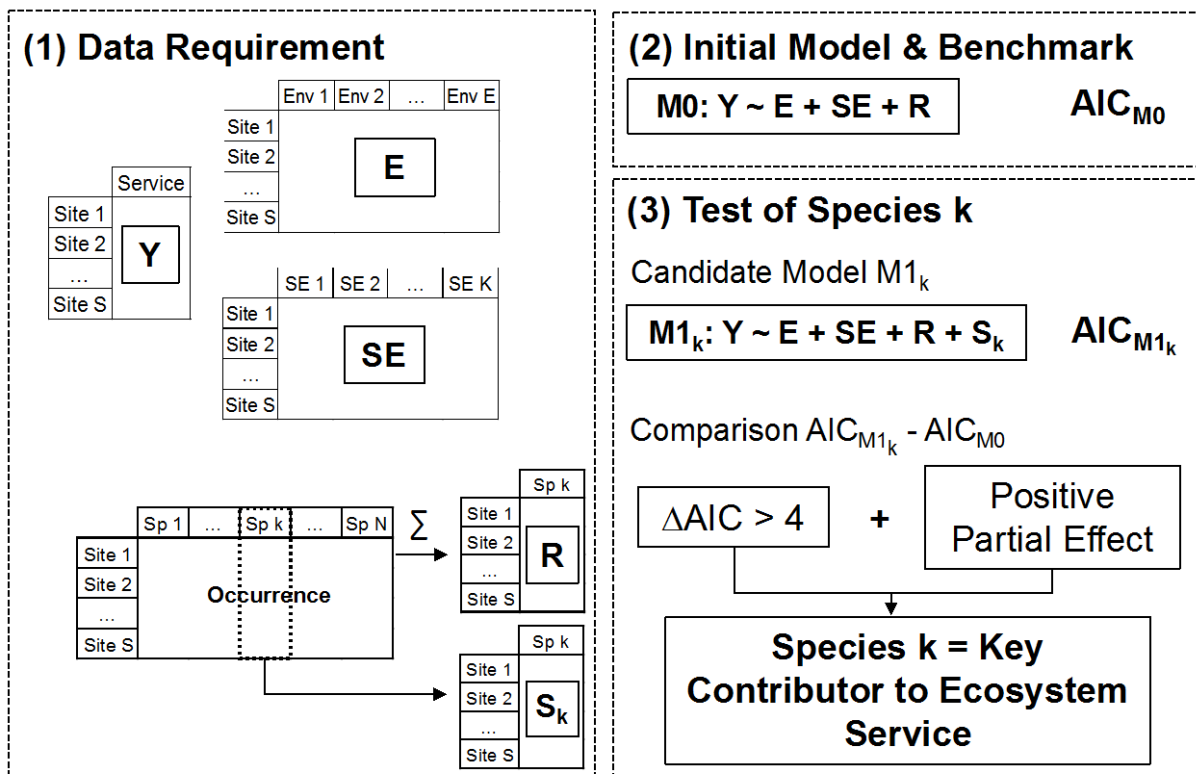
797

798

799

800

801



802

803 Figure 1: Statistical framework to assess the significant contribution of species to
 804 ecosystem services beyond the effects of environmental and socio-economic
 805 conditions and species richness.

806 Step 1: Collecting datasets: for a (large) set of sites, variables describing a given
807 ecosystem service (Y), Environmental (E) and Socio-Economic conditions (SE),
808 and the occurrence of species. Species richness (R) is computed for each site from
809 the Sites-Species matrix as well as the vector (S_k) with presence-absence of each
810 species in sites.

811 Step 2: The goal is to model a given ecosystem service (Y) according to
812 Environmental (E) and Socio-Economic conditions (SE) and species richness (R);
813 to check its relevance according to its explanatory power and to save its Akaike
814 Information Criterion (AIC_{M0}) as a reference for the next step.

815 Step 3: The goal is to identify species key for the studied ecosystem service (Y)
816 adding each candidate species (presence-absence, S_k) as an additional explanatory
817 variable to $M0$ to compute model $M1$ and its associated AIC (AIC_{M1k}). Finally, a
818 species is declared as a key contributor to the ecosystem service if ΔAIC (AIC_{M0} -
819 AIC_{M1k}) > 4 and if its partial effect is positive (positive coefficient in the model).

820

821

822

823

824

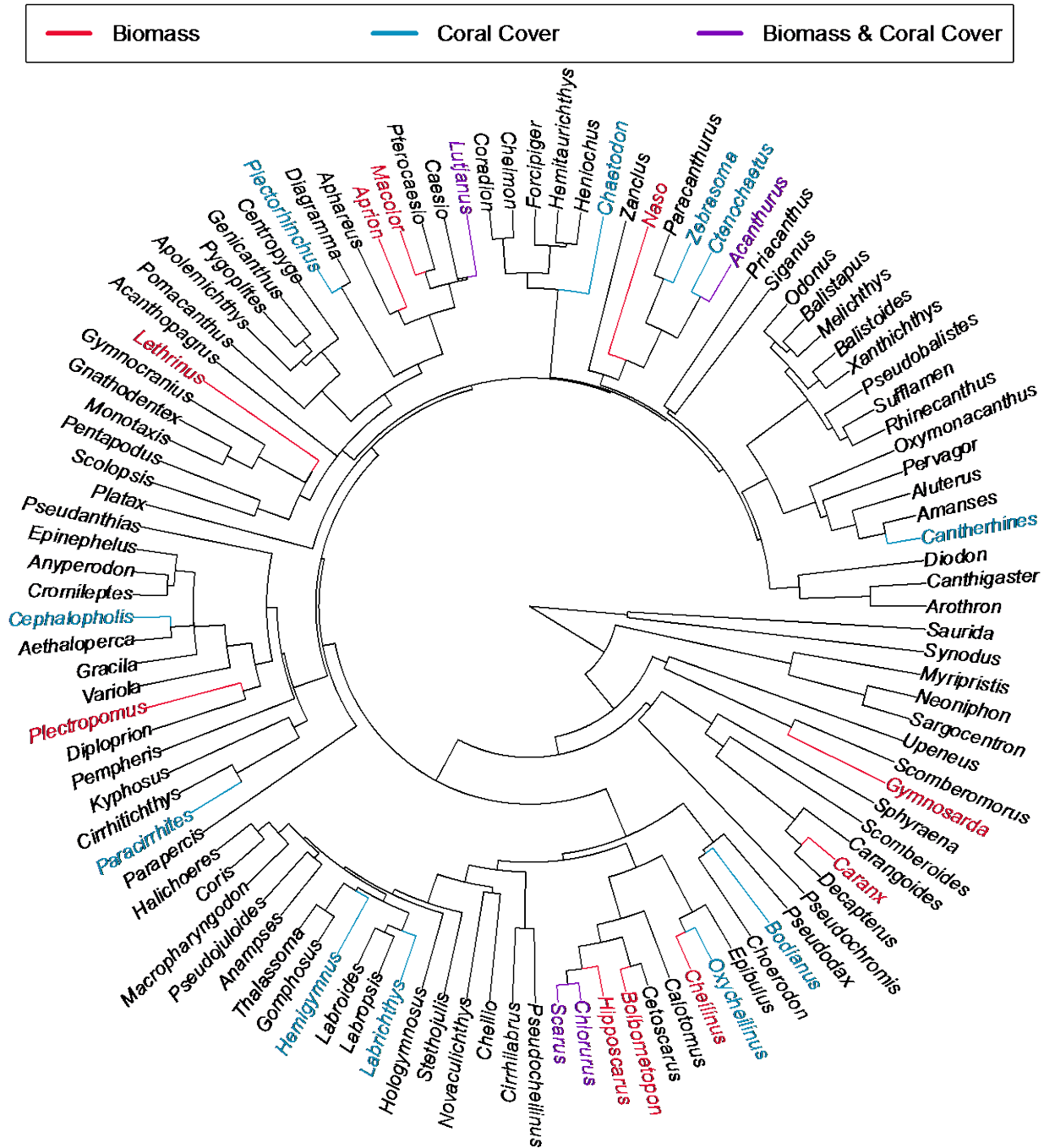
825

826

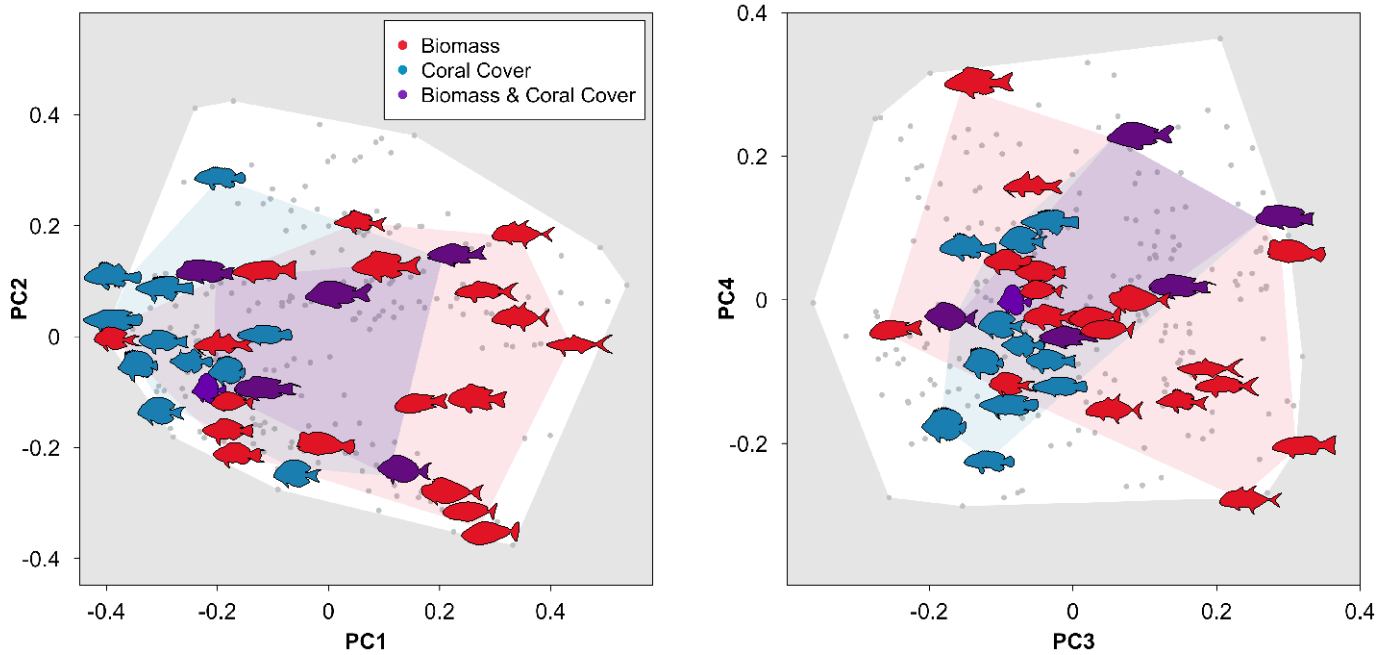
827

828

829

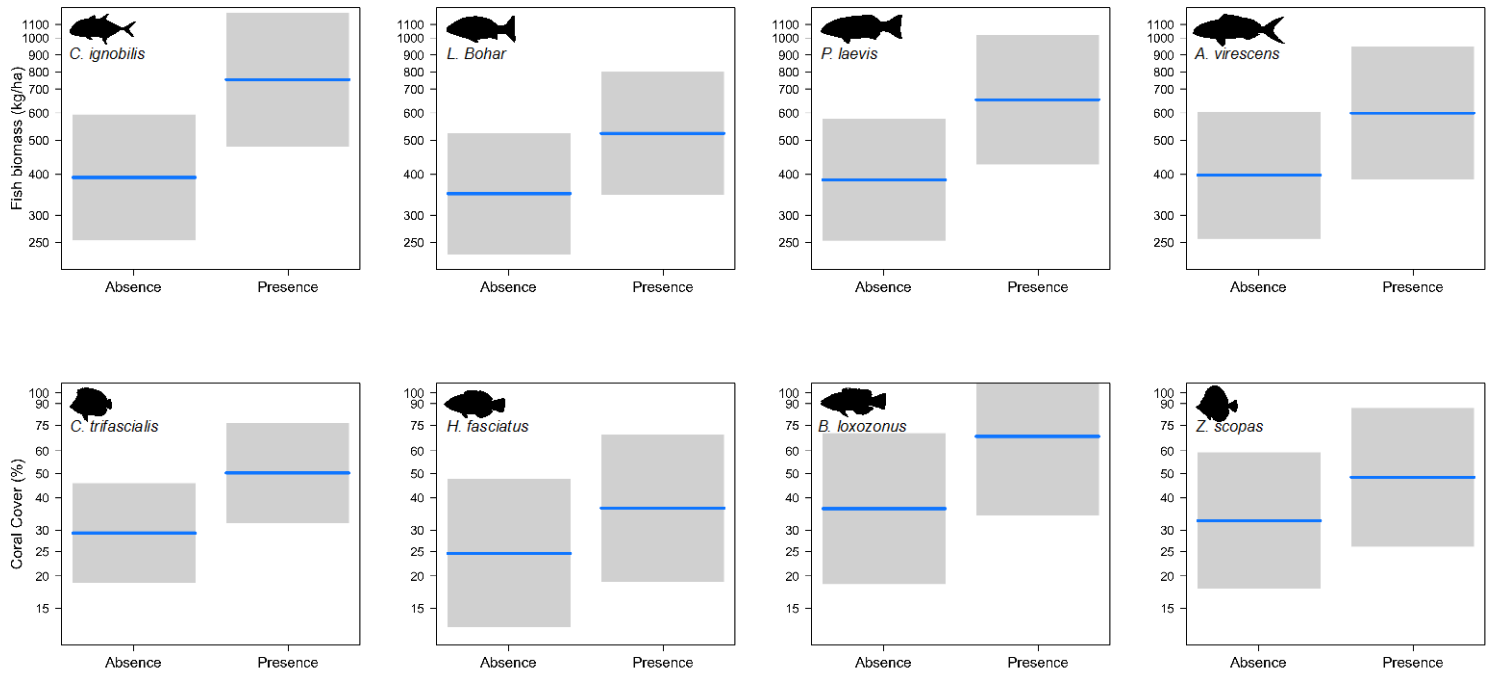


830 Figure 2. Positions of key species for biomass (red), live coral cover (blue) or both
 831 of them (purple) represented as their corresponding fish genera in the Tree of Life
 832 of Coral Reef Fishes, adapted from Near et al. (2013). The 26 key species for
 833 biomass represent 16 genera while the 28 key species for coral cover represent 15
 834 genera with 4 common genera. *Elagatis* and *Parupeneus* genera are missing.



835

836 Figure 3. Functional attributes of key fish species. In total, 51 fish species which
 837 correspond to 35 out of 240 functional entities (15%) have been identified as
 838 strongly related to high biomass (18 red shapes), high live coral cover (11 blue
 839 shapes) or both of them (6 purple shapes). The positions of each of those 35
 840 functional entities in the 4-dimensional functional space is defined according to
 841 species trait values. Fish shapes were chosen to illustrate the main genus of the
 842 species comprised in each functional entity. Other functional entities are
 843 represented with grey dots. Colored areas represent the functional volume filled
 844 by the functional entities that have been identified as strongly related to high
 845 biomass (red, $FRic = 0.2$), high live coral cover (blue, $FRic = 0.05$), both of them
 846 (purple, $FRic = 0.01$) or all functional entities (e.g. all species, white) present in
 847 the dataset. See Table S4 for details of species traits.



848

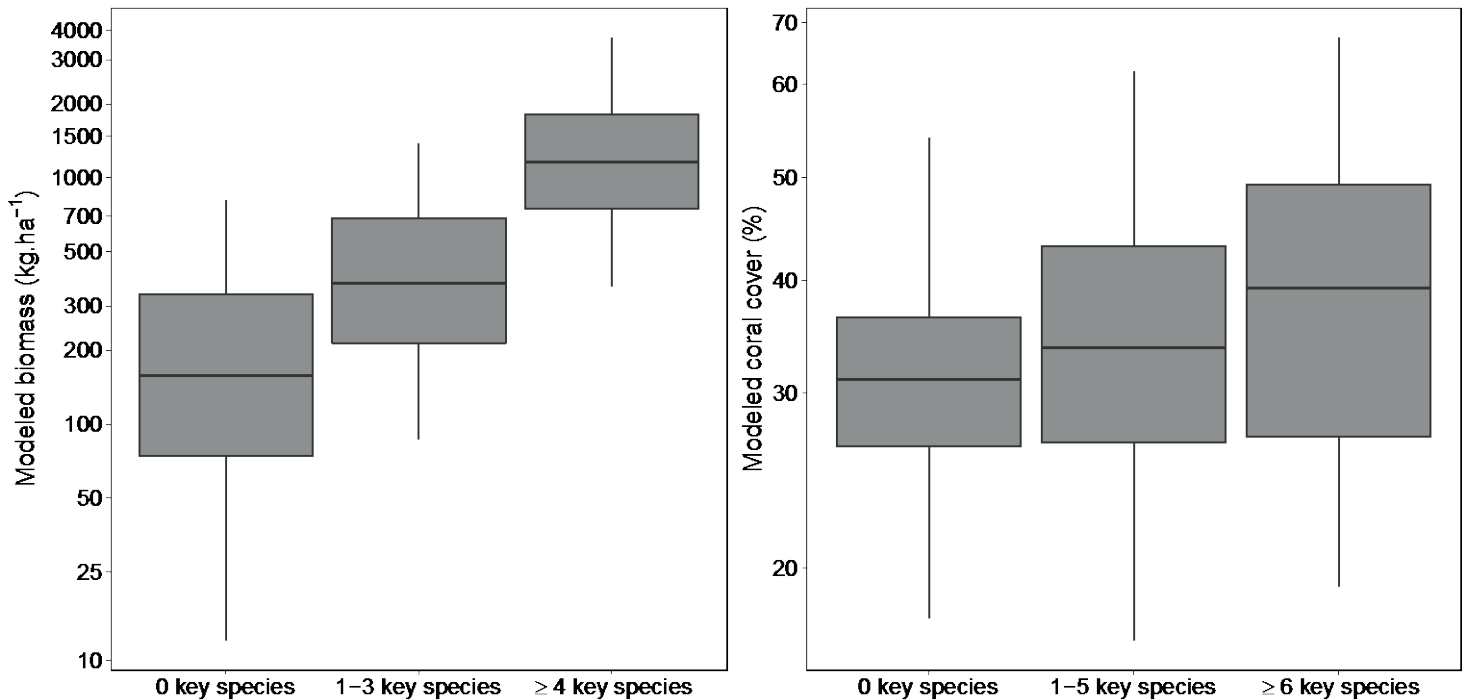
849 Figure 4. Net effect of the 4 most significant key fish species (lowest AIC) for
 850 fish biomass and live coral cover (among the 26 and 28 key species respectively)
 851 using a partial plot from the LMMs while the other variables are held constant.
 852 When present, each key species ensures median level of biomass and live coral
 853 cover significantly ($p < 0.05$) higher than level observed when absent.

854

855

856

857



858

859 Figure 5. The accumulation of key species on coral reefs is positively related to
 860 fish biomass and live coral cover. To control for positive effect of species richness
 861 on ecosystem functioning (biomass and live coral cover), we compared modeled
 862 estimates of fish biomass and live coral between reefs while increasing the
 863 number of key species. 3 categories of reefs were considered based on the 3rd-
 864 quartile of the number of key species as threshold (≥ 4 and ≥ 6 key species for
 865 biomass and coral cover, respectively): i) reefs with no key species, ii) reefs with
 866 at least one key species but below the threshold and iii) reefs above the threshold.
 867 Reef with the highest number of key species reached higher level of biomass
 868 (1,150 kg.ha⁻¹) and coral cover (40%) than their counterparts having no key
 869 species (156 kg.ha⁻¹, 31% respectively). Distributions are represented using 95
 870 percent confidence intervals.